# Status of the Pacific coast groundfish fishery through 2008, stock assessment and fishery evaluation 

## Stock assessments, STAR Panel reports, and rebuilding analyses

October 2008
Pacific Fishery Management Council
Portland, Oregon

# Status of the Chilipepper rockfish, Sebastes goodei, in 2007 

John C. Field<br>Groundfish Analysis Team<br>Fisheries Ecology Division<br>Southwest Fisheries Science Center<br>110 Shaffer Rd.<br>Santa Cruz, CA 95060<br>John.Field@noaa.gov

## EXECUTIVE SUMMARY

Stock Structure: This assessment applies to the chilipepper rockfish (Sebastes goodei) in the waters off of California and Oregon, in the region bounded by the U.S./Mexico border in the south through the Columbia River in the north. Although the distribution is described in the literature as ranging from Queen Charlotte Sound (British Columbia) to Bahia Magdalena (Baja California Sur), the region of greatest abundance is found between Point Conception and Cape Mendocino, California.

Catch History: Chilipepper rockfish have been one of the most important commercial target species in California waters since the 1880s, as well as an important recreational target in Southern California waters historically, and an important recreational target in central and northern California more recently (following the movement of recreational fishing effort to deeper waters in the 1970s and 1980s). Catches were estimated to have begun in 1892, and are estimated to have ranged from several hundred to nearly 1000 tons throughout the first half of the $20^{\text {th }}$ century. Gear types are grouped into four general categories; trawl, hook and line, setnet, and recreational; since World War II a majority has been taken with trawl gear, although hook and line, setnet, and recreational gear have accounted for between 20 and $40 \%$ of landings for most of the last three decades. As early rockfish landings were only reported at the genus level, a combination of historical data and publications, as well as anecdotal accounts of early line, trawl, and recreational fisheries, were used to reconstruct the fraction of catch by gear and sector assumed to be chilipepper. Estimated landings from foreign fisheries from the mid-1960s through the mid-1970s were included as part of the trawl fishery. Throughout most of the past three decades, domestic landings have ranged between approximately 2000 and 3000 tons, however since 2002 landings have averaged less than 100 tons per year (Table E1, Figure E1), primarily a consequence of area closures implemented to rebuild depleted co-occurring species such as bocaccio (S. paucispinis) and canary (S. pinniger) rockfish. Discards are assumed to be negligible in the historical period, however regulatory discards have been substantial in recent years, more than doubling the total catch relative to landings since 2002.

Table E1: Recent commercial and recreational landings (mt, excludes discards)

| Year | Trawl | Hook/line | Setnet | Recreation |
| ---: | ---: | ---: | ---: | ---: |
| 1995 | 1595 | 325 | 94 | 7 |
| 1996 | 1528 | 254 | 58 | 30 |
| 1997 | 1614 | 339 | 83 | 73 |
| 1998 | 1138 | 209 | 78 | 5 |
| 1999 | 839 | 104 | 10 | 24 |
| 2000 | 403 | 51 | 6 | 39 |
| 2001 | 436 | 25 | 5 | 52 |
| 2002 | 162 | 3 | 0.2 | 12 |
| 2003 | 18 | 0.2 | 0.1 | 0 |
| 2004 | 61 | 3 | 1 | 6 |
| 2005 | 60 | 3 | 0.1 | 4 |
| 2006 | 37 | 6 | 0.2 | 1 |



Figure E1: Estimated catches of chilipepper rockfish by major fishery

Data and Assessment: Chilipepper rockfish were last assessed in 1998 (Ralston et al. 1998), at which time they were considered to be above target levels of abundance. From 1978 through 2006, commercial catches and demographic (age and length composition) data for California were obtained from the CalCOM database, those from Oregon were obtained from the PacFIN database, and recreational catches and length composition data were obtained from the RecFIN database beginning in 1981 (with interpolation of landings in missing years). Indices of relative abundance used in the assessment model included a catch per unit effort index from commercial trawl logbooks (from 1980 to 1996, developed and used in the 1998 assessment), an index of relative abundance from a recreational observer program (1987-1998), an index of relative abundance based on the triennial trawl survey (1980-2004), an index of relative abundance based on the Northwest Fishery Science Center Combined Survey (2003-2006), and a coastwide index of pelagic age-0 juvenile abundance developed by combining data from both the SWFSC and NWFSC/PWCC juvenile survey data. Several other potential sources of information were evaluated in earlier models and are discussed in the assessment documentation, although they were not used in the final model. The population was modeled using an age and size structured statistical model, Stock Synthesis II (SS2), version 2.00c, the modeling framework used for most West Coast groundfish assessments.

## Unresolved Problems and Major Uncertainties

The length composition data were down-weighted when associated age-composition data were available, however the approach was acknowledged to be ad-hoc. A more appropriate approach
is to use conditional age-at-length compositions, which should be explored in more detail in future modeling efforts.

The results from the convergence tests with randomly jittered starting parameter values indicated that the likelihood surface is very irregular. In general, biomass trajectories and other critical results do not appear to be sensitive to these differences.

The application of a combined age- and length- based selectivity curve for the recreational CPFV data is somewhat non-traditional and would benefit by either more detailed investigation or an alternative selectivity configuration (an age-based, sex-specific selection curve showed considerable promise).

Future (post-1999) year class strength is highly uncertain; although this model includes highly influential projections through 2006 based on juvenile abundance indices, the failure of the historical (core area) juvenile index to capture much of the year class variability that has been observed is cause for some concern.

The current approach for implementing time-varying growth would benefit by additional data (particularly fishery-independent size at age data), the use of conditional age-at-length data, and more comprehensive efforts to link variability in growth to climate conditions.

Stock Status: This assessment estimates that the spawning biomass of chilipepper rockfish (Sebastes goodei) has increased substantially in recent years, due to a strong 1999 year class as well as greatly reduced harvest rates in commercial and recreational fisheries. The base model result suggests a spawning biomass of 23,889 tons in 2006, corresponding to approximately $70 \%$ of the unfished spawning biomass of 33,390 tons and representing a near tripling of spawning biomass from the estimated low of 8696 tons ( $26 \%$ of unfished) in 1999 (Figure ES-1). As both commercial and recreational fisheries for chilipepper rockfish have been greatly reduced in recent years due to management measures implemented to rebuild depleted rockfish, it is likely that the stock will continue to increase modestly in the longer term under assumptions of equilibrium recruitment.

Table E2: Recent trends in chilipepper rockfish spawning biomass and relative depletion

| year | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Summary biomass | 17008 | 16453 | 15865 | 14578 | 13635 | 13573 | 18556 | 23175 | 27023 | 30022 | 31509 | 32405 | 32401 |
| Spawning biomass | 9812 | 9589 | 9489 | 8968 | 8666 | 9029 | 9536 | 12671 | 17040 | 20229 | 22146 | 23224 | 23827 |
| ~95 confidence limits on spawning biomass |  |  |  |  |  |  |  |  |  |  |  |  |  |
| lower | 8418 | 8033 | 7743 | 7046 | 6608 | 6734 | 7044 | 9281 | 12336 | 14616 | 15984 | 16773 |  |
| upper | 11259 | 11202 | 11296 | 10953 | 10785 | 11379 | 12080 | 16125 | 21830 | 25948 | 28424 | 29797 |  |
| depletion | 0.29 | 0.29 | 0.28 | 0.27 | 0.26 | 0.27 | 0.29 | 0.38 | 0.51 | 0.61 | 0.66 | 0.70 | 0.71 |
| ~95 confidence limits on depletion |  |  |  |  |  |  |  |  |  |  |  |  |  |
| lower | 0.25 | 0.24 | 0.23 | 0.21 | 0.2 | 0.2 | 0.21 | 0.28 | 0.37 | 0.44 | 0.48 | 0.5 |  |
| upper | 0.34 | 0.34 | 0.34 | 0.33 | 0.32 | 0.34 | 0.36 | 0.48 | 0.65 | 0.78 | 0.85 | 0.89 |  |



Figure E2: Estimated trajectory of spawning stock biomass over the modeled period.

## Recruitment

An extremely strong 1999 year class represents the largest estimated historical recruitment, and is the primary cause for the current population trajectory. A year class of comparable strength was also observed in 1984, and the model suggests a series of strong year classes in the late 1960s and early 1970s as well. There are no obvious signs of strong year classes since 1999, and coastwide pelagic juvenile surveys suggest average to low recruitment in recent years, suggesting that the stock may dip slightly in the near term. The projected low recruitments in 2005 and 2006 are based exclusively on the coastwide pelagic juvenile rockfish survey index, which is of short duration and has yet to be validated.

Table E3: Estimated recruitment (1000s) for the recent (1995-2006) period

| year | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| recruits | 15080 | 6555 | 7584 | 12569 | 153415 | 3708 | 15148 | 23831 | 14082 | 25895 | 7647 | 6645 | 32063 |
| $\sim$ 95 confidence limits on recruitment |  |  |  |  |  |  |  |  |  |  |  |  |  |
| lower | 8031 | 1399 | 2723 | 4260 | 104994 | 0 | 9036 | 14220 | 8380 | 15385 | 4546 | 3959 |  |
| upper | 22095 | 11691 | 12465 | 20936 | 202966 | 8023 | 21322 | 33540 | 19842 | 36511 | 10779 | 9358 |  |



Figure E3: Estimated recruitment over the modeled time period

Exploitation Status: Although chilipepper rockfish have been a commercially important species in California waters since well before the second World War, the exploitation rate has rarely exceeded the current target exploitation rate (SPR 50\%). The highest exploitation rates occurred from the late 1980s through the mid 1990s, when they were above target levels and the stock was approaching it's lowest estimated historical levels. From the late 1990s through the present, exploitation rates have been declining significantly, as a result of management measures implemented to rebuild other depleted rockfish species.

Table E4: Estimated exploitation rate (catch/sum bio) for the recent historical period

[^0]

Figure E4: Estimated exploitation rate over the post-World War II period.


Figure E5: SPR relative to stock status through the modeled period

## Reference Points

For rockfish of the genus Sebastes, the proxy for $\mathrm{B}_{\mathrm{MSY}}$ is estimated to be $40 \%$ of the unfished spawning stock biomass $\left(\mathrm{SSB}_{0}\right)$, and the stock is considered to be overfished if the SSB drops below $25 \%$ of $\mathrm{SSB}_{0}$. The proxy for MSY is estimated to be the harvest rate associated with a spawning potential ratio (\%SPR) of $50 \%$, which is a measure of the expected spawning biomass per recruit at the current population level relative to that at the stock's unfished condition (allowing for direct comparison of fishing mortality rates among fisheries with different selectivity patterns). The estimated MSY proxy (harvest associated with an SPR of 50\%) for this assessment is 2099 tons, based on the relative proportion of total catches by fishery assumed in the last year for which data were available (2006), however this in no way intended to imply a de facto sector allocation. The estimated MSY value will change modestly depending upon allocation among fisheries with differing selectivity curves. With a greater proportion of catch allocated to fisheries that are selective at younger ages (trawl and recreational fisheries) the total yield would increase slightly, while if a greater fraction were allocated to hook and line or setnet fisheries, the total equilibrium yield would decrease slightly. Estimates of maximum sustainable yield based on a target equilibrium spawning biomass of $40 \%$ of the unfished spawning biomass, or on the model-estimated MSY, were very modestly greater than the $\mathrm{F}_{50 \%}$ SPR proxy for MSY.

## Table E5: Summary of reference points for chilipepper rockfish

|  |  | $\sim$ O9\% Confidence Limits |  |
| ---: | ---: | ---: | ---: |
| Unfished Stock | Estimate |  | Upper |

## Forecasts

Projections of future biomass were made for two possible catch stream scenarios; status quo (2006) catches and the catch associated with $\mathrm{F}_{50 \%}$ fishing mortality. Under all projections, selection curves were unchanged and the relative proportion of the catch by fishery was assumed to be at the 2006 value for ease of computation. In the $\mathrm{F}_{50 \%}$ projections, the 2007 and 2008 catches were assumed to be at status quo (2006 levels), as it is unlikely that catches could be significantly increased prior to the 2009-2010 management cycle, and as the spawning biomass was greater than $40 \%$ of the unfished level the OY was assumed to be equal to the ABC, and assumed to be fully achieved.

Table E6: Two alternative forecasts of Catch, Spawning Biomass and Depletion

| Year | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| status quo catch | 127 | 127 | 127 | 127 | 127 | 127 | 127 | 127 | 127 | 127 | 127 | 127 |
| SSB | 23827 | 23285 | 22379 | 21574 | 21199 | 21226 | 21531 | 22011 | 22587 | 23211 | 23846 | 24473 |
| Depletion | 0.71 | 0.70 | 0.67 | 0.65 | 0.63 | 0.64 | 0.64 | 0.66 | 0.68 | 0.70 | 0.71 | 0.73 |
|  |  |  |  |  |  |  |  |  |  |  |  |  |
| F50\% catch | 127 | 127 | 3037 | 2576 | 2229 | 2013 | 1901 | 1852 | 1831 | 1822 | 1814 | 1804 |
| SSB | 23827 | 23285 | 22379 | 19139 | 16940 | 15629 | 14911 | 14530 | 14312 | 14164 | 14041 | 13928 |
| Depletion | 0.71 | 0.70 | 0.67 | 0.57 | 0.51 | 0.47 | 0.45 | 0.44 | 0.43 | 0.42 | 0.42 | 0.42 |

## Decision Table

The alternative states of nature used in the decision table were developed in conjunction with the STAR Panel, which considered a variety of potentially appropriate sources of uncertainty. As steepness was thought to be poorly specified for this model (perhaps more so than the natural mortality rate), the lower and upper $25 \%$ of the prior probability distribution for steepness based on the informative prior developed (but not used) in the assessment represented a reasonable means of bracketing uncertainty. As steepness was fixed at the mean value of the prior probability ( 0.57 ) in the base model, the alternative states of nature were consequently 0.34 (low productivity) and 0.81 (high productivity). The three catch streams used in the decision table were developed in coordination with the Groundfish Management Team (GMT) and Groundfish Advisory Subpanel (GAP) representatives to the STAR Panel, and represented "status quo" catches (based on estimates of the 2006 catch, including estimates of discards), equilibrium MSY catches (based on the SPR 0.50 harvest strategy), and ABC catches (based on the $40: 10$ harvest control rule). In all cases, the 2006 total catch estimates were used to apportion theoretical future catches among gear types, importantly this was done to facilitate comparable evaluation of plausible stock trajectories under different states of nature, and in no way implies a recommended or de facto sector allocation.

## Rebuilding Projections

The chilipepper rockfish stock is estimated to be well above the overfished level, such that no rebuilding is required.

Table E7: Decision Table

| year | "Status quo" (2006) catches |  |  | Rec | Low Productivity$h=0.34$ |  | BASE MODEL$h=0.57$ |  | High Productivity$h=0.81$ |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Trawl | Hook/line | Net |  | SpawnBio | epletio | pawnBi | epletio | pawnB | epleti |
| 2007 | 105 | 18 | 0.5 | 4 | 18542 | 0.46 | 23827 | 0.71 | 26482 | 0.87 |
| 2008 | 105 | 18 | 0.5 | 4 | 17887 | 0.44 | 23285 | 0.70 | 25949 | 0.85 |
| 2009 | 105 | 18 | 0.5 | 4 | 16995 | 0.42 | 22379 | 0.67 | 24991 | 0.82 |
| 2010 | 105 | 18 | 0.5 | 4 | 16255 | 0.40 | 21574 | 0.65 | 24072 | 0.79 |
| 2011 | 105 | 18 | 0.5 | 4 | 15929 | 0.39 | 21199 | 0.63 | 23526 | 0.77 |
| 2012 | 105 | 18 | 0.5 | 4 | 15966 | 0.39 | 21226 | 0.64 | 23347 | 0.77 |
| 2013 | 105 | 18 | 0.5 | 4 | 16239 | 0.40 | 21531 | 0.64 | 23436 | 0.77 |
| 2014 | 105 | 18 | 0.5 | 4 | 16645 | 0.41 | 22011 | 0.66 | 23704 | 0.78 |
| 2015 | 105 | 18 | 0.5 | 4 | 17118 | 0.42 | 22587 | 0.68 | 24082 | 0.79 |
| 2016 | 105 | 18 | 0.5 | 4 | 17624 | 0.43 | 23211 | 0.70 | 24522 | 0.80 |
| 2017 | 105 | 18 | 0.5 | 4 | 18141 | 0.45 | 23846 | 0.71 | 24986 | 0.82 |
| 2018 | 105 | 18 | 0.5 | 4 | 18661 | 0.46 | 24473 | 0.73 | 25451 | 0.83 |
| "MSY" catches (base model) |  |  |  |  |  |  |  |  |  |  |
| year | Trawl | Hook/line | Net | Rec | SpawnBio | pletio | pawnBi | pletio | awnBi | epletio |
| 2007 | 105 | 18 | 0.5 | 4 | 18542 | 0.46 | 23827 | 0.71 | 26485 | 0.87 |
| 2008 | 105 | 18 | 0.5 | 4 | 18325 | 0.45 | 23917 | 0.72 | 26652 | 0.87 |
| 2009 | 1735 | 292 | 7 | 64 | 17684 | 0.44 | 23385 | 0.70 | 26111 | 0.86 |
| 2010 | 1735 | 292 | 7 | 64 | 15560 | 0.38 | 21270 | 0.64 | 23899 | 0.78 |
| 2011 | 1735 | 292 | 7 | 64 | 14111 | 0.35 | 19814 | 0.59 | 22259 | 0.73 |
| 2012 | 1735 | 292 | 7 | 64 | 13216 | 0.33 | 18934 | 0.57 | 21149 | 0.69 |
| 2013 | 1735 | 292 | 7 | 64 | 12644 | 0.31 | 18440 | 0.55 | 20424 | 0.67 |
| 2014 | 1735 | 292 | 7 | 64 | 12199 | 0.30 | 18171 | 0.54 | 19956 | 0.65 |
| 2015 | 1735 | 292 | 7 | 64 | 11776 | 0.29 | 18019 | 0.54 | 19650 | 0.64 |
| 2016 | 1735 | 292 | 7 | 64 | 11333 | 0.28 | 17921 | 0.54 | 19446 | 0.64 |
| 2017 | 1735 | 292 | 7 | 64 | 10863 | 0.27 | 17845 | 0.53 | 19302 | 0.63 |
| 2018 | 1735 | 292 | 7 | 64 | 10369 | 0.26 | 17779 | 0.53 | 19194 | 0.63 |
| 40:10 Catches |  |  |  |  |  |  |  |  |  |  |
| year | Trawl | Hook/line | Net | Rec | SpawnBio | pletio | pawnBi | pletio | pawnBi | pletio |
| 2007 | 105 | 18 | 0.5 | 4 | 18652 | 0.46 | 23827 | 0.71 | 26366 | 0.86 |
| 2008 | 105 | 18 | 0.5 | 4 | 17994 | 0.44 | 23285 | 0.70 | 25836 | 0.85 |
| 2009 | 2507 | 429 | 12 | 89 | 17099 | 0.42 | 22379 | 0.67 | 24882 | 0.82 |
| 2010 | 2127 | 364 | 11 | 75 | 13923 | 0.34 | 19139 | 0.57 | 21533 | 0.71 |
| 2011 | 1847 | 308 | 9 | 65 | 11785 | 0.29 | 16940 | 0.51 | 19164 | 0.63 |
| 2012 | 1679 | 266 | 8 | 60 | 10501 | 0.26 | 15629 | 0.47 | 17650 | 0.58 |
| 2013 | 1594 | 241 | 7 | 59 | 9739 | 0.24 | 14911 | 0.45 | 16734 | 0.55 |
| 2014 | 1558 | 228 | 6 | 60 | 9204 | 0.23 | 14530 | 0.44 | 16194 | 0.53 |
| 2015 | 1543 | 223 | 6 | 61 | 8719 | 0.21 | 14312 | 0.43 | 15874 | 0.52 |
| 2016 | 1535 | 220 | 5 | 62 | 8208 | 0.20 | 14164 | 0.42 | 15681 | 0.51 |
| 2017 | 1528 | 219 | 5 | 62 | 7654 | 0.19 | 14041 | 0.42 | 15561 | 0.51 |
| 2018 | 1520 | 218 | 5 | 62 | 7068 | 0.17 | 13928 | 0.42 | 15486 | 0.51 |

Table E8: Summary Table for chilipepper rockfish

| year | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Summary biomass | 17008 | 16453 | 15865 | 14578 | 13635 | 13573 | 18556 | 23175 | 27023 | 30022 | 31509 | 32405 | 32401 |
| Spawning biomass | 9812 | 9589 | 9489 | 8968 | 8666 | 9029 | 9536 | 12671 | 17040 | 20229 | 22146 | 23224 | 23827 |
| $\sim 95$ confidence limits on spawning biomass |  |  |  |  |  |  |  |  |  |  |  |  |  |
| lower | 8418 | 8033 | 7743 | 7046 | 6608 | 6734 | 7044 | 9281 | 12336 | 14616 | 15984 | 16773 |  |
| upper | 11259 | 11202 | 11296 | 10953 | 10785 | 11379 | 12080 | 16125 | 21830 | 25948 | 28424 | 29797 |  |
| depletion | 0.29 | 0.29 | 0.28 | 0.27 | 0.26 | 0.27 | 0.29 | 0.38 | 0.51 | 0.61 | 0.66 | 0.70 | 0.71 |
| ~95 confidence limits on depletion |  |  |  |  |  |  |  |  |  |  |  |  |  |
| lower | 0.25 | 0.24 | 0.23 | 0.21 | 0.2 | 0.2 | 0.21 | 0.28 | 0.37 | 0.44 | 0.48 | 0.5 |  |
| upper | 0.34 | 0.34 | 0.34 | 0.33 | 0.32 | 0.34 | 0.36 | 0.48 | 0.65 | 0.78 | 0.85 | 0.89 |  |
| recruits | 15080 | 6555 | 7584 | 12569 | 153415 | 3708 | 15148 | 23831 | 14082 | 25895 | 7647 | 6645 | 32063 |
| ~95 confidence limits on recruitment |  |  |  |  |  |  |  |  |  |  |  |  |  |
| lower | 8031 | 1399 | 2723 | 4260 | 104994 | 0 | 9036 | 14220 | 8380 | 15385 | 4546 | 3959 |  |
| upper | 22095 | 11691 | 12465 | 20936 | 202966 | 8023 | 21322 | 33540 | 19842 | 36511 | 10779 | 9358 |  |
| ABC | 4000 | 4000 | 4000 | 3400 | 3724 | 3681 | 2700 | 2700 | 2700 | 2700 | 2700 | 2700 | 2700 |
| OY |  |  |  |  | 3724 | 2000 | 2000 | 2000 | 2000 | 2000 | 2000 | 2000 | 2000 |
| total catch | 2021 | 1870 | 2110 | 1430 | 977 | 499 | 517 | 329 | 21 | 236 | 192 | 127 | n/a |
| expl. rate | 0.119 | 0.114 | 0.133 | 0.098 | 0.072 | 0.037 | 0.028 | 0.014 | 0.001 | 0.008 | 0.006 | 0.004 | n/a |
| SPR | 0.40 | 0.37 | 0.45 | 0.55 | 0.72 | 0.72 | 0.84 | 0.99 | 0.93 | 0.95 | 0.96 | 0.97 | 0.97 |

## Research and Data Needs

Additional investigations into the catch history should be made, ideally as a part of a greater reconstruction of historical rockfish landings done comprehensively across all species.

Greater exploration of methods for modeling time-varying growth as influenced by environmental factors should be a key research area for future assessments, and would benefit greatly from data from historical (triennial trawl) and recent (NWC combined) surveys.

The effects of spatial management measures on patterns of vulnerability and selectivity over time have not been evaluated, and would benefit from generic simulation studies of the consequences of spatially explicit management measures to the basic assumptions of stock assessment models.

## Regional Management Concerns

There are insufficient data to consider spatial structure in the model. Although the CalCOFI time series (which was not used in the final model) might suggest greater relative depletion south of Point Conception, this time series has some unusual characteristics that undermine its utility as an index of abundance. As there is only very limited fisheries dependent information in this region, and only a very short (four years) time series of fishery independent information (with low sampling density), there is insufficient information to assess regional concerns. However, as abundance appears to drop sharply towards the U.S./Mexico border, transboundary issues are minimal for this stock.

# Status of the Chilipepper rockfish, Sebastes goodei, in 2007 

## Introduction and distribution

Chilipepper rockfish (Sebastes goodei) are described as an elongate fish with reduced head spines similar in appearance to both shortbelly rockfish (at smaller sizes, although shortbelly tend to be slimmer) and bocaccio rockfish (bocaccio tend to have larger mouths). The latin name honors that $19^{\text {th }}$ century ichthyologist and fisheries biologist David Brown Goode (Love et al. 2002), while the common name was derived from the observation that long strings of these bright red fish resemble a string of drying chilis (Davis 1978). They have been one of the most important commercial target species in California waters since the 1880s, particularly in this core region, and were historically an important recreational target in Southern California waters. Their importance in recreational fisheries in northern waters followed the movement of recreational fishing effort to deeper waters in the 1970s and 1980s, prior to which catches were apparently minimal.

The distribution is described in the literature as ranging from Queen Charlotte Sound (British Columbia) to Bahia Magdalena (Baja California Sur)(Westrheim 1965; Eschmeyer 1983; Love et al. 2002), however they are uncommon north of Cape Blanco (Oregon) and south of Punta Colnett (Baja California Norte). The region of greatest abundance is found between Point Conception and Cape Mendocino, California. Alverson et al. (1964) reported only trace catches of chilipepper rockfish in resource surveys conducted in the 1960s off of Oregon and Washington, all of which was noted between 100 and 150 fathoms. Adult fish tend to be most abundant in large schools between 100 and 300 meters, often in midwater. Settled juveniles tend to be found in shallow water, and move to greater depths with size and age. Love et al. (2002) describe the habitat of adult schools as including boulder fields and other high relief substrata, and occasionally low-relief cobblestones.

Like all rockfish, chilipepper are primitively viviparous and bear live young at parturition. They copulate during September-October and extrude their larvae from December-February (Wyllie Echeverria 1987). Larvae and juveniles have an extended pelagic phase of about 150 days, consequently the spatial dispersal of larvae likely links recruitment among areas. Field and Ralston (2005) evaluated spatial patterns in recruitment variability based on regional catch at age data and concluded that recruitment is largely synchronous throughout most of the range of chilipepper in the California Current between Cape Blanco and Point Conception, although there were insufficient data to evaluate chilipepper south of Point Conception. Wishard et al. (1980) conducted the only known study of stock structure, from samples collected between 34 and 40 N , and they concluded that chilipepper was unusual in its very low levels of allozyme variability, with no suggestion of population substructure. In an extensive review of phylogenetic relationships among Sebastes, Hyde and Vetter (2007) found that chilipepper rockfish were most closely related to both shortbelly (S. jordani) and bocaccio (S. paucispinis) rockfish, with a lineage that dated back approximately 6 million years.

Although there are no quantitative food habits studies of this species, they are described as midwater foragers, with euphausiids, forage fishes (such as anchovies, Pacific hake, and
mesopelagic fishes), and small squids among key prey items (Love et al. 2002). Pelagic juveniles are preyed upon by a wide range of predators, including seabirds, salmon, lingcod and marine mammals. Larger piscivorous fishes, marine mammals, and in recent years jumbo squid are among the predators of larger adults.

## Growth and Maturity

The most recent assessment (Ralston et al. 1998) provides a summary of previous estimates of chilipepper growth parameters, dating back to Phillips (1964). Age and length data were available for over 16,000 males and 30,000 females, however most of these data were fisheries derived. The external fits are shown (Figures 1a and 1b), comparable parameter values estimated internally from an early draft of the model that included conditional catch-at-age information were used in the base model as fixed parameters. As the previous assessment reported significant variation in size at age, potentially confounded with changes in selectivity over time, time varying growth was explored in some detail for this assessment. Figures 2a and 2b shows the average size at age from the commercial trawl fishery over time, as both annual averages and a 3 -year running mean for fish ages 3 , 6 and 9 . These data suggest a gradual decline in size at age from the late 1970s and early 1980s, with a slight bump in the late 1980s, followed by low values in the 1990s and increasing values since 1999. Consequently, changes in the size at age were explored in this model.

Weight at length was estimated separately for males and females, based on data from 233 females and 220 males for which this information was collected during triennial trawl surveys (Figures 3a and 3b). Although maturity could vary both as a function of length and age, for the purposes of this model, maturity was fit with a logistic regression model as a function of length (Figure 4).

## Natural Mortality

In the last chilipepper stock assessment, Ralston et al. (1998) estimated sex-specific values of natural mortality internally; for females the model estimated a natural mortality rate of $0.223 / \mathrm{yr}$ and for males the model estimated $\mathrm{M}=0.253$. Prior to that assessment, Rogers and Bence (1993) assumed a natural mortality rate of 0.15-0.20, and Henry (1986) had used a value of 0.20. In earlier assessments, the maximum observed age of chilipepper was 35 years, which corresponds to an estimate of $\mathrm{Z}=0.12$ from Hoenig's (1983) equation. However, Ralston et al. (1998) also note that application of the Jensen (1997) equation to the estimated K values obtained for the two sexes yielded M values in the range of $0.28-0.34$. In order to evaluate Beverton's (1992) approach relating the age at $50 \%$ maturity to the natural mortality rate, we compiled data on age at maturity and estimated natural mortality for all West Coast groundfish stocks as well as four Gulf of Alaska rockfish stocks (Figure 5). The resulting relationships were used to develop point estimates of natural mortality for chilipepper rockfish, based on an estimated age at $50 \%$ maturity of 2.5 yr . These provided point estimates of M of 0.17 based on all West Coast and Gulf of Alaska Sebastes ( $\mathrm{n}=15$ ), and 0.24 based on all West Coast groundfish ( $n=22$ ). Despite the fact that each relationship had an $R^{2}$ of $\sim 0.75$, no attempt was made to develop confidence intervals or informative priors based on any of these estimates, in keeping with the guidance developed in the Harvest Policy workshop. This report emphasized the
significant limitations associated with deriving a relationship between M and life history characteristics, and stressed that in the absence of a genuine scientific advance in estimating natural mortality rates, continuity in assumptions regarding natural mortality has a greater priority than any preferences developed by assessment authors.

Despite this, the natural mortality rate used in the last assessment was considered to be too high by the STAT team and the STAR Panel during the review of this stock assessment. Part of the rationale for this likely includes the age data for 1978-1981 that were used or considered in this model, which suggested a greater proportion of older fish in the early years of the fishery. Based on model estimates and model profiles of alternative natural mortality rates conducted prior to and during the stock assessment review, $M$ was fixed at 0.16 for females, and 0.202 for males. The higher natural mortality rate estimated for male chilipepper is somewhat unusual given the assumptions of a higher natural mortality rate for older females estimated or assumed in assessments of canary, black and yellowtail rockfish.

## Aging Precision

As surface ageing often underestimates ages of older individuals, the 1980 and 1981 age data (which were originally surface read) were not included in the 1998 model. These samples were re-aged using break and burn methods, and samples from 1978 and 1979 were also aged using break and burn methods, these data are now included in the model. The ages available for four years of the triennial trawl survey were all surface read and are no longer available (to re-read and evaluate for a potential bias correction), and consequently these too are not used in this model. The precision of the age determination process was measured by both comparing the independent readings of two age readers of samples collected in 2004 ( $\mathrm{n}=95$ ), as well as comparing independent readings by the same reader ( $\mathrm{n}=97$ ), as reported in the 1998 assessment). The standard deviation by age for each double read was estimated, and as there was no evidence of bias or of an increasing CV with age, a constant CV based on pooling the two samples was used to project the standard deviation by age in the aging error matrix. However, the precision could be overestimated as the high agreement at older ages could also be due to the small sample sizes, as most fish with two reads were less than $\sim 7$ years of age.

## Regulatory History

Chilipepper have long been an important element of California fisheries, however with the exception of excluding foreign fishing effort from the U.S. EEZ in the late 1970s, management actions were modest (and usually general to all rockfish and other groundfish) prior to the implementation of the Groundfish Fishery Management Plan in 1982. When the Groundfish FMP was implemented, management for the groundfish trawl fishery was based on individual vessel trip limits, which were set at 40,000 lbs per trip on the Sebastes (all rockfish species) complex. These limits were maintained until 1991, when they were reduced to 25,000; in 1993 the trip limit system was revised from daily to biweekly trip limits, which were set at $50,000 \mathrm{lbs}$ (south of Cape Mendocino). The trip limit regime continued to evolve in their absolute amounts and temporal duration (monthly, bimonthly) throughout the 1990s, with a general trend towards lower limits as conservation concerns arose for other rockfish species (particularly bocaccio rockfish in the region south of Mendocino). Consequently, landings for chilipepper rockfish
declined significantly during this period, falling well below the ABCs and OYs implemented by the PFMC. Figure 6 summarizes the major management actions for chilipepper (and rockfish regulations more generally), Table 1 summarizes the ABC and OY values adopted by the Council and the subsequent estimates of total catches (including discards), while Appendix A provides an extensive summary of the management actions relevant to chilipepper rockfish since the implementation of the FMP.

For the current management cycle, the Pacific Fishery Management Council has specified status quo alternatives for chilipepper rockfish south of Cape Mendocino for 2007 and 2008 (ABC 2,700; OY 2,000). Chilipepper rockfish within the Eureka INPFC region are managed within the minor rockfish North category (an assumption that they account for approximately 32 tons of that OY has been made). Recent catches are well below these levels due to the constraints imposed by the rockfish conservation areas, and low trip limits in open areas implemented to ensure low bycatch rates of rebuilding species that co-occur with chilipepper (particularly bocaccio, but including canary, widow, cowcod and yelloweye). Although proposals have been repeatedly developed that would facilitate accessing the existing chilipepper OY, a paucity of bycatch data in southern areas for many gear types as well as coastwide bycatch constraints have repeatedly prevented liberalization of trip limits or approval of Experimental Fishing Permits (EFPs) in recent years.

## Commercial Fisheries Landings

Chilipepper have historically been one of the most important rockfish species in California fisheries. Commercial landings from 1978 to the present were obtained directly from the California Cooperative Survey (CALCOM) database using expansion procedures from sampling commercial market categories (Pearson and Erwin 1997). Chilipepper have been landed primarily in chilipepper, bocaccio and mixed rockfish market categories. In a recent evaluation of market categories of the commercial fishery, chilipepper rockfish scored high on an index of reliability (D. Pearson, NMFS/SWFSC, pers. comm.), and landings from 1978 to the present are consequently considered to be accurate.

Landings of rockfish (all species combined) in California were recorded in CDFG Fisheries Bulletins from 1928 through 1978 by region (Del Norte/Eureka, San Francisco, Monterey, Santa Barbara, Los Angeles, and San Diego), shown as Figure 7a and 7b (digitized summaries of these catches can be queried online http://las.pfeg.noaa.gov:8080/las_fish1/servlets/dataset). We used these landings to derive catch estimates for the early time period. For the period prior to 1928, we used rockfish landings reported by Sette and Fiedler (1928), who report landings irregularly from 1892 through 1926. Landings are interpolated between unreported years, and assumed to be zero prior to 1892. Although paranzella trawling (and later otter-board trawling) have been an important source of marine fisheries landings in California since 1876, most of the trawl catch in early years was composed of flatfish (petrale and English sole) fished over soft bottom (Clark 1936). Wolford (1930) describes hook and line, set lines, long lines, and hand lines as being the primary gears used in rockfish fisheries prior to World War II, and Phillips (1949) estimates that only about $5 \%$ of the early rockfish landings were from trawl-caught fish. Thus, we assume $95 \%$ of all rockfish landings prior to 1943 to be hook and line caught, and 5\% to be trawl caught.
Table 2 provides estimates based on Sette and Fiedler from 1880 to 1927. Table 3 provides the

CDF\&G Fisheries Bulletin summaries of total rockfish catch by region, and the assumed proportion of these catches by gear type, and the assumed proportion of each catch estimated to be chilipepper rockfish by region based on the following analysis.

There is little in the way of species composition information for these early fisheries, however Phillips (1939) reported on the species composition of rockfish from the Monterey wholesale fish markets between April 1937 and March 1938, in which $30.8 \%$ of the landings by weight were chilipepper rockfish (with $39.4 \%$ bocaccio and $7.9 \%$ yellowtail rockfish). Monterey Bay ports were the most productive along the coast during that period, accounting for $51 \%$ of all landings between 1936 and 1940, with San Francisco accounting for another 20\%. Consequently, as landings of rockfish in the Eureka area were minimal until the introduction of the trawl fishery in the 1940 s, we assume that $30.8 \%$ of California rockfish landings from Santa Barbara north to the Del Norte/Eureka area were chilipepper rockfish until the introduction of the balloon trawl fishery in 1943. Based on the earliest estimate of species composition in the Del Norte/Eureka area (see below), we assume that $5.7 \%$ of rockfish landed in this region were chilipepper (note that landings in this region were minimal until 1943). The species composition of southern California rockfish fisheries is not quantified in historical accounts, however chilipepper are cited by Wolford (1930) as being the "second most important rockfish in southern California rockfish fisheries (vermillion are described as the "most important" and bocaccio as "important"). Similarly, Roedel (1948) described chilipepper as "one of three leading Southern California species" (along with vermillion and bocaccio). Even earlier, Jordan and Evermann (1898) had described chilipepper as being "taken in abundance about the Coronados Islands, Santa Catalina, and the Cortez Banks." The 1930s was a period in which landings in Los Angeles and San Diego regions dominated southern California landings, as the Santa Barbara region, including Morro Bay, accounts for only 12\% of Southern California landings during this period. Consequently, chilipepper seem to have been historically a significant component of hook and line fisheries throughout Los Angeles and San Diego regions, and we assume that chilipepper accounted for $20 \%$ of all Los Angeles and San Diego region rockfish landings from 1928 through 1963.

In 1943 the balloon trawl was introduced to northern California waters from Oregon, in association with a strong market for frozen rockfish by the military to support the war effort. Trawl gear rapidly surpassed hook and line gear in accounting for the majority of California rockfish landings, particularly in the northern ports of Eureka and Fort Bragg (Scofield 1948; Phillips 1949). Thus, through 1940 we assume that $95 \%$ of chilipepper were hook and line caught, and we assume that by 1944 some $90 \%$ of the total rockfish (and subsequently, chilipepper) catch was trawl (based on the percentage trawl in later years, see below). Between 1940 and 1944 we assume that rockfish catches were 25,50 and $75 \%$ trawl in 1941, 1942 and 1943 respectively. Trawl caught rockfish continued to comprise approximately 85 to $90 \%$ of all rockfish landings throughout the 1960s and early 1970s, and we used the ratio of trawl caught rockfish reported by Nitsos (1965), Orcutt (1969) and Gunderson et al. (1974) to total rockfish landings from CDFG bulletin to apportion the chilipepper catch by gear from 1953 through 1977 based on these observed fractions and interpolation between unobserved periods (Table 4).

To assess the fraction of trawl caught rockfish that were chilipepper, we relied on the very sparse species composition reports included in Nitsos (1965) and Gunderson et al. (1974). Nitsos
reported the 1962-1963 species composition by port complex for most California ports (as trawling was then prohibited in nearshore waters south of Santa Barbara, no species composition was reported for that region), these are reported in bold font in Table 4, and these values were used for both trawl and hook-line fisheries from 1942 through 1963 (during the period in which trawl landings dominated). Gunderson et al. (1974) also reported trawl species composition for the year 1973. For all intervening years between 1963 and 1978, the fraction of the catch that was chilipepper rockfish was interpolated between these observed catch compositions and the CalCOM estimates for 1978-1979. Accounting for the catch composition in the Los Angeles and San Diego regions since 1963 is tricky, as most landings were hook and line in this region and no hook and line data for this period are available. However Gunderson et al. (1974) described chilipepper as accounting for $26.4 \%$ of the Conception area trawl catch in 1973, and chilipepper continued to be described as important to Santa Barbara hook and line fisheries during this period, although they were not as valuable as the more brightly colored vermillion and other species (Love 1991; Kronman 1999). Consequently we assume the Gunderson et al. (1974) catch proportion for all fisheries throughout Santa Barbara, Los Angeles and San Diego; and interpolate catch proportions from $20 \%$ in 1963 to $26.4 \%$ in 1973. From 1974 to 1977 we interpolate the $26.4 \%$ in Southern California fisheries reported by Gunderson to the CalCOM estimates of 2\% of Santa Barbara, 8.1\% of Los Angeles, and 2.2\% of San Diego rockfish catches. There is clearly a great deal of uncertainty over whether this decline is an artifact of the means by which catches were reconstructed, or reflects changes in abundance or target fisheries, and we acknowledge that the relative importance of chilipepper in Southern California fisheries throughout this period is highly uncertain. For Oregon landings, PacFIN estimates were used for landings from 1981-present, and for 1963-1980 estimated are based on Douglas (1998), who report minimal (and sporadic) chilipepper landings on the Pacific Ocean perch and other rockfish market categories. We assume landings were negligible in Oregon waters prior to 1963. The resulting estimates of chilipepper catch are reported in Tables 5-6 and Figures 8a and 8b.

An alternative catch stream for the period between 1953-1977 was also developed, based on retroactively applying market category species compositions from the 1978-1984 period to CDFG landings data by market category extending back to 1953 (D. Pearson, pers. com.). Based on recently digitized CDFG landings information by block and market category, and applying the species composition for market categories from the 1978-1983 period, the catch of chilipepper rockfish was reconstructed from the period 1953-1968, and CalCOM reconstructions from 1969-1977 were used based on Pearson (in prep). The corresponding total catch estimate is compared to the earlier reconstruction in Figure 9. As these values differed only modestly, the first catch stream was used in the base model, to maintain consistency with the approach used to estimate landings prior to 1953.

Prior to the STAR Panel meeting, but following the distribution of the draft assessment, the STAR Panel Chair (Dr. David Sampson) pointed out that records of rockfish catches (at the genus, not species level) by gear and by region were also available for much of the historical period, as published in Bureau of Commercial Fisheries Reports. A subset of the relative proportion of catch by gear and by region was developed from these records, which reflect strong geographical differences in historical gear type use, with a shift to primarily trawl-caught rockfish in the north to almost exclusively hook and line caught rockfish in the south. Figures 10a-10e show the relative rockfish catch by gear type and district for select years in this period,
however there was insufficient time to re-define the initial catch statistics in a timely fashion for consideration in the final model. Modest changes between the proportion of catch by gear type are not anticipated to have a major influence on the model results. A number of STAR Panel reviews have lamented the lack of a comprehensive reconstruction of historical rockfish catches by species for California waters, similar to that of Rogers (2003) for foreign fishery catches, and this remains a key stock assessment need. Currently, California fish ticket information with associated market category and CDF\&G block number is in the process of being digitized for the period 1928-1977, and a comprehensive rockfish historical catch reconstruction will benefit greatly from the results of this effort. Finally, comparison of the catch estimates used in this model to those used in Ralston et al (1998) are presented as Figures 11a through 11d, which show that although some catch estimates have varied modestly over time, the time series track each other very closely.

## Commercial Discards

Heimann and Miller (1960) reported a bycatch rate of approximately 0.8\% for chilipepper rockfish taken in 64 bottom trawls off of Morro Bay, California between August 1957 and July 1958. Similarly, Heimann (1963) reported extremely low discard rates for chilipepper rockfish, of approximately $0.4 \%$ for a series of 19 intermediate depth tows made between Pigeon Point and Point Sur, California in 1960. Aside from these observations, there is essentially no data available on potential discard rates for any but the most recent years for chilipepper rockfish. As chilipepper are a desirable market category, discards have been assumed to be negligible in past assessments (Ralston 1998), and with the exception of the recent years in which regulatory changes have resulted in high discard rates, we will continue with that assumption. The estimated commercial discard rates for chilipepper and bocaccio in the Monterey and Conception INPFC areas, derived primarily from observations of the trawl fleet, were 46\%, $11 \%, 70 \%$, and $65 \%$ from 2002 through 2005 respectively (as a \% of discard+landed). Catches for all gear types for these four years were adjusted proportionately, with the $65 \%$ discard rate from 2005 carried over into 2006 (based on Hastie and Bellman 2006, and comparable reports). As the total landings have been minor relative to historical landings in this period, adjustments to this rate for recent years would not be expected to have major consequences to the model results.

## Recreational Fishery Landings

Recreational fishing effort in California for fishes other than big game fish such as tunas and salmon was relatively modest in California until about 1928, when Commercial Passenger Fishing Vessels (CPFVs) popularized recreational fishing (Scofield 1928; Croker 1940; Young 1969). Initially, most effort was in the waters of the Southern California Bight, however party boat fisheries soon became popular in Monterey, and although these fisheries were suspended during World War II, effort increased rapidly shortly after the war ended. CPFV captains have been required to submit logbooks detailing catches since 1936, in which species resolution is typically low (typically only "rockfish" is recorded, although some rockfish targets such as cowcod were usually identified to species). Reported CPFV catches in numbers of fish for most years between 1936 and 2000 were available from the CPFV database (Hill and Schneider 1999), with missing years and region-specific information filled in from Young (1969) and Best (1963). Although this database has no estimate of private vessel catches or other fishing modes (shore,
pier, neither of which catch chilipepper), and compliance rates have typically been less than $100 \%$, this is the only source of recreational catches prior to 1980, and catch estimates are based on this information as tuned to more recent estimates.

For 1980 through 2006, catches in both numbers of fish and weight of fish were obtained from the RecFIN database. RecFIN data are based on Marine Recreational Fisheries Statistics Survey (MRFSS) catch estimates, which are based on a combination of angler field surveys and randomized telephone surveys from 1980 through 2006 (with a hiatus from 1990 through 1992), with four primary fishing modes; CPFV, private vessel, pier, and shore (only the first two catch notable quantities of chilipepper). Spatial resolution of these catch estimates is limited to northern and southern California (north and south of Point Conception). Table 7 provides RecFIN catch information for chilipepper rockfish in northern and southern (south of Conception) recreational fisheries in numbers, total weight, and average weight from 1980-2006 (with the years 1990-1992 interpolated) by mode (CPFV and private/rental only). Figure 12 also shows the percentage of all rockfish that were estimated to be chilipepper rockfish by region and mode from RecFIN data as well as CDFG observer program data collected from 1975-1978 and 1986-1989 in the south, and 1987-1998 in the north. These percentages were critical to reconstructing historical estimates of chilipepper catches in recreational fisheries.

The reconstruction of recreational catches prior to 1980 is highly dependent on assumptions about the spatial development of this fishery to deeper water over time, particularly in the north, (reconstructions were made separately both north and south of Point Conception). North of Point Conception, it is widely held that CPFV fisheries moved from nearshore habitat and target species to deeper and deeper waters over time. Miller and Gotshall (1965) report on the landings, weights, and species composition of northern California recreational fisheries from 1957 through 1961, in which blue, yellowtail, olive, and bocaccio rockfish were among the most important (together accounting for $\sim 65 \%$ of the total catch by number). Chilipepper were reported in only trace amounts, accounting for $0.321 \%$ of the total observed CPFV rockfish catch (2165 out of 674,678 rockfish reported), and were even more scarce in the private/rental boat (skiff) fishery, where they accounted for $0.004 \%$ of observed rockfish (7 out of 157,257 rockfish reported). Similarly, Heimann and Miller (1960) described chilipepper as being a very minor species in Morro Bay party boat fisheries in the late 1950s; this fleet too was clearly targeting nearshore assemblages (blue, olive, yellowtail, and vermillion rockfish comprised over $80 \%$ of the catch). However, chilipepper appear to have been sporadically important, at least in the Monterey Bay area recreational fisheries, in the years between this report and the RecFIN time period; Mason (1995) describes wide fluctuations in the CPFV catches of deepwater rockfish, with chilipepper being a key recreational species in 1962, 1964 and 1977-1978. As no species composition data are available, nor is it clear whether this reflected local or coastwide shifts in fishing spots and methods, we interpolated the percentage of rockfish landings (in numbers of fish) thought to be chilipepper from the 1957-1961 point estimate ( $0.321 \%$ ) to the 1980-1982 RecFIN average (3.84\%). This in turn was scaled upwards by the ratio of RecFIN estimated CPFV catches over logbook CPFV catches from 1980-1982 to develop an expansion factor for the historic CPFV fishery (1.87), which provided an estimate of the historical CPFV (and other fishery modes) total rockfish catches in numbers (Table 8; Figure 13). Finally, as the average weight of chilipepper reported in Miller and Gotshall ( 1.2 kg ) was significantly greater than the average weight of fish reported by RecFIN in the 1980-1982 period ( 0.72 kg ), we interpolated
the average weight between these periods to arrive at the tonnage of total catch. To account for the presumably modest CPFV chilipepper catches in the north prior to 1957, we assume that chilipepper catches were $0 \%$ of the total rockfish catch at the initiation of the fishery in 1928, and interpolate from 0 to $0.331 \%$ in 1957. As the private boat fishery represented a trivial source of mortality in both the 1957-61 period and the 1980-82 period, we do not account for possible private vessel landings in the north prior to 1980.

For southern recreational fisheries, we used RecFIN data from 1980 through 2006, an expansion factor for historical CPFV logbook data as was done in the north (estimated at 1.98), and supplemented with observations of the percentage of the CPFV catch listed as chilipepper from the 1975-1979 onboard observer program. As this program tended to record a higher (and less variable) percentage of chilipepper rockfish relative to the total rockfish catch, we used the average proportion of the total rockfish catch observed to be chilipepper from the 1975-1979 observer data and the 1980-1982 RecFIN data to interpolate the fraction of historical catches that were chilipepper, assuming a ramp up from 0\% chilipepper in 1928 (when CPFV fishing began, presumably with a focus on shallow water targets) to $11.3 \%$ in 1974. As chilipepper have long been described as an important recreational fish in Southern California (Wolford 1930; Roedel 1948; Davis 1977; Love 1991), and tend to be more important over deeper reefs, this is a reasonable approximation of recreational fisheries development. As private vessel landings of chilipepper estimated by RecFIN were significant in the early 1980s (estimated at 38,000 fish per year between 1980-1982), we assumed that private vessels began catching chilipepper in the post-world war II era, and interpolated landings from 0 in 1947 to 38,000 fish per year in 1979. As the average weights of chilipepper in the early 1980s were comparable in the north and south in the RecFIN database, we used the same average weight estimated for central California fisheries (above) for southern California fisheries.

The total estimated catches in the recreational fishery are shown as Figure 14, the total catches by all fisheries are shown in Figure 15, and these catches by fishery are also shown relative to catches estimated in the 1998 assessment in Figure 10 (referred to earlier). The number of subsamples and length measurements in the RecFIN database are included as Table 9.

## Trawl Logbook CPUE Data

A catch per unit effort index was developed in the last assessment by Ralston et al. (1998), and was included in this assessment in the same form, as management constraints have likely biased the assumptions that would be necessary to update this index. Ralston (1999) further developed the trawl CPUE time series using alternative weighting regimes; these two time series as well as the time series from the 1998 model are presented as Table 10 and Figure 16. The 1998 estimates were assumed to have a CV of 0.10 in the 1998 model, however this CV was largely arbitrary. As the indices developed in 1999 had CVs on the order of 0.25 to 0.35 , and model runs consistently estimated an effective RSME of $\sim 0.25-0.28$ when the initial CV was set at 0.1 , we used 0.25 as the assumed CV.

## Commercial age and length composition data

Expanded length composition data for the three commercial fisheries were extracted from the CalCOM database (Pearson and Erwin 1997) for all years from 1978 through 2006. Length data were pooled into 2 cm groups with accumulator groups representing sizes less than 16 cm and greater than 52 cm . Age data were aggregated into 21 age groups, comprised of ages 1-20 and an accumulator age of 21 and older fish. Age composition data by commercial gear type are shown in Figures 17-19, and length composition data are shown as Figures 20-22. Although earlier years of the fishery had significant proportions of older fish, less than $1 \%$ of all (expanded) fish were older than age 20 (although this fraction was somewhat higher for earlier years in which catch at age data were available). Starting values for multinomial sample for both age and length composition data were based on the number of port samples taken that included chilipepper age structures or lengths, respectively. Table 11 provides the sample sizes and total number of fish by year and gear type used in the expansions.

A comparison of raw (unexpanded) catch-at-length data from port samples that included age information and those that did not suggested some potential discrepancies between the length composition of aged versus un-aged fish, which may have been a (minor) contributing factor to the complications encountered with the conditional catch-at-age data. A more likely complicating factor may have been the approach used to generate the effective sample sizes as well as for tuning the effective sample sizes of the conditional age-at-length data.
Recommendations for future efforts to incorporate conditional age-at-length information, as well as innovative approaches that could be used to link the likelihood components between length frequency and age-length data, are included in the STAR Panel report as well as the recommendations section of this document. As a result of potential biases in the age composition subsampling, the effective sample sizes were set to negative numbers (resulting in a zero emphasis for those combinations in the likelihood function) for the following gear/year combinations; trawl (1978-1979, 1998-2000), hook and line (1998-2002), and setnet (1983, 1992). These data should be revisited for potential bias (by evaluating the expanded, rather than raw, catch at length for both aged and un-aged fish) prior to the next assessment. Additionally, the length frequency data for the 1992 setnet fishery suggested catches of a large number of very large males, which were sufficiently suspect to warrant exclusion of these data from the model.

## Recreational length composition data and CPUE time series

Recreational length data from the RecFIN database were based on a query of coastwide length composition data from March of 2007, and are presented as Figure 23 (northern and southern separate) and Figure 24 (combined). As these data were not associated with sex information, they were included in the model as combined sex length composition data associated with the recreational fishery. In evaluating the potential for developing a CPUE time series for chilipepper rockfish using RecFIN observer data, we found that chilipepper were only recorded in 52 of the thousands of observed trips. Attempting to identify appropriate trips using the approach of Stephens and MacCall (2004) resulted in a subset of nearly 250 trips that could be identified as those in which chilipepper catches were likely, however there were unusual species co-occurrences that lead to this approach being suspect. As chilipepper rockfish tend to only be encountered in deeper water recreational trips, and the depth distribution of recreational effort
has changed markedly over time, RecFIN catch rate data were not evaluated further in this assessment.

The California Department of Fish and Game conducted on-board monitoring of partyboat catches in Northern California from 1987 to 1998, which includes catch, angler effort, size composition of catches, location information and, more importantly, depth information (Deb Wilson-Vandenberg, CDFG, pers. comm.). Between 1987 and 1998 some 2267 recreational fishing trips were observed from Morro Bay (649) to Eureka and Crescent City (12), however the majority of observed trips originated from Monterey (821), San Francisco (444), and Bodega Bay (269) area ports. CDFG block information, as well as fishing site ( 457 sites) and the maximum and minimum observed depth information (ranging from 2 to 150 fathoms), was also available for all trips. Locations represented 68 separate CDF\&G blocks, but $90 \%$ of the trips took place in just 27 of these blocks. Between 1987 and 1998 most of the trips were in the 20 to 60 fathom range, however there was a slight increase in the percentage of trips in the 0 to 20 fathom range and a slight decrease in the percentage of trips in the 60 to 100 fathom range. Overall, the latter represented less than $15 \%$ of all trips observed.

The total number of observed trips, binned by the average depth for the trip, for each year are given in Table 12. Chilipepper were ranked third in terms of the total number of rockfish caught in observed trips (27,690 out of 313,752), after blue and yellowtail rockfish, however they were ranked $21^{\text {st }}$ in terms of the most frequently occurring species. This seems to be a consequence of fishing location. Chilipepper were frequently encountered in trips that fished at greater depths, occurring in only $1 \%$ of trips that fished less than 40 fathoms, but in $68 \%$ of trips that fished in 60 to 80 fathoms and $92 \%$ of trips that fished greater than 80 fathoms. The number of chilipepper caught per year and depth bin are included as Table 13. Clearly, depth is an important variable in the GLM, although when site-specific location information was explored as a variable, the variance explained by depth decreased substantially. This reinforced the decision to exclude RecFIN data. Consequently, due to concerns discussed during the STAR Panel review regarding possible impacts of changing depth strategies over time, all trips at depths greater than 80 fathoms were excluded from the final model. We used the average depth per location, binned into 20 fathom depth intervals for the GLM. Ultimately, trips taken at less than 20 fathoms average depth were also excluded due to the very low frequency of positives for chilipepper. For location information, we considered site specific information, CDF\&G block information, and port-group information as possible factors in exploratory models. All explained a moderate fraction of the variance, and all resulted in very similar results with respect to year effects, however using site as a variable resulted in the loss of a substantial number of records.

The logistic regression method of Stephens and MacCall (2004) was also evaluated to obtain a subset of the trip data that would be appropriate for calculating chilipepper CPUE from the observer data. This method uses the species composition from each trip to determine whether chilipepper rockfish were likely to have been encountered on that trip, however this method is more commonly used for datasets in which location information is unavailable or unreliable (such as sampling and interviews conducted at the end of a fishing trip, used for MRFSS dataseries). One reason for this was to evaluate whether this approach resulted in different inferences with respect to trend, and to evaluate whether the resulting species coefficients from this approach were consistent with those obtained from a similar effort using the MRFSS data.

The top 50 species in frequency of occurrence were extracted, chilipepper were separated as being the target species, and species that co-occurred with chilipepper less than two times were excluded (four species). The remaining 45 species served as potential explanatory variables. Logistic regression of chilipepper presence/absence on categorical presence/absence of these explanatory species provided predicted probabilities that chilipepper would be taken on a trip, given the other species that were taken on that trip. The resulting species associations (coefficients from the logistic regressions) are shown in Figure 25. The threshold probability for inclusion in the selected set was set at 0.35 as this was the probability that resulted in the lowest average CV of the annual indexes. However, the results of using the filtered dataset relative to the entire dataset were nearly identical (discussed below), as the logic behind the filter was to provide proxy information for habitat (area, depth) in datasets without data on these factors. When location and depth information are included, the filter is essentially unnecessary.

Consequently, the final model used all of the available trip information, the year effects are the relative CPUE index (Figure 26), with precision estimated using a jackknife procedure. The other fixed effects were block information (11 blocks with sufficient data, Figure 27) and depth (three bins, 20 to 39, 40 to 59, and 60 to 79 fathoms, Figure 28). A large number of sensitivity runs suggested highly similar, if not virtually identical, results when either higher resolution (site-specific) or lower resolution (port group) location information was used, as well as month or season, or other changes in the resolution of these bins was altered. The AIC values for a suite of models are reported in Table 14, which demonstrates that year, depth, block and season information contributed to an improved model fit. Although the results varied only modestly, the AIC also suggested that a gamma error distribution fit the data better than a lognormal distribution for the base models. Furthermore, the resulting trend when the Stephens/MacCall filter was developed and used to filter trips was nearly identical to the trend without this filter when all trips positive for chilipepper or with a threshold of 0.35 or above were used. The coefficient of variation (CV) estimated in the jackknife routine was also very similar with all of these runs, and between the gamma and lognormal error distribution, although the CV was considerably greater when depth information was excluded.

Length frequency information from chilipepper measured in the observer program was converted from total length to fork length, using the conversions provided by D. Pearson (pers. com.), where
$F l=0.977 * T L-0.977$

The resulting length compositions by year, for fish caught within the depth ranges used to develop the relative abundance index, are shown in Figure 29. The number of trips in which chilipepper were caught was used as the sample size in the length composition data. As sex information was not included, the resulting length frequencies were used in the model with the unknown gender code. These data suggest that the high value in the index during 1987-1988 represented the abundance of the 1984 year class, which is identifiable in other age and length time series. As this age class grew, it likely moved into deeper water, consistent with the shift to greater depths with size observed in the triennial length composition data and consistent with similar ontogenetic movement for many other rockfish and groundfish. Similarly, the increase in abundance in 1992 may have been a function of a relatively strong 1988 or 1989 year class. This
also suggests that a dome-shaped selectivity curve is likely to be appropriate for these length data, given the changing spatial distribution of animals with size.

## Triennial Trawl Survey

A primary source of fishery independent information for most managed and assessed groundfish species in the California Current is the West Coast triennial trawl survey conducted between 1977 and 2004 (Weinberg et al. 2002). As the general consensus from recent data workshops has been to exclude 1977 data, we obtained both stratum-specific area swept biomass estimates and haul-specific survey data from 1980 to 2004 (M. Wilkins, AFSC, pers. com; B. Horness, NWFSC, pers. com), both of which were generated after excluding bad performance tows and "waterhauls," in which few benthic organisms were noted (Zimmermann et al. 2001). Tow specific CPUEs from this survey by year are shown in Figure 30, which also illustrates the variation in the latitudinal range of this survey over time (These Figures include a "cap" on the relative size of the largest tows, to maintain a constant scale across all of the Figures). Areaswept biomass indices by INPFC area and depth strata are presented as Table 15. To develop a consistent area-swept biomass index that represented all years, we compiled biomass estimates for all stratum between $36^{\circ} 48^{\prime} \mathrm{N}$ and $43^{\circ} 00 \mathrm{~N}$ (55m-366 m depth)(Figure 31).

Another comparable index was developed by T. Helser (NWFSC, pers. com.) using the Generalized Linear Mixed Model (GLMM) approach described in Helser (2003) and Helser et al. (2005). This model uses depth strata and latitude (or INPFC latitude proxies) as fixed effects, and vessel as a random effect. This index more explicitly accounts for the area of the given strata, as well as integrates uncertainty across both the proportion positive and the positive catch rate indices (such that both the variance due to vessel and residual variances are estimated, with the assumption of a log-normal error variance assumption for the positive observations). Point estimates of biomass and the associated CVs are based on the median of the marginal posterior density from MCMC, however to develop these estimates the model needs a high density of positive tows per strata (at least 2, preferably 3 for each year, depth, latitude combination). The strata used for this index were from $34.5^{\circ} \mathrm{N}$ to $38^{\circ} \mathrm{N}$, and from $38^{\circ}$ to $41^{\circ} \mathrm{N}$. The region north of 41 was excluded due to the rarity of positive tows in that area, inclusion of this area could result in a bias by extrapolating the larger CPUEs observed south of this region. Depth strata were 50 to 155 m , and 156 to 366 m .

As seen in Table 16 and Figures 31 and 32, there is a relatively large difference between the design-based estimate and the GLMM estimates, due primarily to the fact that the mean from the standard approach is heavily influenced by a small number of tows with very large positive catches; the influence of these tows is reduced in the GLMM under the assumption of a lognormal error distribution. This is a common challenge in developing indices of abundance from trawl surveys for semi-pelagic rockfish species with very patchy distributions and often highly specific habitat associations. By contrast, modeling of absolute abundance using design-based versus GLMM approaches tends to produce very similar trends for most flatfish species. Consequently, survey biomass indices are often more appropriately treated as indices of relative, rather than absolute biomass, and both the triennial trawl survey index and the combined survey index are treated in this matter in this assessment.

Length frequencies for the triennial survey were calculated based on standard estimation methods (Dark and Wilkins, 1994), and are presented as Figure 33. Additionally, these data are pooled over all years and shown aggregated into depth bins to demonstrate a clear movement to deeper water with size, as shown for many other Sebastes species (Figure 34). Otoliths collected in 1977, 1980, 1992 and 1995 were surfaced aged, and the samples have since been lost or destroyed; there are no available data with which to bias-correct these estimates and they were consequently not used in the model. The number of hauls was used for the initial effective multinomial sample size in the length compositional data.

## Northwest Center Trawl Survey

Data were provided for area-swept biomass estimates from 2003 to 2006, and associated length frequency compositions, were provided by Beth Horness (NWFSC). A summary of methods used to derive these data is available from O. Hamel (Calculation of summary statistics for the Pacific West Coast upper continental slope trawl survey of groundfish resources off Washington, Oregon and California, in prep, available on request). Catch per unit effort estimates from this survey by latitude and depth are shown as Figure 35. The total area swept biomass estimates ranged from a high of 129,000 tons in 2003 to a low of 69,200 tons in 2006, with the vast majority of the biomass in the shallow stratum of the Monterey INPFC area (Table 17). However, there is no obvious overall trend in the results, particularly given the high uncertainty in the estimates, although there may be a possible suggestion of a decline in recent years. As with the triennial survey index, another comparable index was developed by T. Helser (NWFSC, pers. com.) using the GLMM methods described above for the triennial survey index. The stratification for this index differed, as there was greater spatial coverage in the southern area, and consequently this index estimated biomass for three latitudinal strata, from 32-36 N, 36-40 N, and 40-43 N, with depth strata 50-155, and 156-400. The resulting index is provided in Table 18, which also includes the comparable design-based estimates. As shown in Figure 36, the two indices both appear to be somewhat noisy, with substantial interannual variability from which no obvious trends can be detected; although the GLMM index does seem somewhat better behaved, and may be indicative of a modest population decline over the (short) duration of that time series. The length data for all years, and the age data for 2004, all suggest that the biomass vulnerable to this survey in this period was very strongly dominated by the 1999 year class (Figure 46). Approximately 700 to 1000 chilipepper otoliths have been collected in each year of this survey, however only 850 ages for 2004 were available for this model. These were expanded by the NWFSC and entered into the model as catch at age data.

## Juvenile rockfish survey

The Fishery Ecology Division of the Southwest Fishery Science Center has conducted a standardized midwater trawl survey during May-June aboard the NOAA R/V David Starr Jordan every year since 1983. The primary purpose of the survey is to estimate the abundance of pelagic juvenile rockfishes (Sebastes spp.) and to develop indices of year-class strength for use in groundfish stock assessments on the U. S. west coast. This is possible because the survey samples young-of-the-year rockfish when they are ~100 days old, an ontogenetic stage that occurs after year-class strength is established, but well before cohorts recruit to commercial and recreational fisheries. Chilipepper rockfish are the second most frequently encountered species
in the survey, accounting for $\sim 4.3 \%$ of the total number of rockfish caught from 1983-2006 (shortbelly accounting for just over $85 \%$ of the rockfish identified to species since 1983, excluding shortbelly, chilipepper account for nearly $31 \%$ of the remaining rockfish). This survey has encountered tremendous interannual variability in the abundance of the ten species that are routinely indexed, as well as high apparent synchrony in abundance among the ten most frequently encountered species. Past assessments have used this survey as an index of year-class strength, including assessments for widow rockfish (He et al. 2005), Pacific hake (Helser et al. 2005), shortbelly rockfish (Field et al. 2007) and past assessments of chilipepper rockfish (Ralston et al. 1998).

Historically, the survey was conducted between $36^{\circ} 30^{\prime}$ to $38^{\circ} 20^{\prime} \mathrm{N}$ latitude (approximately Carmel to just north of Point Reyes, CA), but starting in 2004 the spatial coverage expanded to effectively cover the entire range of shortbelly rockfish indexed in this model, from Cape Mendocino in the north to the U.S./Mexico border. Additionally, since 2001 juvenile rockfish data are available from a comparable survey conducted by the Pacific Whiting Conservation Cooperative and the Northwest Fisheries Science Center (spanning from just south of Monterey Bay to Westport, WA; see Sakuma et al. 2007). Comparison of the coastwide data have revealed two types of shifts in the distribution of most pelagic species, in which species characterized by a more southerly geographic range (e.g., bocaccio, shortbelly, and squarespot rockfish) were caught in relatively large numbers south of Point Conception, while species with more northerly distributions (widow, canary, and yellowtail rockfish) were caught in moderate numbers north of Cape Mendocino. The near absence of fish in the core survey area was associated with an apparent redistribution of fish, both to the north and the south, as well as overall lower abundances.

The survey index is calculated after the raw catch data are adjusted to a common age of 100 days to account for interannual differences in age structure. For this assessment cycle, a number of survey indices were developed by S. Ralston (SWFSC) using both the historical (core) survey area and a combined index that uses both SWFSC and NWFSC/PWCC survey data. The indices prepared for chilipepper are presented in Table 19 and shown in Figure 37, and the methods are described in the 2007 stock assessment cycle background materials. One shortcoming of the core index that has been noticed in past assessments has been the failure of the core area survey to capture the magnitude of the 1999 year class for most stocks, the strength of which has since been demonstrated for most recently assessed species. Based on the strong evidence for a very strong 1999 year class, and the recommendations from the juvenile rockfish survey workshop, the core juvenile index was not included in the final model. However, the coastwide juvenile index developed by integrating the results of both surveys in an ANOVA model with year, latitude, vessel, period, and depth effects, was used to inform the relative year class strength for the years 2001-2006. Past assessments have used a power coefficient to transform the index (He et al. 2006), based on the assumption of a compensatory relationship between pelagic juvenile abundance and subsequent recruitment to the adult population following settlement (Adams and Howard 1996). However, due to the short duration of the time series, a power transformation was not estimated for the coastwide survey.

## CalCOFI larval abundance data

Egg or larval abundance data from the California Cooperative Oceanic and Fisheries Investigations (CalCOFI) surveys have been used in stock assessments for a number of commercially important west coast species, including northern anchovy (Jacobson and Lo 1994), Pacific sardine (Conser et al. 2002), bocaccio rockfish (MacCall 2003), shortbelly rockfish (Field et al. 2007) and sheephead (Alonzo et al. 2004). Although a larval abundance index was developed in the first stock assessment for cowcod (S. levis, Butler et al. 1999), this index was not included in the most recent assessment (Piner et al. 2006) out of concerns for the rarity of cowcod in sampled tows. Only a small number of Sebastes larvae can readily be identified to species, including bocaccio, shortbelly, cowcod, splitnose, and chilipepper. Chilipepper rockfish larvae were not identified to the species level in initial plankton sorting efforts. However, morphological characteristics were developed in recent years that allowed for identification, and they were consequently identified in all samples in the CalCOFI core area, and are currently in the process of being enumerated in CalCOFI tows taken in northern stations (W. Watson, SWFSC, pers. comm.). The distribution of chilipepper larvae catches between 1951 and 1969 demonstrates higher catches in northern transects, with catches generally greatest within 75 miles of the mainland (Figures 38 and 39).

As with other indices, we used tow specific information and a delta-GLM approach to derive an index of spawning biomass. Fixed effects in the model included year (fixed to spawning season, such that a year is the October-April spawning period), latitude ( $30^{\prime}$ bins), month (OctoberApril), and distance from shore ( 25 mile bins). These estimates and the associated standard errors estimated from a jackknife routine were used in the model as an index of population fecundity (spawning biomass). Figures 40-42 show the resulting latitude, distance from shore, and month effects; Figure 43 shows the year effects (with standard error) for the resulting model. In general, high levels of abundance were observed throughout most of the 1950s and 1960s, sporadic catches were observed through the 1970s and 1980s (recall that the survey was triennial between 1971 and 1984), and very few larvae were observed in the 1990s. Larvae have been more frequently encountered between 2002-2006. Although the CalCOFI time series is not inconsistent with other data series, the fact that these data are taken from the southern periphery of the stock's range indicates that this may not be an appropriate index of abundance for a coastwide model. Additionally, the lack of estimates throughout most of the period between the early 70s and 2000 (associated with few or no catches of larvae) are troublesome. Consequently, these data were not used in the final model.

## History of Modeling Approaches

Chilipepper rockfish were last assessed by Ralston et al. (1998) using the stock synthesis agestructured model (Methot 2000) for the combined Eureka, Monterey, and Conception areas. The 1998 model began in 1970, but assumed a starting biomass below the unfished equilibrium (based on using the estimated landings from 1960-69 to generate an initial equilibrium population in 1970). The 1998 model also made no assumptions regarding a stock-recruit relationship; recruitment strengths were estimated based on free parameters. Natural mortality rates were estimated internally at 0.22 for females and 0.25 for males. The structure of the data in this assessment is consistent with that assessment, as both assumed four distinct fisheries
(trawl, hook-and-line, setnet and recreational). Landings, age, length, and length-at-age data from these four fisheries were included in the model based on similar expansion routines, age data were limited to 1982-1996 but length data were available from 1980-1996. Estimates of landings changed little between the 1998 and current assessments (Figures 11a-11d, discussed in the catch reconstruction). Similarly, the 1998 model included survey indices from a catch-per-unit-effort index derived from the California commercial trawl logbook data base (which remains unchanged in this assessment), an index of abundance from the triennial trawl survey (which has an extended time series and was been modeled using a different GLM approach than that used in this assessment), and a time series of pelagic juvenile abundance, although the current time series is considerably shorter (2001-2006) than the core index used in the 1998 assessment (19831997). However, the 1998 assessment explicitly described significant changes in mean size at age, which were raised as an important research question, but ultimately applied an approach utilizing time-varying selectivity to fit the length composition data. New indices used in this assessment include the recreational CPUE time series based on CDF\&G monitoring data, and the 2003-2006 NWFSC combined survey index (also modeled using a GLMM approach).

The results of the 1998 assessment suggested that chilipepper were at a moderate level of biomass and were not estimated to be overfished. The 1998 model estimated that spawning biomass had declined from $\sim 48,000$ tons during the 1970's to a low of 22,000 tons in 1987, before increasing as a result of the 1984 year class (which was apparent in both the 1998 and 2006 models). The unfished spawning biomass in the 1998 model was estimated at $58,500 \mathrm{mt}$. The 1998 model estimated that the total exploitation rate ranged from a low of $4.2 \%$ in 1970 to a peak of $19.8 \%$ in 1989, although the exploitation rate had been below the target fishing mortality rate since 1993. Primary sources of uncertainty in the 1998 assessment included the statistical uncertainty associated with the fit of the various data sources to the base model, the conflict between the two principle sources of information (logbook and triennial trawl survey indices), the difficulty in projecting future recruitment for a stock characterized by high recruitment variability, and the difficulty in distinguishing potential changes in selectivity from apparently substantial declines in the mean size at age for fish collected in the post-1993 period.

Prior to the 1998 assessment, Rogers and Bence (1993) conducted a similar length-based assessment (using the length-based version of stock synthesis, Methot 1990) for which the modeled time period began in 1980. Their model included a triennial trawl survey index and a recreational CPUE index, but did not include either a trawl logbook CPUE or a pelagic juvenile survey index. The 1993 assessment also included age and length data from commercial fisheries (modeled as the same four fisheries as in Ralston et al. 1998 and this assessment), including data from fish that had their otoliths surface aged (rather than break-and-burn), and used estimates of natural mortality rate that ranged from 0.15 to 0.20 . Rather than present the results of a single base model, the authors presented the results of a suite of three models, in which the 1992 biomass ranged from 40,000 to $87,000 \mathrm{mt}$, and the equilibrium yield (based on the then proxy for FMSY of F35\%) ranged from 3,941 to $6,729 \mathrm{mt}$. Their general conclusions were that the existing ABC of 3600 mt was sufficient to protect the fishery at the F35\% level, and that raising the ABC above this level could be "somewhat optimistic."

Prior to the 1993 assessment, a stock assessment had been developed by Henry (1986), who used the age composition data in a cohort analysis model to estimate upper and lower bounds on
fishing mortality rates and population abundance (Deriso et al. 1985). The author then applied an age-structured deterministic population model (GENMOD; Hightower and Lenarz 1989) to estimate MSY and equilibrium yields with two alternative models. The data used in that model included total catch (modeled as a single fishery), catch at age (1978-1982, surface read ages), catch at length (1978-1985), and triennial survey abundance point estimates from 1977, 1980 and 1983. The results indicated that the stock was moderately exploited, with "good recent recruitment and the absence of apparent biological stress," and the author recommended an ABC level set at the midpoint of two alternative MSY estimates, which was 3563 mt (the ABC was ultimately set at $3,600 \mathrm{mt}$ ). A precursor to the 1986 assessment was performed in 1985 (Henry 1985) using a cohort analysis, however this assessment did not result in a clear picture of stock status and did not recommend changes in the ABC levels.

## Previous STAR Panel Suggestions

The prioritized STAR Panel recommendations from the 1998 assessment included:

- Aging otoliths collected from research surveys (the triennial trawl survey)
- Investigating differences between the trawl logbook and the shelf trawl survey index
- Continuation of the midwater trawl survey for pelagic juveniles
- Continuing to monitor the age and length composition of the fishery catch
- Reporting of logbook catches of rockfish by species rather than unspecified rockfish.

For the first priority, only a very limited number of otoliths were aged in time to incorporate in this assessment, these from the 2004 NWFSC combined survey. Ageing of both historical and recent otoliths from resource surveys remains a key priority, unfortunately most of the historically collected otoliths from the triennial survey (4 survey years) were surface aged and their whereabouts are no longer known. As a result, these samples are not available to re-age using break-and-burn methods. For the second priority, the triennial survey index was developed using a somewhat different means in for this assessment, however the major data conflicts in this assessment were among the recreational CPUE survey (which tended to be in agreement with the trawl survey) and the trawl fishery catch at age data (and to a lesser extent the trawl CPUE index).

The third recommendation was to maintain the midwater trawl survey for pelagic juveniles; this survey has been maintained and in fact expanded spatially (including a second survey that is used to develop a combined coastwide index). Additional details, analysis and recommendations related to the application of juvenile indices were the subject of a Council-sponsored workshop, and recommendations in the report to the PFMC (Hastie and Ralston 2007) should be consulted for details. One recommendation was to exclude the historical (core area) index unless a strong relationship between the index and subsequent year class strength could be demonstrated. Consequently, as the core area index failed to capture the magnitude of the 1999 year class, this index was not used in the final model.

With respect to the fourth recommendation, continued data collection of age and length data from fisheries has been well maintained, and otoliths aged in a timely fashion. With respect to the reporting of logbook catches by species, it is generally agreed that the substantial impact of management measures implemented to rebuild depleted rockfish in the post-1998 era have undermined the assumptions that would allow for continuation of a trawl logbook CPUE index. Finally, while not explicitly stated in the list of prioritized research recommendations, the recognition and consideration of time-varying growth was a key uncertainty in the 1998 assessment, and remains a key research priority in this most recent review.

## Consultations with the Groundfish Advisory Subpanel (GAP) and with Fishers

Due to time and budget constraints, a pre-assessment data workshop was not held for the chilipepper and bocaccio stock assessments. Consultations with members of the GAP representatives did not suggest major concerns regarding the data available or considered for the chilipepper assessment, as there was a general sense that this stock would be shown to be above target levels. One issue raised was the question of historical discard rates, which were described as negligible by fishers prior to the implementation of highly restrictive management measures beginning in the late 1990s due to the desirability of chilipepper by processors. Consequently, discards were assumed to be zero prior to the collection of observer data in 2002.

## Model

The population was modeled using an age and size structured statistical model, Stock Synthesis II (SS2), version 2.00c, the modeling framework used for most West Coast groundfish assessments. This modeling framework was developed with the intent of allowing the complexity of the model to be consistent with the quantity and quality of the data commonly available for West Coast groundfish. The model treats a cohort as a collection of fish whose size-at-age is characterized by a mean and a variance, such that the numbers at age are distributed across defined length bins- similar to a length-age transition matrix, although with the potential to account for the effects of size-specific survivorship. The model also allows for growth, mortality, selectivity and other functions to be time varying, and time varying growth is explored in this model. A full description of the population dynamics, selectivity and catch equations, and associated likelihood functions are given in Methot (2005), while a more practical guide to using this modeling framework is provided in Methot (2006).

The base model developed here is based on equal emphasis factors (lambdas=1.0) for most likelihood components, with the exception that lambda's are set at 0.1 for length composition data where age composition data are used (trawl, hook and line, and setnet fisheries, as well as the NWFSC Combined survey). This downweighting is acknowledged to be an ad-hoc approach, to lessen the possible effects of double-use of data from the same fish. This was considered to be a reasonable interim approach based on the STAR Panel recommendations. A more appropriate approach would be to use conditional age-at-length compositions, which would also facilitate the estimation of growth (including time-varying growth) internally, however early efforts to apply conditional age-at-length information were unsuccessful and were postponed for future work. The approach used for iteratively re-weighting standard errors (for indices) and sample sizes (for catch at age, catch at length information) was based on the recommendations of
R. Methot (OST/NMFS). For standard errors, the model estimated root mean squared error (RSME) was compared to the input error, and where the model RSME was greater (lower), a scalar was added to the CVs in the data file. However, in cases where the model fits to surveys had very large input CVs (considerably larger than the model estimated RSMEs), the input CVs were reduced externally using multiplicative scalars, as the subtraction of a scalar to the input CV could result in a negative CV for some index/year combinations.

An additional problem noted during the assessment review is that the model tuning process that adjusted for inconsistencies between the "input" and "effective" sample sizes for length and age compositions treated the age- and length-compositions as independent even though length/age data for some fish were included in both length- and age-compositions.

## Prior Probabilities

Based on the recommendations from the Groundfish Harvest Policy Evaluation Workshop, a prior probability for steepness was developed by M. Dorn (AFSC, pers. comm.) for consideration in the stock assessment model. This resulted from an updated meta-analysis comparable to that developed in Dorn (2002), but excluding the contribution of chilipepper rockfish to avoid double use of stock information. The prior developed for chilipepper rockfish had a mean value of 0.573 with a standard deviation of 0.183 , very comparable to the prior probability for previously unassessed rockfish with a mean value of 0.58 and a standard deviation of 0.181 . Ultimately, steepness was fixed at this point estimate, and no other prior probabilities were used in the model, however the standard deviation of the prior probability was used to bracket uncertainty in the decision table.

## Major changes since last assessment

- Change in modeling platform to Stock Synthesis 2 v2.00c
- Catch reconstruction revised, with catch history extended back to 1892 rather than starting at an initial equilibrium in 1970 (fleet structure is unchanged).
- Length composition data extended back to 1978 (and forward to 2006), new age data include years 1978-1981 and 1998-2005. Some of these years were not used in final model.
- Relative abundance indices developed using CPFV observer data (1987-1998) and CalCOFI larval abundance data (1951-2006), although the latter were not used in the final model.
- Juvenile survey indices revised from index used in 1998 model; but excluded from the final model due to the failure of the index to capture the magnitude of the 1999 year class. A new coastwide index, based on the expanded SWFSC survey and a new NWFSC/PWCC survey, was used for the last six years of the model (2001-2006).
- Steepness fixed at 0.57 (there was no explicit spawner-recruit relationship in the 1998 model), natural mortality fixed at 0.16 for females, 0.20 for males (values in 1998 were 0.22 and 0.25 for females and males respectively).
- Selectivity curves are modeled using a double-normal selectivity curve for recreational fisheries and CPUE index.
- Time varying growth estimated internally in the model, implemented with a time-varying growth coefficient, K, using five time period blocks that were informed by major shifts in the signal for the Pacific Decadal Oscillation.


## Base Model Selection

The initial (draft) base model was developed under the assumption that a reasonable starting point would be to include all of the relevant sources of information and examine their influence on the model in the sensitivity analysis by sequentially removing time series. The model assumed a single stock, with two sexes, which had differential growth and natural mortality. Several of the time series, including the CalCOFI larval abundance index and the core juvenile rockfish survey index, were excluded from the final base model during this examination. Similarly, early exploration of alternative values for steepness, natural mortality and other parameters led to these parameters being estimated in the draft model, and fixed in the final model. Sigma-R was fixed at 1, a value consistent with the effective Sigma-R in the results, and recruitment deviations were estimated for 1965-2006. Age frequency data in this assessment were initially treated as conditional age-at-length data, an approach recommended by the developers of SS2 in order to improve the ability to fit growth curves internally and avoid problems associated with weighting of the length and age likelihood components. However, efforts to model conditional age-at-length data, and in particular efforts to tune the effective sample sizes for these data, led to a decision to use traditional catch-at-age data along with catch-at-length information.

As time-varying growth was described as a key uncertainty in the last (1998) assessment, there were numerous efforts to develop a reasonable approach to estimating time-varying growth (primarily by allowing the growth coefficients K to vary), including exploration of annual deviations, offsets staggered in three year time blocks, linking growth directly to climate indices, and allowing time-varying blocks of years that are informed by major shifts in climate indices. All improved the model fit by dozens to several hundred likelihood units, most of which was accounted for in length frequency information.

Due to both the tremendous discrepancy between design-based and GLMM-based estimates of biomass from the trawl surveys, the inconsistencies in the relative values for each survey using each estimation approach, and the observed patchiness of the data, the trawl survey indices were treated as relative abundance indices with no estimated catchability coefficients. There was general agreement that the index should provide a meaningful index of relative abundance, and consequently this index was evaluated carefully with respect to the raw data used to develop the index as well as the model fit to the index. Initial fits were quite poor, and reflected another unusual characteristic of the early versions of the model, the failure of the model to capture an
increase in relative abundance in the late 1980s as a result of the strong 1984 year class, a phenomena that was puzzling given the widespread evidence for an increase in stock abundance in most of the data.

Logistic and dome-shaped selectivity were explored for all fleets and surveys. For most fleets there was little or no improvement in fit by using dome-shaped selectivity, however the fits to the recreational fishery and CPUE data both improved significantly with dome-shaped selectivity. In the draft model and the model evaluated early in the STAR process, the setnet fishery showed strong signs of dome-shaped selectivity, within a relatively narrow size band. However, changes made during the end of the STAR week led to a selectivity curve with a double-normal parameterization that seemed to be "truncated" prior to reaching the ascending asymptote.

Developing an appropriate means of modeling selectivity to the recreational CPUE time series was widely acknowledged to be key to incorporating the index into the model, and upon exploration of various combinations of sex- and age-specific selectivity curves, a combination of size and age-based selectivity (non sex-specific) was ultimately used for this index. The ability of the model to capture the increase and subsequent stock decline associated with the strong 1984 year class, including the bimodality present in the observed length data (indicative of the dimorphic growth rates by sex of that year class), contributed to the decision to use this somewhat nontraditional approach to modeling selectivity. The model predicted lengthcompositions using length-based selectivity alone, including sex-specific length-based selectivity, failed to replicate the length composition data. However, exploration of sex-specific age selectivity curves during the STAR Panel review suggested that such an approach held promise for replacing the age- and length-based, sex-specific selectivity curve; although successful implementation would have required additional (unavailable) time.

## Base model results

For the final base model, the total number of parameters estimated in this model was 80, including $\mathrm{R}_{0}$, time-varying growth ( K offsets, 5 ), parameters for logistic selectivity curves for trawl and hook and line fisheries and the two trawl surveys (8), parameters for the double-normal selectivity curves for the setnet fishery, recreational fishery, and recreational CPUE index (18), parameters for double-normal age selectivity for the recreational CPUE index (6), and recruitment deviation values for the years 1965-2006 (42). Table 20 provides the estimates for all of these parameters, as well as the model estimated standard deviation values for most of these parameters. However, in order for the model to be able to invert the Hessian matrix, selectivity for the triennial trawl survey as well as the age selectivity for the recreational CPUE index were fixed at their estimated values and the model was re-run.

The final base model used five offsets for K that were based on intervals informed by major shifts in the Pacific Decadal Oscillation (PDO) index, with the years grouped according to a fiveblock pattern based on major changes in the PDO index (1970-1979, 1980-1988, 1989-1991, 1992-1998, and 1999-2006). The PDO has been widely described as the dominant low frequency signal in Northeast Pacific Ocean, and is essentially the leading principal component of North Pacific Ocean temperatures above $20^{\circ} \mathrm{N}$ latitude. This climate signal has been linked
to zooplankton abundance and productivity, salmon smolt survival, halibut recruitment, and other indices of marine productivity (Mantua et al. 1997; Francis et al. 2001; Clark and Hare 2002; Peterson and Schwing 2003; Logerwell et al. 2003). Consequently this approach was considered to be preferable to arbitrary multi-year bins and provided a comparable improvement in the fit to the data (on the order of 90 likelihood units at the cost of five parameters, and noting that the length frequency data were downweighted for many data sources). Other growth parameters were estimated externally.

The base model estimates of total biomass, spawning biomass, depletion, recruitment, total catch, exploitation rate, spawning biomass per recruit (SPR) are provided in Tables 21a and 21b. The model estimated an unfished spawning biomass ( $\mathrm{SSB}_{0}$ ) of 33,390 metric tons, an unfished summary biomass of 45,057 , and a 2007 spawning biomass of 23,827 , which results in a relative spawning biomass estimate of 0.71 . Figures $44-47$ show the total biomass, spawning biomass, depletion (with reference $25 \%$ and $40 \%$ of unfished biomass references), and depletion with a ten year forecast (based on 2006 status quo catches). The depletion level at its lowest point (1999) was estimated to be 8,666 tons, or $26 \%$ of $\mathrm{SSB}_{0}$. Thus, based on the base model result, the spawning biomass has nearly tripled in a relatively short ( 8 year) time period, due primarily to a very strong 1999 year class (the strongest year class estimated by the model) and greatly reduced harvest levels in recent years. Figures 48 and 49 show estimated annual recruitment values over the time period with $95 \%$ asymptotic confidence limits, and Figures 50-51 show the recruitment deviations and deviation variance checks. Figure 52 shows the estimated harvest rate by year and fishery, and Figure 53 shows the model estimated spawner recruit relationship.

The SPR was well above (current) target levels throughout most of the historical period, but was below (current) target levels between 1983 and 1997, with a low of 0.32 in 1990. The SPR has ranged between 0.72 and 0.99 since 1999, reflecting the lack of fishing mortality and fishing opportunities for chilipepper rockfish (Figures 54-55). The model estimated proxy MSY based on an $\mathrm{F}_{50 \%}$ SPR, the current (1999-2006) growth conditions, and an allocation regime consistent with the catch composition of the final year (2006) of the fishery, was estimated to be 2099 metric tons. This value was associated with an exploitation rate (catch over summary biomass) of 0.088 , and an equilibrium spawning biomass of 15,482 , which corresponds to $46 \%$ of the unfished biomass. Based on the fishing mortality rate that would cause the spawning biomass to maintain an equilibrium value of $40 \%$ of the unfished level ( $\mathrm{B}_{40 \%}$ ), the MSY proxy would be slightly greater, at 2155 metric tons, corresponding to an exploitation rate of 0.102 and an SPR of 0.45 . When the model estimated MSY internally the estimated value was very slightly greater, at 2164 metric tons (corresponding to an exploitation rate of 0.112 and an SPR of 0.43 ). Table 22 provides a more comprehensive summary of all of the relevant MSY proxy reference points.

The selectivity curves for the six fisheries are shown in Figures 56-63. Model estimated numbers at age over time, and the average age of fish in the population are shown separately for both females and males (Figures 64-67). Fits to each of the relative abundance indices (in both arithmetic and log scale) as well as scatterplots of observed versus predicted indices are shown as Figures 68-87. Figures 88 and 89 show time varying growth and Figure 90 shows model estimates of the von Bertalanffy growth coefficient (K) over time, with the mean annual winter PDO and a running three year mean of the winter PDO, which were used to inform the
designation of the time blocks. Fits to catch at length data by fleet are shown as Figures 91 through 128, including Pearson residual plots and observed versus effective sample sizes. Fits to catch at age data by fleet are shown as Figures 129 through 150, including Pearson residual plots and observed versus effective sample sizes.

Time-varying growth was included in the base model as offsets from the base K parameter for five time blocks that were structured around major changes in the Pacific Decadal Oscillation (PDO). Inclusion of time varying growth in this manner improved the overall model fit by nearly 100 likelihood units, primarily in the trawl and recreational CPUE length composition data as well as the recreational CPUE index. There were modest degradation of fits to survey length composition data and fishery age composition data. Inclusion of time-varying growth also captured a significant amount of the observed variability in the size at age of fish from commercial fisheries (Figures 151-152). However, the approach used to model time-varying growth would benefit by additional data and analyses, as discussed in greater detail in the sections that follow.

## Forecasts and decision table

The alternative states of nature used in the decision table (Table 23) were developed in conjunction with the STAR Panel, which considered a variety of potentially appropriate sources of uncertainty. As steepness was generally thought to be poorly specified for this model, the lower and upper $25 \%$ of the prior probability distribution for steepness based on the informative prior probability developed (but not used) for the assessment represented a reasonable means of bracketing uncertainty. As steepness was fixed at the mean value of the prior probability (0.57) in the base model, the alternative states of nature were consequently 0.34 (low productivity) and 0.81 (high productivity). The three catch streams used in the decision table were developed in coordination with the Groundfish Management Team (GMT) and Groundfish Advisory Subpanel (GAP) representatives to the STAR Panel, and represented "status quo" catches (based on estimates of the 2006 catch), equilibrium MSY catches (based on the SPR 0.50 harvest strategy), and ABC catches (based on the 40:10 harvest control rule). In all cases, the 2006 total catch estimates were used to apportion theoretical future catches among gear types. This was done to facilitate comparable evaluation of plausible stock trajectories under different states of nature, and in no way implies a recommended or de facto sector allocation.

The forecast scenarios included in the decision table provide a sense of the likely population trajectories under alternative fishing regimes. In all examples, it seems likely that the sharp increase in spawning biomass associated with the 1999 year class will taper off, with the stock taking a slight (under status quo fishing effort) or moderate (under equilibrium MSY or higher catches) dip in abundance in the near term. Under status quo catches, none of the states of nature suggest the possibility of the stock declining below target biomass levels ( $40 \%$ of unfished) within the next ten years. Only the low productivity scenario coupled with MSY catches or 40:10 catches (fishing down to MSY) show any risk of dipping below target levels, and even under this low productivity scenario only the very high catch stream might cause the stock to fall below the overfished limit within the next ten years. In general, the stock is above target levels and expected to remain so within the foreseeable future.

## Sensitivity Analysis

To evaluate model convergence during the model review, starting values were randomly adjusted ("jittered") between a range of starting values. During the assessment review, convergence problems were evident as indicated by irregular profile plots and other analyses. This seems to reflect an irregular likelihood surface related to conflicting signals from various data sources. Although a cause for some concern, the effects of this did not seem to be severe with respect to the model results. To evaluate the effect, twelve simulations were done with "jittered" initial values, and the resulting equilibrium recruitment estimates and likelihood estimates were plotted against each other (Figure 151). These results suggest two relatively localized minima in the likelihood surface, one very close to the minimum likelihood of the base model, the other associated with a slightly lower equilibrium recruitment value, but a considerably higher total likelihood value. The latter seemed to be associated with very poor fits to the recreational CPUE index and associated length composition data (Table 24), and may reflect the difficulty in achieving convergence with combined age and length-based selectivity for that index. However, the effects did not appear too severe for most other indices, and the model results varied only slightly even among the simulations with considerably higher likelihood values.

The sensitivity analyses reported here provided an opportunity to compare the results from the base model in terms of measures of the model fit (in likelihood units) when key parameters that were fixed at assumed values in the model were varied, as well as the changes in model results. Table 25 presents the likelihood values by data type for the two states of nature, the high steepness ( $\mathrm{h}=0.81$ ) and low steepness ( $\mathrm{h}=0.34$ ) scenarios, as well as very high ( $\mathrm{h}=0.99$ ) and very low ( $\mathrm{h}=0.21$ ) scenarios. Similarly, the Table includes likelihood estimates when female natural mortality is varied from 0.12 to 0.2 . In all examples, the male offset is $1.26 *$ Female_M, as in the base model. Likelihood profiles for steepness (h) and natural mortality (M) are presented as Figures 154 and 155, and a likelihood surface is presented as Figure 156. For all of these values, each run was "jittered" no less than ten times, and the model run with the lowest likelihood of the ten was reported for the likelihood values and profiles. The results of the sensitivity and the profiling on steepness suggests that estimates of steepness lower than the base case (0.57) are increasingly unlikely, while higher values of steepness are increasingly (but very modestly) more likely.

Overall, these results suggest that steepness is likely to be greater than approximately 0.4 , but that the model is otherwise relatively uninformative with respect to steepness. The improvement in likelihood with higher steepness values is found primarily in the trawl fishery length and age frequency data, as well as in the trawl CPUE index. By contrast the triennial survey index and the recreational CPUE index are more consistent with lower steepness values. This tension characterizes the strongest inconsistencies among the various sources of data used in this model. Consequently, the steepness value assumed for the base model is reasonable, as high steepness values for Sebastes are generally considered to be less consistent with their long-lived, slow growing life history characteristics (although chilipepper rockfish are among the faster growing species with relatively higher turnover rates), and lower levels are not consistent with the likelihood profile. Figures 157 and 158 show the resulting estimates of spawning biomass and recruitment over time with the high and low productivity scenarios, with the intuitive result that
the historical biomass is scaled upwards in the low productivity scenario, with current abundance at a slightly lower level than in the base model, while historical abundance is slightly lower in the high productivity model, and current abundance is even closer to the unfished level.

As with the previous assessment, the choice (or estimation) of M has a strong impact on the model results, and as with the previous assessment, lower natural mortality rates are associated with less severe declines in biomass over time (with a smaller overall stock size), while higher natural mortality rates are associated with greater declines in spawning biomass and higher overall stock sizes. Consequently, natural mortality is a key uncertainty in the model. Figures 159 and 160 show the estimated spawning biomass and recruitment over time with the lower (0.12) and higher (0.20) assumed values for female natural mortality; although the historical estimates of abundance change little, recent estimates are (intuitively) far more dynamic for the higher natural mortality assumption relative to the lower natural mortality assumption. The likelihood profile for $M$ suggests that the fixed (assumed) value is close to the local minima for M (Figure 153), suggesting that the assumed value is reasonable. Similarly, the likelihood surface (Figure 154) demonstrates that the gradient in likelihood is consistent across all assumed values of $h$, implying that the model is relatively more informative for natural mortality.

Another means of evaluating the sensitivity of the model is to sequentially remove datasets from the base model. Table 26 provides the likelihood values and point estimates of unfished spawning biomass and recruitment, while Figures 161-172 show the estimated trends in spawning biomass and recruitment for a suite of runs in which individual data sources are excluded or model structure otherwise altered. For most data, the consequence of removal was relatively modest, for example there were only very modest changes in estimates of $\mathrm{B}_{0}$, biomass trend and end-year depletion with removal of the trawl CPUE time series, the NWC combined survey time series, the setnet fishery length and age composition data, and the assumption of asymptotic versus dome-shaped selectivity for the setnet fishery (which in retrospect would have been a more reasonable assumption given the shape of the final selectivity curve, however the effect on the model estimates is virtually nonexistent). With the exclusion of other sources of data, there were often more noteworthy effects on model estimates of the unfished spawning biomass and the depletion trend, although none of these had a major impact on the general population trend or depletion level. For example, exclusion of the recreational CPUE index resulted in a slight scaling upwards of the unfished spawning biomass level (from ~33,400 to $\sim 35,300$ ), a flattening of the population trend during the 1990s relative to the base model (Figure 162) which suggests continued population declines in this period, and a greater population increase during the early 2000s to end at a final (2006) depletion level of $84 \%$ of the unfished level (rather than $70 \%$ in the base model). By contrast, when the trawl fishery length and age frequency data are excluded (Figure 163), the recreational CPUE data are more influential in the 1990s, such that depletion is lower in both the late 1990s (16\% rather than $26 \%$ of unfished biomass in 1998) and 2006 (53\% rather than 70\%). A similar, but less significant, result occurred when the hook and line length and age frequency data were excluded, although this result was also associated with a general scaling downward of the total spawning biomass throughout the duration of the time series.

In general, this reflects the greatest sources of tension in the model, both the trawl CPUE and length/age frequency data, as well as the hook and line length frequency and age frequency data,
were generally in conflict with the recreational CPUE data and (to a lesser extent) the triennial survey data. The latter two sources suggested greater population declines during the 1990s, while the former sources were more consistent with a relatively level biomass trend throughout the 1990s. The major effect of not including time-varying growth was a general scaling upward of the historical biomass (Figure 167), consistent with the lower productivity that this would have assumed as the growth deviations were generally all in the positive direction during the period in which they were estimated. Reconciliation of the most appropriate approach for modeling time varying growth is a key research and modeling priority for future assessments.

For the coastwide juvenile survey time series, Figure 164 shows only the estimates of SSB and recruitment from 1990 but includes a ten year forecast (assuming status quo catches), as the primary effect of this survey is to invert the recruitment estimates for 2002-2004, which are very weakly informed by the NWC combined survey length composition data, and reduce the estimates of the 2005 and 2006 year classes, which have very little data that might inform the model otherwise. As this dataset is of short duration, has not necessarily been validated, and the previous (core area, longer time series) failed to capture the magnitude of the 1999 year class (the index is moderately well correlated with year class strength estimates for other years), the inferences resulting from inclusion of the coastwide survey index should be treated with some apprehension. However, the overall effect of including this dataset is negligible with respect to estimates of reference points and biomass trend through the present period, and is relatively modest with respect to the forecast of future biomass trends. Importantly however, all of the data sources seemed to be consistent with a population increase in the early 2000s, as in none of these sensitivity runs did the end year depletion fall below $50 \%$ of the unfished population level.

A final sensitivity test evaluated the consequences of either doubling or halving the estimates of historical (pre-1978) landings of chilipepper rockfish (Figures 171-172). As described in the section on catch reconstructions, the estimated proportion of historical catches that are likely to have been chilipepper are highly uncertain for most of the pre-1978 period, including the period of foreign fisheries through the mid-1960s to the early 1970s. Doubling or halving these estimates is an ad-hoc approach to evaluating the sensitivity of the model to the exploitation history, but provides reasonable bounds on the plausible impacts. The results are consistent with the base model, with a general scaling upwards (for the doubling) and downwards (for the halving) of the historical trend, however the trend over the past 25 years and the ending depletion levels are virtually unchanged.

## Summary of Responses to STAR Panel requests

The draft assessment distributed to the STAR Panel included conditional age-at-length compositions rather than age-compositions, however problems with tuning this model resulted in a model revision that was based on both length- and age-compositions without conditional age-at-length compositions. The STAT also proposed that the core area juvenile survey index be removed from the SS2 analysis, largely as a result of the failure of that index to capture the magnitude of the extremely strong 1999 year class. In discussing the significant limitations of the CalCOFI index, both the STAT and the STAR Panel agreed that this index too was not suitable for chilipepper rockfish, primarily as the survey misses much of the spatial range of the stock. The STAR Panel accepted these initial revisions to the base model, and proposed down-
weighting those length-compositions for which there were also age-compositions. The STAR Panel also suggested fixing, rather than estimating, both steepness and natural mortality in the revised model. The mean value of steepness based on the Dorn prior probability was used for steepness, while 0.16 was used for female natural mortality (based on profiles of M in the draft model).

Among the first requests made by the Panel was the review of the length composition data for both aged and unaged fish, which uncovered some potentially imbalanced age composition subsampling and resulted in removing selected years of data from the model (although the overall influence of these data on the model was minimal). The STAR Panel and STAT also spent considerable time reviewing the data that contributed to the CPFV index, ultimately arriving at a new approach for estimating the index based on excluding the deeper depths (which had limited sampling) and considering a suite of alternative approaches for modeling selectivity, including age-based, sex-based and length-based dome-shaped selectivity curves. Considerable effort was also expended on evaluating an appropriate means of modeling time-varying growth. For both of these issues, the current approaches should be considered placeholders until more appropriate means of modeling selectivity to the recreational index and time-varying growth can be developed. The STAR Panel also provided additional guidance for future modeling efforts with respect to tuning the effective sample sizes in a model in which sampled fish contribute to both length- and age-compositions (see the STAR Panel report). This summary highlights the key issues that were raised and considered during the model review, a more detailed accounting of the requests and responses is included as Appendix C.

## Comparison with the last assessment

The major differences between the 1998 assessment and the current assessment were summarized earlier, and Figures 173 and 174 show the major differences in the results of the base models for each assessment. There is a substantial difference in the scale of the total biomass between the two models, with the 1998 model estimating a considerably larger (approximately double) spawning biomass than the current model in the early period (the 1998 model was initiated in 1970). However, the "low natural mortality rate" model run as a sensitivity test in the 1998 assessment (in which M was set to 0.16 , which is the base model M for this assessment) predicted an early 1970s total biomass of approximately $35,000 \mathrm{mt}$, much closer to $30,000 \mathrm{mt}$ total biomass estimated in the base model for this assessment (Ralston et al. 1998, Figure 38). The 1998 model also suggested a greater relative decline throughout the early 1980s, and a proportionately greater (but slightly lagged) response in the spawning biomass through the late 1980s into the 1990s. These results are also consistent with the sensitivity tests that assumed a higher natural mortality rate in this assessment (Figure 160). Estimates of recruitment in the two models were nearly identical throughout the overlapping time period (Figure 174), demonstrating consistency in both the estimation of recruitment strengths and variability. Estimates of exploitation rates and harvest projections were also similar, although estimates of both were slightly higher in the 1998 assessment.

## Retrospective analysis

A retrospective analysis was conducted by sequentially removing the most recent two years of data, such that models included data through 2004 only (Figure 175), through 2002 only (Figure 176), through 2000 only (Figure 177) and through 1998 only (Figure 178). As with other sensitivity runs, the runs were "jittered" at least 8-10 times, and the model with the lowest likelihood was presented in the comparison. The historical spawning biomass and recruitment trajectories changed very little with each analysis, which is not a terribly surprising result in a model for which steepness and natural mortality were fixed, and catches in the past 5-8 years have been minimal. Interestingly, the strength of the 1999 year class was very evident in the data by as early as 2002, and the 2000 retrospective may have mistakenly attributed an apparent abundance of small fish associated with the 1999 recruitment year (these fish were just beginning to appear in trawl catches) to a strong 1998 year class.

## Technical Deficiencies

During the STAR Panel review, the length composition data were down-weighted when associated age-composition data were available, however the approach (a lambda of 0.1 for length data where age data also exist, and 1 for the associated age data) was acknowledged to be ad-hoc and lacking a solid theoretical basis. A more appropriate approach is to use conditional age-at-length compositions, which was attempted in early runs but led to a suite of problems in model tuning.

In evaluating possible causes of these problems, the raw length composition data by fishery for years with both aged and non-aged fish was evaluated on a year-by-year basis, and where the length compositions seemed inconsistent, the emphasis on the data was effectively set to zero. For some years, there seems to be evidence that there was some geographic bias in the sampling of aged versus un-aged fish that could have been internally inconsistent, and there was at least one example of samples that had large numbers of male chilipepper that were of unreasonably large size and must have represented identification errors of some sort. However, as this evaluation was based on unexpanded length compositions, it is possible that good lengthcomposition data may have been excluded from the model. A re-evaluation of these length composition data, improved efforts to incorporate conditional age-at-length information, and approaches to model tuning that account for joint tuning of co-dependent age and length frequencies are all priorities for future assessments.

The model tuning process that adjusted for inconsistencies between the model fits to surveys (RMSE) and the input CVs took an ad hoc approach with surveys that had very large CVs for some index values. The input CVs were reduced proportionally, which was somewhat inconsistent with the normal process of adding a constant to account for process error.

The estimated growth curves had kinks that could probably be eliminated by reducing the lower bound of the smallest length bin. This would also improve estimation of the selectivity curves for the two fisheries independent trawl surveys, for which the smallest ( $<16 \mathrm{~cm}$ ) fish appear to be fully, or near fully, selected. This in turn would negate the need to fix the parameters for the triennial survey selectivity, which was necessary to invert the Hessian matrix.

The results from the convergence tests with randomly jittered starting parameter values indicated that the likelihood surface is very irregular. The final runs, as well as sensitivity runs, were "jittered" 10 to 12 times in order to better ensure convergence, however the conflicting signals of some data sources is a source of some concern. In general, biomass trajectories and other critical results do not appear to be sensitive to these differences.

Although there is a clear progression from shallow to deeper water with age and size, the application of a combined age- and length- based selectivity curve for the recreational CPFV data is somewhat non-traditional and would benefit by either more detailed investigation or an alternative selectivity configuration (an age-based, sex-specific selection curve showed considerable promise).

Although the setnet fishery was modeled with dome-shaped (double logistic) selecitivity, which indicated declining selectivity at the very largest size classes for early model configurations, the ultimate shape of the selectivity curve suggested a more monotonic increase in selectivity with largest sizes. Consequently, a logistic selectivity curve may have been more appropriate for modeling the selectivity of this fishery, although a sensitivity analysis suggest that the significance of such a change would be negligible.

## Key Uncertainties

This stock has increased substantially in recent years due to the strength of the 1999 year class, which is strongly visible in age and length composition data from both fisheries and resource surveys. Future (post-1999) year class strength is highly uncertain; although this model includes highly informative projections through 2006 based on juvenile abundance indices, the failure of the historical (core area) juvenile index to capture much of the year class variability that has been observed is troublesome.

Early catch histories are fairly uncertain. Although it is common knowledge that chilipepper have been historically important, and reasonable estimates of the total rockfish catch estimates exist, estimates of the percentage of historical catches that were chilipepper, and how that percentage may have changed over time, are based primarily on anecdotal information.

Lack of fishery-independent age data is problematic; as the four years of triennial age data were surface read, they were not used in the model (the ages up to age 8 were used in estimating the external growth curves, based on the common assumption that surface ages tend to be consistent with break and burn ages up to approximately age 10). Such data would be particularly useful in estimating time-varying growth, which seems to be an important factor for chilipepper rockfish. As the 1970-1979 estimated K is quite high (approximately 0.32), alternative approaches for estimating growth prior to the period in which most data are available should be explored. Additionally, the estimates of yield and productivity will be based in part on future assumptions regarding growth. Similarly, while there is a paucity of smaller fish in the commercial fisheries, there are indications of smaller individuals in the surveys, and including a broader range of length bins (smaller than 16 cm ) or exploring a younger minimum age ( $\mathrm{A}_{\text {min }}$ ) for the Schnute growth curve formulation could lead to improvements in how growth is estimated.

There are insufficient data to consider spatial structure in the model; although the CalCOFI time series might suggest greater relative depletion south of Point Conception, this time series has some unusual characteristics that undermine its utility as an index of abundance. As there is only very limited fisheries dependent information in this region, and only a very short (four years) time series of fishery independent information (with low sampling density), spatial features have been ignored in this model.

Discards are assumed to be negligible until 2002, when catches were scaled upwards to account for the discard rates estimated by the West Coast groundfish observer program. This assumption may be incorrect, as regulatory impacts may have resulted in an increase in discarding as management measures evolved from the mid to late 1990s to 2002 to rebuild overfished and depleted stocks. In the earlier historical period, even negligible to modest estimates of discarding in some fisheries could potentially be developed based on observed discard rates in other fisheries for earlier time periods. Average size data from the observer program have not been developed or integrated into the model, and could be evaluated in the future.

There is considerable uncertainty associated with the coastwide juvenile index as this dataset is of short duration, has not necessarily been validated, and the previous (core area survey) failed to capture the magnitude of the 1999 year class. Although the current influence of the survey is modest, and there is currently little information in the model to counter the influence of this index, it is also likely that the CVs in the coastwide index may be constraining (currently the average CV is approximately 0.037 ) as the time series lengthens and begins to overlap temporally with length and age data from fisheries and surveys. Re-evaluation of the coastwide juvenile index should be an important element of both future research and future assessments.

Since 2003, the Rockfish Conservation Areas (RCAs) have been the primary management tool implemented to protect rebuilding species that co-occur with chilipepper, such as bocaccio, widow, and canary rockfish. As a result of these management measures and reductions in trip limits, catches of chilipepper rockfish have declined significantly, limiting the amount of fisherydependent information (age and length frequency information) available to the assessment model. However, such measures have also likely resulted in a bias in those age and length frequency information that do exist, as such data are derived from fish that were caught either shoreward or seaward of the RCAs, while the areas of greatest chilipepper abundance are within the RCAs. As a result, and further complicated by the clear ontogenetic shift to deeper water with size (and presumably age), these age and length frequency information are not likely to be reflective of the true age and length structure of the population (e.g., Punt and Methot 2004; Field et al. 2006). Such considerations could potentially be addressed by a more rigorous evaluation of the sources of the data, and possibly by including alternative selectivity curves for the post-RCA period, however such approaches were not evaluated in detail in this assessment and should be considered in future assessments.

## Regional Management Concerns

There are insufficient data to consider spatial structure in the model, consequently the resource is modeled as a single stock. Although the stock extends north of Cape Blanco, Oregon, the abundance and catches are minimal and have no significance in the model. Catches and biomass
between Cape Mendocino and Cape Blanco are modest, but noteable and historically accounted for in landings and surveys. By contrast, catches and biomass trends south of Point Conception are poorly quantified and highly uncertain, but anecdotal accounts suggest that chilipepper were historically a relatively important stock in this region. Although the CalCOFI time series (which was not used in the final model) is suggestive of greater relative depletion in this region, this time series has some unusual characteristics that undermine its utility as an index of abundance. As there is only very limited fisheries dependent information in this region, and only a very short (four years) time series of fishery independent information (with low sampling density), there is insufficient information to assess potential regional concerns in this area. Increased sampling of both fisheries data and by resource surveys are critical to any attempts to develop a greater understanding of potential spatial differences in stock status and trends in this region. However, as the Southern California Bight appears to be a region of sharply declining abundance, and abundance appears to drop even more sharply towards the U.S./Mexico border, transboundary issues are minimal for this stock.

## Research and Data Needs

Additional investigations into the catch history should be made, including greater evaluation of detailed historical landings data from fish tickets (ongoing) which should inform catch history reconstructions. As has been recommended previously by both STAT teams and STAR panels, the reconstruction of historical rockfish landings should be done comprehensively across all rockfish species to ensure efficiency and consistency (priority medium, medium to long term).

Information on maturity and fecundity is available, but limited. Additional information should be compiled and carefully evaluated for accuracy, potential changes over time, and potential maternal effects (priority medium, long term).

There is a paucity of length at age information for smaller fish, particularly those collected in fishery independent surveys. Otoliths that are available from past years of the triennial survey, and those available from the combined survey, should be aged to provide better data on the early stages of growth and possible time-variations in growth. Additionally, aging error is poorly estimated, as only a modest number of otoliths were read by two readers, and most of these were relatively young fish. Additional double-reads of break and burn otoliths should be conducted to better estimate ageing error (priority high, short term).

Greater exploration of methods for modeling time-varying growth as influenced by environmental factors should be a key research area for future assessments. Such exploration will benefit substantially from both an increased availability of data from research catches (both historical and recent) as well as a renewed attempt to model age and length data using conditional length-at-age approaches (priority high, short to medium term).

The consequences of the Rockfish Conservation Areas to vulnerability, selectivity patterns and other stock attributes could be significant, and would benefit from greater analysis as well as more generic simulation studies that might inform assessment authors of the consequences of spatially explicit management measures to the basic assumptions of stock assessment models (priority medium, medium to long term).

Additional fisheries dependent and fisheries independent data for the region south of Point Conception (including additional evaluation of historical landings in this region) is essential in evaluating whether the relative stock status may be different in this region relative to the coastwide trend, as might be suggested by a superficial evaluation of the CalCOFI data. Further evaluation of the CalCOFI data, to determine the extent to which these data may or may not inform relative trends at a more spatially explicit level, should also be a research priority (priority medium, medium to long term).

## Acknowledgements

I am grateful to the STAR Panel, chaired by David Sampson and including Patrick Cordue, Norman Hall, Kevin Piner, Gerry Richter, and John DeVore. The Panel provided important contributions, feedback and suggestions that improved the final model considerably. Additionally, I am grateful to a large number of people who provided data and feedback on earlier versions of this model. Steve Ralston provided such feedback in addition to providing the insights from the last (1998) model and developing the juvenile abundance indices, Don Pearson assisted in the CalCOM queries and questions, as well as aging many of the fish used in this model, Alec MacCall provided assistance with the recreational CPUE index as well as feedback on early drafts of the model, E.J. Dick provided a great deal of support in developing the recreational CPUE and CalCOFI indices, Xi He provided both feedback and computer time during the model development, and Meisha Key provided advice and feedback during model development, particularly with respect to RecFIN and CPFV observer data. Ian Stewart and Richard Methot provided valuable support in using SS2 and in graphing the output effectively and rapidly, Beth Horness and Tom Helser provided data, indices and answers to many questions and requests for data from the NWC combined survey, Jim Hastie provided bycatch estimates, and Stacey Miller provided logistical support and coordination of the review Panel. Mark Wilkins provided triennial survey biomass estimates and length composition expansions as well as answers to many questions regarding these data, John DeVore provided the detailed history of regulations that affected chilipepper rockfish management, and Richard Charter and Susie Jacobson provided CalCOFI data and patiently answered questions as well. Armies of port samplers (and of course fishermen) provided the data upon which the entire model is based.

## Sources

Adams, P.B. and D.F. Howard. 1996. Natural mortality of blue rockfish during their first year in nearshore benthic habitats. Fishery Bulletin 94: 156-162.

Alonzo, S.H., M. Key, T. Ish, and A.D. MacCall. 2004. Status of the California sheephead (Semicossyphus pulcher) stock. California Department of Fish and Game.

Alverson, D.L., A.T. Pruter and L.L. Ronholt. 1964. A study of demersal fishes and fisheries of the Northeastern Pacific Ocean. Institute of Fisheries, University of British Columbia:
Vancouver, CA.
Best, E.A. 1963. The California marine fish catch for 1961. Fish Bull. 121. 56 p.

Beverton, R.J.H. 1992. Patterns of reproductive strategy parameters in some marine teleost fishes. Journal of Fish Biology B41: 137-160.

Clark, W.G. and S.R. Hare. 2002. Effects of climate and stock size on recruitment and growth of Pacific halibut. North American Journal of Fisheries Management 22: 852-862.

Conser, R.J., K.T. Hill, P.R. Crone, N.C.H. Lo, and D. Bergen. 2002. Stock assessment of Pacific sardine with management recommendations for 2003. In Appendix to the status of the Pacific coast coastal pelagic species fishery through 2002 and recommended acceptable biological catches for 2003: Pacific Fishery Management Council: Portland, OR.

Croaker, R.S. 1939. Three years of fisheries statistics on marine sport fishing in California. Transactions of the American Fisheries Society 69: 117-118.

Davis, J.C. 1949. Salt water fishing on the Pacific coast. A.S. Barnes and Co. New York.
Deriso, R.B., T.J. Quinn II and P.R. Neal. 1985. Catch-age analysis with auxiliary information. Canadian Journal of Fisheries and Aquatic Sciences 42: 815-824.

Dick, E.J. 2004. Beyond 'lognormal versus gamma': discrimination among error distributions for generalized linear models. Fisheries Research 70: 351-366.

Dorn, M.W. 2002. Advice on West coast rockfish harvest rates from Bayesian meta-analysis of stock-recruit relationships. North American Journal of Fisheries Management 22: 280-300.

Douglas, D.A. 1998. Species composition of rockfish in catches by Oregon trawlers, 1963-1993. Oregon Department of Fish and Wildlife Marine Program Data Series Report.

Field, J.C. and S. Ralston. 2005. Spatial variability in California Current rockfish recruitment events. Canadian Journal of Fisheries and Aquatic Sciences 62: 2199-2210.

Field, J.C., A.E. Punt, R.D. Methot, and C.J. Thomson. 2006. Does MPA mean 'major problem for assessments'? Considering the consequences of place-based management systems. Fish and Fisheries 7: 284-302

Field, J.C. E.J. Dick, M. Key, M. Lowry, Y. Lucero, A. MacCall, D. Pearson, S. Ralston, W. Sydeman, and J. Thayer. 2007. Population dynamics of an unexploited rockfish, Sebastes jordani, in the California Current. Proceedings of the Lowell-Wakefield Symposium on the Biology, Assessment and Management of North Pacific Rockfish.

Francis, R.C., S.R. Hare, A.B. Hollowed, and W.S. Wooster. 1998. Effects of interdecadal climate variability on the oceanic ecosystems of the Northeast Pacific. Fisheries Oceanography 7:1-21.

Gunderson, D. R., J. Robinson, and T. Jow. 1974. Importance and species composition of continental shelf rockfish landed by United States trawlers. International North Pacific Fisheries Commission Report.

Hastie, J. and M. Bellman. 2006. Estimated 2005 discard and total catch of selected groundfish species. Agenda Item E.2.b., attachment 1, Pacific Fisheries Management Council Briefing Book, March 2007.

Hastie, J. and S. Ralston. 2007. Report from Pre-Recruit Survey Workshop September 13-15, 2006, http://www.pcouncil.org/bb/2007/0407/E1b_NWFSC3_sup.pdf.

Heimann, R.F.G. 1963. Trawling in the Monterey Bay Area, with special reference to catch composition. California Department of Fish and Game 49: 152-173.

Heimann, R.F.G. and D.J. Miller. 1960. The Morro Bay otter trawl and party boat fisheries August, 1957 to September, 1958. California Department of Fish and Game 46: 35-67.

Heimann, R.F. and J.G. Carlisle, Jr. 1970. The California marine fish catch for 1968 and historical review 1916-1968. California Department of Fish and Game Fishery Bulletin 149.

Helser, T.E., I.J. Stewart, C. Whitmire, and B. Horness. 2005 draft. Model-Based Estimates of Abundance for 11 species from the NMFS slope surveys. Available from T. Helser, NMFS, NWFSC, 2725 Montlake Blvd.E., Seattle, WA 98112. 142 pp.

Helser, T., I. J. Stewart, G. Fleischer, and S Martell. 2006. Stock Assessment of Pacific Hake (Whiting) in U.S. and Canadian Waters in 2006. In Volume 7: Status of the Pacific Coast Groundfish Fishery Through 2005, Stock Assessment and Fishery Evaluation Portland, OR: Pacific Fishery Management Council.

Henry, F. D., Jr. 1985. A progress report on the status of chilipepper (Sebastes goodei) off California. In: Appendix to the Status of the Pacific Coast Groundfish Fishery through 1985 and Recommended Acceptable Biological Catches for 1986, Stock Assessment and Fishery Evaluation, Pacific Fishery Management Council, Portland, OR.

Henry, F. D., Jr. 1986. Status of the coastwide chilipepper (Sebastes goodei) fishery. In: Appendix to the Status of the Pacific Coast Groundfish Fishery through 1986 and Recommended Acceptable Biological Catches for 1987, Stock Assessment and Fishery Evaluation, Pacific Fishery Management Council, Portland, OR.

Hightower, J.E. and W.H. Lenarz. 1989. Using GENMOD to develop harvesting policies for multiaged fish stocks. American Fisheries Society Symposium 6: 209-210.

Hill, K.H. and N. Schneider.1999. Historical logbook databases from California’s commercial passenger fishing vessel (partyboat) fishery, 1936-1997. SIO Reference Series No. 99-19.

Hoenig, J.M. 1983. Empirical use of longevity data to estimate mortality rates. Fishery Bulletin 81: 898-903.

Hyde, J.R. and R.D. Vetter. 2007. The origin, evolution, and diversification of rockfishes of the genus Sebastes (Cuvier). Molecular Phylogenetics and Evolution. In press.

Jordan, D.S. and B.W. Evermann. 1898. The fishes of North and Middle America. Bulletin of the U.S. National Museum 47, pt. 2.

Kronman, M. 1999. Santa Barbara’s hook-and-line fisheries. Prepared for the Santa Barbara Maritime Museum.

Logerwell E., P. Lawson, N. Mantua, R.C. Francis, and V. Agostini. 2003. Tracking environmental processes in the coastal zone for understanding and predicting Oregon coho (Oncorhynchus kisutch) marine survival. Fisheries Oceanography 12:554-568.

Love, M.S., M. Yolkavich and L. Thorsteinson. 2002. The rockfishes of the Northeast Pacific. University of California Press: Berkeley.

MacCall, A.D. 2003. Status of Bocaccio off California in 2003. In Appendix to the status of the Pacific coast groundfish fishery through 2003: Stock assessment and fishery evaluation. Pacific Fishery Management Council: Portland, Oregon.

MacGregor, J.S. 1986. Relative abundance of four species of Sebastes off California and Baja California. CalCOFI Reports 27: 121-135.

Mantua, N.J., S.R. Hare, Y. Zhang, J.M. Wallace, and R.C. Francis, R.C. 1997. A Pacific interdecadal climate oscillation with impacts on salmon production. Bulletin of the American Meteorological Society 78: 1069-1079.

Maunder, M.N. and G.M. Watters. 2003. A general framework for integrating environmental time series into stock assessment models: model description, simulation testing, and example. Fishery Bulletin 101: 89-99.

Maunder, M.N. and A.E. Punt. 2004. Standardizing catch and effort data: a review of recent approaches. Fisheries Research 70: 141-159.

Miller, D.J. and D. Gotshall. 1965. Ocean sportfish catch and effort from Oregon to Point Arguello, California. California Department of Fish and Game Fish Bulletin 130. 135 p.

Nitsos, R.J. 1965. Species composition of rockfish (Family Scorpaenidae) landed by California Otter Trawl Vessels, 1962-1963. $16^{\text {th }}$ and $17^{\text {th }}$ Annual Reports of the Pacific Marine Fisheries Commission for the years 1963 and 1964.

Orcutt, H. G. 1969. Bottomfish resources of the California Current System. CalCOFI Reports 13: 53-59.

Pearson, D. E., and B. Erwin. 1997. Documentation of California's commercial market sampling data entry and expansion programs. U.S. Dep. Commer., NOAA Tech. Memo. NMFS-SWFSC-240, 62 p.

Peterson, W.T. and F.B. Schwing. 2003. A new climate regime in the northeast Pacific ecosystems. Geophysical Research Letters 30: 17528-17533.

Phillips, J.B. 1939. The rockfish of the Monterey wholesale fish markets. California Fish and Game 25: 214-225.

Philips, J.B. 1964. Life history studies on ten species of rockfish (genus Sebastodes). California Department of Fish and Game Fish Bulletin 126. 63 p.

Pikitch, E.K., D.L. Erickson and J.R. Wallace. 1988. An evaluation of the effectiveness of trip limits as a management tool. NWAFC Processed Report 88-27.

Punt, A.E. and Methot, R.D. 2004. Effects of marine protected areas on the assessment of marine fisheries. In Shipley, J.B. (editor), Aquatic Protected Areas as Fisheries Management Tools. American Fisheries Society Symposium 42, pp. 133-154.

Ralston, S. 1999. Trends in standardized catch rate of some rockfishes (Sebastes spp.) from the California trawl logbook database. SWFSC Admin. Rep. SC-99-01.40p.

Ralston, S. and Howard, D.F. 1995. On the development of year-class strength and cohort variability in two northern California rockfishes. Fishery Bulletin 93: 710-720.

Ralston, S., D. Pearson and J. Reynolds. 1998. Status of the chilipepper rockfish in 1998. In Appendix to Status of the Pacific Coast Groundfish Fishery through 1998 and recommended acceptable biological catches for 1999. Stock Assessment and Fishery Evaluation. Pacific Fishery Management Council, Portland OR.

Roedel, P.M. 1948. California Department of Fish and Game Fish Bulletin No. 68. Common Marine Fishes of California.

Rogers, J. B., and J. R. Bence. 1992. Review of fishery and auxiliary data for chilipepper rockfish in the Eureka/Conception/Monterey INPFC areas: a qualitative assessment of the status of the stock in 1992. In: Appendices to the status of the Pacific coast groundfish fishery through 1992 and recommended acceptable biological catches for 1993, Stock Assessment and Fishery Evaluation, Pacific Fishery Management Council, 2000 SW First Avenue, Suite 420, Portland, OR, 97201.

Rogers, J. B., and J. R. Bence. 1993. Status of the chilipepper rockfish stock in 1993. In: Appendices to the status of the Pacific coast groundfish fishery through 1993 and recommended acceptable biological catches for 1994, Stock Assessment and Fishery Evaluation, Pacific Fishery Management Council, 2000 SW First Avenue, Suite 420, Portland, OR, 97201.

Rogers, J.B. 2003. Species allocation of Sebastes and Sebastolobus sp. caught by foreign countries from 1965 through 1976 off Washington, Oregon, and California, USA. NOAA Tech. Memo. NMFS-NWFSC-57.

Sakuma, K. M., S. Ralston, and V. G. Wespestad. 2006. Interannual and spatial variation in young-of-the-year rockfish, Sebastes spp.: expanding and coordinating the sampling frame. CalCOFI Reports 47: 127-139.

Schnute, J.T. 1981. A versatile growth model with statistically stable parameters. Canadian Journal of Fisheries and Aquatic Sciences. 38:1128-1140.

Scofield, W.L. 1948. Trawling gear in California. California Department of Fish and Game Fish Bulletin 72.

Sette, O.E. and R.H. Fiedler. 1928. Fishery Industries of the United States, 1927. In Report of the United States Commissioner of Fisheries for the Fiscal Year 1928. U.S. Department of Commerce.

Walford, L.A. 1930. California Department of Fish and Game Fish Bulletin No. 28. Handbook of Common Commercial and Game Fishes of California.

Westrheim, S.J. 1965. Northern range extensions for four species of rockfish (Sebastodes goodei, S. helvomaculatus, S. rubrivinctus, and S. zacentrus) in the north Pacific Ocean. Journal of the Fisheries Research Board of Canada 22: 231-235.

Young, P.H. 1969. The California Partyboat Fishery 1947-1967. California Department of Fish and Game Fish Bulletin 145. 91 p.

Table 1: Management performance in obtaining the ABC and OY for chilipepper rockfish (catch includes all catches in all areas, commercial and recreational, as well as estimated discards from 2002-2006; discards prior to 2002 are assumed to be negligible, although some regulatory discarding was likely).

| Year | ABC | OY | Catch | \%ABC | \%OY |
| ---: | ---: | ---: | ---: | ---: | ---: |
| 1982 | - |  | 2492 |  |  |
| 1983 | 2300 |  | 2465 | 107 |  |
| 1984 | 2300 |  | 2923 | 127 |  |
| 1985 | 2300 |  | 3182 | 138 |  |
| 1986 | 2300 |  | 3147 | 137 |  |
| 1987 | 2300 |  | 2059 | 90 |  |
| 1988 | 3600 |  | 2691 | 75 |  |
| 1989 | 3600 |  | 3395 | 94 |  |
| 1990 | 3600 |  | 3110 | 86 |  |
| 1991 | 3600 |  | 3311 | 92 |  |
| 1992 | 3600 |  | 2753 | 76 |  |
| 1993 | 3600 |  | 2393 | 66 |  |
| 1994 | 4000 |  | 1877 | 47 |  |
| 1995 | 4000 |  | 2021 | 51 |  |
| 1996 | 4000 |  | 1870 | 47 |  |
| 1997 | 4000 |  | 2110 | 53 |  |
| 1998 | 3400 |  | 1430 | 42 |  |
| 1999 | 3724 | 3724 | 977 | 26 | 26 |
| 2000 | 3681 | 2000 | 499 | 14 | 25 |
| 2001 | 2700 | 2000 | 517 | 19 | 26 |
| 2002 | 2700 | 2000 | 329 | 12 | 16 |
| 2003 | 2700 | 2000 | 21 | 1 | 1 |
| 2004 | 2700 | 2000 | 236 | 9 | 12 |
| 2005 | 2700 | 2000 | 192 | 7 | 10 |
| 2006 | 2700 | 2000 | 127 | 5 | 6 |
| 2007 | 2700 | 2000 | - |  |  |

Table 2: Estimated chilipepper rockfish landings by gear type for the early period (1892-1927), based on reported estimates of total rockfish landings by Sette and Fiedler (1928, bold under "all rockfish"), interpolated estimates for intervening years, the estimated ratio of chilipepper to all rockfish in 1928 based on the regional landings data, and the assumption that $95 \%$ of rockfish landings were hook and line until 1943.

|  | trawl | hookline | total |
| :---: | :---: | :---: | :---: |
| 1892 | 11 | 206 | 217 |
| 1893 | 10 | 195 | 205 |
| 1894 | 10 | 183 | 193 |
| 1895 | 9 | 171 | 180 |
| 1896 | 9 | 162 | 170 |
| 1897 | 8 | 152 | 160 |
| 1898 | 8 | 143 | 150 |
| 1899 | 7 | 133 | 140 |
| 1900 | 8 | 147 | 155 |
| 1901 | 8 | 161 | 170 |
| 1902 | 9 | 176 | 185 |
| 1903 | 10 | 190 | 200 |
| 1904 | 11 | 204 | 215 |
| 1905 | 11 | 218 | 229 |
| 1906 | 12 | 232 | 244 |
| 1907 | 13 | 246 | 259 |
| 1908 | 14 | 260 | 274 |
| 1909 | 15 | 292 | 308 |
| 1910 | 17 | 325 | 342 |
| 1911 | 19 | 358 | 376 |
| 1912 | 21 | 390 | 411 |
| 1913 | 22 | 423 | 445 |
| 1914 | 24 | 455 | 479 |
| 1915 | 26 | 488 | 513 |
| 1916 | 33 | 633 | 666 |
| 1917 | 41 | 778 | 819 |
| 1918 | 49 | 924 | 972 |
| 1919 | 32 | 605 | 637 |
| 1920 | 33 | 631 | 665 |
| 1921 | 28 | 534 | 562 |
| 1922 | 25 | 483 | 509 |
| 1923 | 30 | 571 | 601 |
| 1924 | 28 | 532 | 560 |
| 1925 | 32 | 615 | 648 |
| 1926 | 44 | 845 | 890 |
| 1927 | 38 | 716 | 754 |

Table 3: Total California rockfish catches by region (based on CDF\&G Fisheries Bulletin reports) and as estimated by gear type.

| Year | Trawl | \% trawl | \% hookline | Eureka | $\begin{array}{r} \text { San } \\ \text { Francisco } \end{array}$ | Monterey | Santa Barbara | Los <br> AngelesSan | Diego | CA Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1928 |  | 0.050 | 0.950 | 49 | 453 | 1037 | 47 | 770 | 555 | 2911 |
| 1929 |  | 0.050 | 0.950 | 117 | 487 | 745 | 45 | 687 | 642 | 2723 |
| 1930 |  | 0.050 | 0.950 | 114 | 466 | 1282 | 21 | 906 | 478 | 3268 |
| 1931 |  | 0.050 | 0.950 | 48 | 473 | 1162 | 31 | 1183 | 400 | 3298 |
| 1932 |  | 0.050 | 0.950 | 40 | 451 | 930 | 35 | 798 | 299 | 2552 |
| 1933 |  | 0.050 | 0.950 | 14 | 516 | 734 | 47 | 588 | 253 | 2152 |
| 1934 |  | 0.050 | 0.950 | 58 | 414 | 762 | 128 | 511 | 130 | 2001 |
| 1935 |  | 0.050 | 0.950 | 73 | 402 | 976 | 178 | 374 | 78 | 2080 |
| 1936 |  | 0.050 | 0.950 | 85 | 391 | 1189 | 182 | 123 | 70 | 2039 |
| 1937 |  | 0.050 | 0.950 | 61 | 470 | 955 | 166 | 157 | 65 | 1875 |
| 1938 |  | 0.050 | 0.950 | 248 | 254 | 839 | 73 | 126 | 34 | 1573 |
| 1939 |  | 0.050 | 0.950 | 342 | 176 | 603 | 91 | 141 | 92 | 1445 |
| 1940 |  | 0.050 | 0.950 | 264 | 206 | 753 | 136 | 153 | 67 | 1579 |
| 1941 |  | 0.250 | 0.750 | 206 | 205 | 662 | 132 | 203 | 42 | 1451 |
| 1942 |  | 0.500 | 0.500 | 123 | 32 | 298 | 38 | 74 | 10 | 576 |
| 1943 |  | 0.750 | 0.250 | 624 | 92 | 311 | 39 | 89 | 5 | 1160 |
| 1944 |  | 0.900 | 0.100 | 2506 | 31 | 332 | 22 | 10 | 5 | 2907 |
| 1945 |  | 0.900 | 0.100 | 5315 | 84 | 534 | 45 | 27 | 5 | 6009 |
| 1946 |  | 0.900 | 0.100 | 4007 | 100 | 508 | 49 | 80 | 9 | 4752 |
| 1947 |  | 0.900 | 0.100 | 2497 | 96 | 690 | 27 | 132 | 9 | 3450 |
| 1948 |  | 0.900 | 0.100 | 1595 | 123 | 748 | 36 | 200 | 24 | 2726 |
| 1949 |  | 0.900 | 0.100 | 1275 | 236 | 611 | 62 | 259 | 37 | 2481 |
| 1950 |  | 0.900 | 0.100 | 1556 | 449 | 1107 | 86 | 294 | 34 | 3525 |
| 1951 |  | 0.900 | 0.100 | 2052 | 1000 | 1441 | 122 | 329 | 15 | 4958 |
| 1952 |  | 0.900 | 0.100 | 1090 | 1625 | 1677 | 108 | 219 | 9 | 4728 |
| 1953 |  | 0.900 | 0.100 | 1336 | 1892 | 1954 | 89 | 179 | 15 | 5466 |
| 1954 | 4899 | 0.892 | 0.108 | 1263 | 1354 | 2349 | 263 | 247 | 14 | 5491 |
| 1955 | 5035 | 0.899 | 0.101 | 1225 | 709 | 1887 | 1533 | 199 | 48 | 5601 |
| 1956 | 5897 | 0.887 | 0.113 | 1305 | 1335 | 2548 | 1169 | 258 | 35 | 6650 |
| 1957 | 6396 | 0.886 | 0.114 | 1676 | 1279 | 2482 | 1523 | 228 | 32 | 7220 |
| 1958 | 6486 | 0.814 | 0.186 | 1610 | 1903 | 2657 | 1426 | 229 | 141 | 7967 |
| 1959 | 5534 | 0.818 | 0.182 | 1366 | 2233 | 2132 | 671 | 265 | 95 | 6761 |
| 1960 | 5352 | 0.889 | 0.111 | 1300 | 1493 | 1617 | 1281 | 239 | 90 | 6019 |
| 1961 | 4037 | 0.862 | 0.138 | 885 | 1008 | 1465 | 1053 | 175 | 99 | 4684 |
| 1962 | 3538 | 0.849 | 0.151 | 808 | 903 | 1295 | 917 | 172 | 70 | 4166 |
| 1963 | 4445 | 0.883 | 0.117 | 1332 | 1070 | 1119 | 1181 | 221 | 112 | 5034 |
| 1964 | 3078 | 0.864 | 0.136 | 768 | 794 | 987 | 719 | 208 | 87 | 3562 |
| 1965 | 3481 | 0.838 | 0.162 | 1082 | 715 | 1188 | 786 | 249 | 133 | 4153 |
| 1966 | 3856 | 0.861 | 0.139 | 822 | 732 | 1536 | 1027 | 226 | 136 | 4480 |
| 1967 |  | 0.860 | 0.140 | 1075 | 389 | 1156 | 1313 | 251 | 167 | 4351 |
| 1968 |  | 0.860 | 0.140 | 1272 | 265 | 1087 | 1188 | 243 | 126 | 4180 |
| 1969 | 3434 | 0.860 | 0.140 | 1340 | 276 | 932 | 1133 | 227 | 86 | 3994 |
| 1970 | 4109 | 0.866 | 0.134 | 1694 | 350 | 1305 | 1115 | 172 | 108 | 4744 |
| 1971 | 4018 | 0.809 | 0.191 | 2098 | 565 | 1088 | 869 | 197 | 150 | 4968 |
| 1972 | 5969 | 0.829 | 0.171 | 2734 | 736 | 1669 | 1493 | 301 | 267 | 7200 |
| 1973 | 7958 | 0.823 | 0.177 | 2371 | 1391 | 3528 | 1759 | 277 | 344 | 9671 |
| 1974 |  | 0.832 | 0.168 | 3277 | 984 | 2723 | 1809 | 224 | 584 | 9602 |
| 1975 |  | 0.841 | 0.159 | 3679 | 1014 | 2732 | 2168 | 369 | 445 | 10407 |
| 1976 |  | 0.851 | 0.149 | 4410 | 1105 | 2193 | 2652 | 328 | 460 | 11147 |
| 1977 |  | 0.860 | 0.140 | 3183 | 826 | 2292 | 2514 | 214 | 407 | 9435 |

Table 4: Fraction of rockfish landings by region assumed to be chilipepper, based on analysis in text (where bold early years represent fractions supported by literature estimates, and 1978-1979 fractions are based on CalCOM estimates).

|  | Eureka | San Francisco | Monterey | Santa Barbara | Los <br> AngelesSan | Diego |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1928 | 0.057 | 0.308 | 0.308 | 0.308 | 0.200 | 0.200 |
| 1929 | 0.057 | 0.308 | 0.308 | 0.308 | 0.200 | 0.200 |
| 1930 | 0.057 | 0.308 | 0.308 | 0.308 | 0.200 | 0.200 |
| 1931 | 0.057 | 0.308 | 0.308 | 0.308 | 0.200 | 0.200 |
| 1932 | 0.057 | 0.308 | 0.308 | 0.308 | 0.200 | 0.200 |
| 1933 | 0.057 | 0.308 | 0.308 | 0.308 | 0.200 | 0.200 |
| 1934 | 0.057 | 0.308 | 0.308 | 0.308 | 0.200 | 0.200 |
| 1935 | 0.057 | 0.308 | 0.308 | 0.308 | 0.200 | 0.200 |
| 1936 | 0.057 | 0.308 | 0.308 | 0.308 | 0.200 | 0.200 |
| 1937 | 0.057 | 0.308 | 0.308 | 0.308 | 0.200 | 0.200 |
| 1938 | 0.057 | 0.308 | 0.308 | 0.308 | 0.200 | 0.200 |
| 1939 | 0.057 | 0.308 | 0.308 | 0.308 | 0.200 | 0.200 |
| 1940 | 0.057 | 0.308 | 0.308 | 0.308 | 0.200 | 0.200 |
| 1941 | 0.057 | 0.308 | 0.308 | 0.308 | 0.200 | 0.200 |
| 1942 | 0.057 | 0.331 | 0.213 | 0.341 | 0.200 | 0.200 |
| 1943 | 0.057 | 0.331 | 0.213 | 0.341 | 0.200 | 0.200 |
| 1944 | 0.057 | 0.331 | 0.213 | 0.341 | 0.200 | 0.200 |
| 1945 | 0.057 | 0.331 | 0.213 | 0.341 | 0.200 | 0.200 |
| 1946 | 0.057 | 0.331 | 0.213 | 0.341 | 0.200 | 0.200 |
| 1947 | 0.057 | 0.331 | 0.213 | 0.341 | 0.200 | 0.200 |
| 1948 | 0.057 | 0.331 | 0.213 | 0.341 | 0.200 | 0.200 |
| 1949 | 0.057 | 0.331 | 0.213 | 0.341 | 0.200 | 0.200 |
| 1950 | 0.057 | 0.331 | 0.213 | 0.341 | 0.200 | 0.200 |
| 1951 | 0.057 | 0.331 | 0.213 | 0.341 | 0.200 | 0.200 |
| 1952 | 0.057 | 0.331 | 0.213 | 0.341 | 0.200 | 0.200 |
| 1953 | 0.057 | 0.331 | 0.213 | 0.341 | 0.200 | 0.200 |
| 1954 | 0.057 | 0.331 | 0.213 | 0.341 | 0.200 | 0.200 |
| 1955 | 0.057 | 0.331 | 0.213 | 0.341 | 0.200 | 0.200 |
| 1956 | 0.057 | 0.331 | 0.213 | 0.341 | 0.200 | 0.200 |
| 1957 | 0.057 | 0.331 | 0.213 | 0.341 | 0.200 | 0.200 |
| 1958 | 0.057 | 0.331 | 0.213 | 0.341 | 0.200 | 0.200 |
| 1959 | 0.057 | 0.331 | 0.213 | 0.341 | 0.200 | 0.200 |
| 1960 | 0.057 | 0.331 | 0.213 | 0.341 | 0.200 | 0.200 |
| 1961 | 0.057 | 0.331 | 0.213 | 0.341 | 0.200 | 0.200 |
| 1962 | 0.059 | 0.365 | 0.230 | 0.389 | 0.200 | 0.200 |
| 1963 | 0.054 | 0.297 | 0.196 | 0.293 | 0.200 | 0.200 |
| 1964 | 0.057 | 0.331 | 0.213 | 0.341 | 0.206 | 0.206 |
| 1965 | 0.066 | 0.327 | 0.224 | 0.332 | 0.213 | 0.213 |
| 1966 | 0.076 | 0.323 | 0.234 | 0.322 | 0.219 | 0.219 |
| 1967 | 0.086 | 0.319 | 0.245 | 0.312 | 0.225 | 0.225 |
| 1968 | 0.095 | 0.315 | 0.256 | 0.302 | 0.232 | 0.232 |
| 1969 | 0.105 | 0.311 | 0.266 | 0.293 | 0.238 | 0.238 |
| 1970 | 0.114 | 0.307 | 0.277 | 0.283 | 0.245 | 0.245 |
| 1971 | 0.124 | 0.303 | 0.288 | 0.273 | 0.251 | 0.251 |
| 1972 | 0.134 | 0.299 | 0.299 | 0.264 | 0.257 | 0.257 |
| 1973 | 0.134 | 0.299 | 0.299 | 0.264 | 0.264 | 0.264 |
| 1974 | 0.143 | 0.283 | 0.308 | 0.215 | 0.227 | 0.215 |
| 1975 | 0.152 | 0.268 | 0.317 | 0.166 | 0.190 | 0.167 |
| 1976 | 0.162 | 0.252 | 0.326 | 0.117 | 0.154 | 0.119 |
| 1977 | 0.171 | 0.237 | 0.335 | 0.069 | 0.117 | 0.071 |
| 1978 | 0.181 | 0.222 | 0.344 | 0.020 | 0.081 | 0.022 |
| 1979 | 0.209 | 0.194 | 0.337 | 0.019 | 0.080 | 0.021 |

Table 5: Estimated landings of chilipepper rockfish by California region, 1928-1979, including Oregon and Foreign Fishery landings, and by gear type.

| Year | Eureka | San <br> Francisco | Monterey | Santa Barbara | $\begin{array}{r} \text { Los } \\ \text { Angeles } \end{array}$ | San Diego | Oregon | Foreign Fisheries | Trawl | Hook-line |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1928 | 3 | 140 | 320 | 14 | 154 | 111 |  |  | 37 | 701 |
| 1929 | 7 | 150 | 229 | 14 | 137 | 128 |  |  | 33 | 626 |
| 1930 | 6 | 144 | 395 | 7 | 181 | 96 |  |  | 41 | 781 |
| 1931 | 3 | 146 | 358 | 10 | 237 | 80 |  |  | 42 | 788 |
| 1932 | 2 | 139 | 286 | 11 | 160 | 60 |  |  | 33 | 623 |
| 1933 | 1 | 159 | 226 | 14 | 118 | 51 |  |  | 28 | 539 |
| 1934 | 3 | 127 | 235 | 39 | 102 | 26 |  |  | 27 | 503 |
| 1935 | 4 | 124 | 301 | 55 | 75 | 16 |  |  | 29 | 541 |
| 1936 | 5 | 120 | 366 | 56 | 25 | 14 |  |  | 29 | 552 |
| 1937 | 3 | 145 | 294 | 51 | 31 | 13 |  |  | 27 | 508 |
| 1938 | 14 | 78 | 258 | 22 | 25 | 7 |  |  | 20 | 371 |
| 1939 | 19 | 54 | 186 | 28 | 28 | 18 |  |  | 17 | 299 |
| 1940 | 15 | 64 | 232 | 42 | 31 | 13 |  |  | 20 | 362 |
| 1941 | 12 | 63 | 204 | 41 | 41 | 8 |  |  | 92 | 268 |
| 1942 | 7 | 11 | 63 | 13 | 15 | 2 |  |  | 55 | 52 |
| 1943 | 35 | 30 | 66 | 13 | 18 | 1 |  |  | 123 | 32 |
| 1944 | 142 | 10 | 71 | 8 | 2 | 1 |  |  | 210 | 9 |
| 1945 | 301 | 28 | 114 | 15 | 5 | 1 |  |  | 418 | 16 |
| 1946 | 227 | 33 | 108 | 17 | 16 | 2 |  |  | 362 | 18 |
| 1947 | 141 | 32 | 147 | 9 | 26 | 2 |  |  | 322 | 22 |
| 1948 | 90 | 41 | 159 | 12 | 40 | 5 |  |  | 313 | 26 |
| 1949 | 72 | 78 | 130 | 21 | 52 | 7 |  |  | 325 | 29 |
| 1950 | 88 | 149 | 235 | 29 | 59 | 7 |  |  | 510 | 48 |
| 1951 | 116 | 331 | 307 | 42 | 66 | 3 |  |  | 778 | 75 |
| 1952 | 62 | 538 | 357 | 37 | 44 | 2 |  |  | 935 | 98 |
| 1953 | 76 | 627 | 416 | 30 | 36 | 3 |  |  | 1069 | 111 |
| 1954 | 72 | 448 | 500 | 90 | 49 | 3 |  |  | 1037 | 118 |
| 1955 | 69 | 235 | 402 | 523 | 40 | 10 |  |  | 1149 | 122 |
| 1956 | 74 | 442 | 542 | 399 | 52 | 7 |  |  | 1344 | 163 |
| 1957 | 95 | 423 | 528 | 520 | 46 | 6 |  |  | 1434 | 174 |
| 1958 | 91 | 630 | 565 | 487 | 46 | 28 |  |  | 1504 | 326 |
| 1959 | 77 | 740 | 454 | 229 | 53 | 19 |  |  | 1286 | 271 |
| 1960 | 74 | 494 | 344 | 437 | 48 | 18 |  |  | 1258 | 149 |
| 1961 | 50 | 334 | 312 | 359 | 35 | 20 |  |  | 956 | 146 |
| 1962 | 48 | 330 | 297 | 357 | 34 | 14 |  |  | 917 | 156 |
| 1963 | 72 | 318 | 219 | 346 | 44 | 22 | 14.9 |  | 917 | 111 |
| 1964 | 43 | 263 | 210 | 245 | 43 | 18 | 0.1 |  | 711 | 106 |
| 1965 | 72 | 234 | 266 | 261 | 53 | 28 | 0 |  | 765 | 136 |
| 1966 | 62 | 236 | 360 | 331 | 50 | 30 | 0 | 985 | 1905 | 140 |
| 1967 | 92 | 124 | 283 | 410 | 57 | 38 | 0.3 | 1634 | 2498 | 127 |
| 1968 | 121 | 83 | 278 | 359 | 56 | 29 | 0 | 671 | 1468 | 113 |
| 1969 | 140 | 86 | 248 | 332 | 54 | 20 | 0 | 53 | 810 | 104 |
| 1970 | 194 | 107 | 362 | 316 | 42 | 27 | 0 | 1 | 908 | 114 |
| 1971 | 260 | 171 | 313 | 238 | 50 | 38 | 0 | 2 | 867 | 155 |
| 1972 | 365 | 220 | 498 | 394 | 77 | 69 | 0 | 26 | 1372 | 215 |
| 1973 | 317 | 416 | 1054 | 464 | 73 | 91 | 0 | 907 | 2893 | 371 |
| 1974 | 469 | 279 | 838 | 389 | 51 | 126 | 0.2 | 1403 | 3193 | 282 |
| 1975 | 561 | 272 | 865 | 360 | 70 | 74 | 1.5 | 734 | 2588 | 260 |
| 1976 | 713 | 279 | 714 | 311 | 50 | 55 | 0 | 529 | 2335 | 210 |
| 1977 | 545 | 196 | 767 | 172 | 25 | 29 | 0 |  | 1491 | 167 |
| 1978 | 618 | 284 | 500 | 45 | 33 | 9 | 0 |  | 1293 | 169 |
| 1979 | 1005 | 417 | 694 | 51 | 56 | 12 | 0 |  | 2004 | 177 |

Table 6: Estimates of chilipepper landings by region and gear type in California area (based on CalCOM), including Oregon (based on PacFIN), 1978-2006. Excludes 2002-2006 discards.

| year | Eureka | San Francisco | Monterey | Santa Barbara | Los <br> Angeles | San Diego | Oregon | Trawl | Hook-line | Net |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1978 | 618 | 284 | 500 | 45 | 33 | 9 | 0 | 1293 | 169 | 169 |
| 1979 | 1005 | 417 | 694 | 51 | 56 | 12 | 0 | 2004 | 177 | 177 |
| 1980 | 783 | 835 | 1157 | 31 | 52 | 5 | 0 | 2721 | 96 | 45 |
| 1981 | 713 | 874 | 772 | 32 | 68 | 23 | 23.4 | 2295 | 139 | 71 |
| 1982 | 369 | 508 | 1087 | 37 | 75 | 23 | 23.2 | 1681 | 356 | 85 |
| 1983 | 558 | 950 | 717 | 11 | 38 | 22 | 9.8 | 1879 | 80 | 345 |
| 1984 | 573 | 1141 | 908 | 43 | 81 | 29 | 2.1 | 2448 | 98 | 231 |
| 1985 | 421 | 872 | 1386 | 19 | 91 | 35 | 2.1 | 1807 | 279 | 739 |
| 1986 | 404 | 1353 | 940 | 29 | 28 | 6 | 1.1 | 1269 | 331 | 1161 |
| 1987 | 506 | 522 | 827 | 59 | 21 | 11 | 0.5 | 1314 | 173 | 461 |
| 1988 | 741 | 689 | 889 | 65 | 11 | 5 | 0.2 | 1778 | 333 | 289 |
| 1989 | 721 | 989 | 1210 | 193 | 30 | 3 | 4.5 | 2363 | 426 | 361 |
| 1990 | 926 | 1174 | 722 | 95 | 1 | 2 | 2.3 | 2317 | 232 | 373 |
| 1991 | 814 | 1411 | 774 | 155 | 10 | 1 | 14 | 2229 | 618 | 332 |
| 1992 | 377 | 1489 | 717 | 63 | 15 | 6 | 13.1 | 1330 | 1053 | 297 |
| 1993 | 595 | 963 | 761 | 41 | 3 | 7 | 6.1 | 1282 | 861 | 233 |
| 1994 | 498 | 608 | 723 | 13 | 1 | 3 | 13.9 | 1267 | 485 | 108 |
| 1995 | 606 | 564 | 819 | 8 | 3 | 4 | 9.5 | 1595 | 325 | 94 |
| 1996 | 451 | 606 | 748 | 19 | 2 | 4 | 9.3 | 1528 | 254 | 58 |
| 1997 | 486 | 840 | 681 | 17 | 4 | 2 | 7.3 | 1614 | 339 | 83 |
| 1998 | 319 | 644 | 449 | 2 | 3 | 1 | 5.8 | 1138 | 209 | 78 |
| 1999 | 411 | 358 | 175 | 2 | 1 | 3 | 3.3 | 839 | 104 | 10 |
| 2000 | 177 | 213 | 68 | 1 | 0 | 0 | 0.7 | 403 | 51 | 6 |
| 2001 | 116 | 144 | 72 | 0 | 1 | 0 | 132.7 | 436 | 25 | 5 |
| 2002 | 67 | 61 | 37 | 0 | 0 | 0 | 0.3 | 162 | 3 | 0 |
| 2003 | 10 | 2 | 5 | 0 | 0 | 0 | 0.7 | 18 | 0 | 0 |
| 2004 | 38 | 18 | 9 | 0 | 0 | 0 | 0.2 | 61 | 3 | 1 |
| 2005 | 43 | 11 | 8 | 0 | 0 | 0 | 0.7 | 60 | 3 | 0 |
| 2006 | 19 | 14 | 10 | 0 | 0 | 0 | 0.1 | 37 | 6 | 0 |

Table 7: RecFIN catch information for chilipepper rockfish, 1980-2006.

| Private/Rental |  |  |  |  |  |  |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
|  | North | South | CPFV 1000s |  | Total metric tons |  | Mean weight (kg) |  |
| North | South | North | South | North | South |  |  |  |
| 1980 | 0 | 50 | 50 | 385 | 30 | 362 | 0.60 | 0.83 |
| 1981 | 0 | 27 | 105 | 252 | 61 | 210 | 0.58 | 0.75 |
| 1982 | 0 | 36 | 181 | 246 | 178 | 192 | 0.98 | 0.68 |
| 1983 | 1 | 6 | 110 | 100 | 100 | 60 | 0.90 | 0.57 |
| 1984 | 0 | 3 | 201 | 28 | 127 | 19 | 0.63 | 0.60 |
| 1985 | 2 | 3 | 218 | 253 | 156 | 202 | 0.70 | 0.79 |
| 1986 | 21 | 6 | 342 | 183 | 276 | 110 | 0.76 | 0.58 |
| 1987 | 12 | 6 | 146 | 6 | 109 | 3 | 0.69 | 0.23 |
| 1988 | 14 | 25 | 679 | 51 | 264 | 26 | 0.38 | 0.35 |
| 1989 | 15 | 21 | 289 | 195 | 150 | 95 | 0.49 | 0.44 |
| 1990 | 15 | 23 | 261 | 159 | 114 | 74 |  |  |
| 1991 | 8 | 25 | 232 | 122 | 79 | 52 |  |  |
| 1992 | 5 | 28 | 203 | 86 | 43 | 31 |  |  |
| 1993 | 15 | 30 | 174 | 50 | 7 | 10 | 0.50 | 0.32 |
| 1994 | 0 | 37 | 146 | 14 | 0 | 17 | 0.09 | 0.34 |
| 1995 | 3 | 26 | 117 | 2 | 2 | 5 | 0.62 | 0.21 |
| 1996 | 1 | 20 | 88 | 1 | 21 | 10 | 0.48 | 0.45 |
| 1997 | 0 | 1 | 1 | 1 | 73 | 1 | 0.82 | 0.40 |
| 1998 | 0 | 6 | 24 | 9 | 1 | 4 | 0.75 | 0.61 |
| 1999 | 0 | 12 | 49 | 9 | 18 | 6 | 0.75 | 0.28 |
| 2000 | 1 | 9 | 50 | 7 | 31 | 8 | 0.63 | 0.44 |
| 2001 | 1 | 6 | 28 | 11 | 51 | 1 | 1.01 | 0.16 |
| 2002 | 0 | 3 | 5 | 14 | 6 | 6 | 0.97 | 0.37 |
| 2003 | 0 | 0 | 0 | 0 | 0 | 0 | 0.37 | 0.38 |
| 2004 | 0 | 0 | 0 | 0 | 15 | 0 | 6 |  |
| 2005 | 0 | 0 | 0 | 8 | 0 | 4 | 0.07 | 0.43 |
| 2006 | 0 | 0 | 0 | 4 | 0 | 1 | 0.07 | 0.34 |

Table 8: Reconstructed catches of all rockfish based on CPFV logs and estimated catches of chilipepper rockfish (1000s fish, tons), 1928-1979, based on interpolated species composition and average weight information.

|  | All rockfish <br> Reported CPFV |  | All rockfish Expanded CPFV |  | Chilipepper <br> Private (1000s) |  | Chilipepper CPFV (1000s) |  | Chilipepper <br> Total Tons |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | North | South | North | South | North | South | North | South | North | South |
| 1928 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1929 | 18 | 8 | 34 | 15 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1930 | 36 | 15 | 67 | 30 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1931 | 54 | 23 | 101 | 45 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1932 | 72 | 30 | 135 | 60 | 0 | 0 | 0 | 1 | 0 | 0 |
| 1933 | 90 | 38 | 168 | 75 | 0 | 0 | 0 | 1 | 0 | 1 |
| 1934 | 108 | 46 | 202 | 90 | 0 | 0 | 0 | 1 | 0 | 1 |
| 1935 | 126 | 53 | 236 | 105 | 0 | 0 | 0 | 2 | 0 | 1 |
| 1936 | 144 | 61 | 270 | 120 | 0 | 0 | 0 | 2 | 0 | 2 |
| 1937 | 171 | 72 | 320 | 143 | 0 | 0 | 0 | 3 | 0 | 2 |
| 1938 | 168 | 71 | 314 | 140 | 0 | 0 | 0 | 3 | 0 | 2 |
| 1939 | 147 | 62 | 275 | 123 | 0 | 0 | 0 | 3 | 0 | 2 |
| 1940 | 211 | 90 | 396 | 177 | 0 | 0 | 1 | 5 | 0 | 4 |
| 1941 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1942 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1943 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1944 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1945 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1946 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1947 | 148 | 46 | 277 | 91 | 0 | 0 | 1 | 4 | 0 | 3 |
| 1948 | 295 | 116 | 553 | 228 | 0 | 1 | 1 | 11 | 1 | 8 |
| 1949 | 383 | 188 | 716 | 372 | 0 | 2 | 2 | 18 | 1 | 14 |
| 1950 | 467 | 213 | 873 | 420 | 0 | 3 | 2 | 21 | 2 | 16 |
| 1951 | 533 | 189 | 997 | 374 | 0 | 4 | 3 | 20 | 2 | 15 |
| 1952 | 464 | 242 | 868 | 479 | 0 | 4 | 2 | 26 | 2 | 20 |
| 1953 | 395 | 301 | 739 | 595 | 0 | 5 | 2 | 34 | 1 | 26 |
| 1954 | 491 | 658 | 919 | 1301 | 0 | 6 | 3 | 78 | 2 | 59 |
| 1955 | 585 | 1153 | 1095 | 2278 | 0 | 7 | 3 | 142 | 2 | 107 |
| 1956 | 653 | 1384 | 1223 | 2734 | 0 | 7 | 4 | 176 | 3 | 133 |
| 1957 | 645 | 767 | 1207 | 1516 | 0 | 8 | 4 | 101 | 3 | 77 |
| 1958 | 1052 | 517 | 1968 | 1021 | 0 | 9 | 6 | 71 | 5 | 53 |
| 1959 | 879 | 300 | 1645 | 593 | 0 | 10 | 5 | 42 | 4 | 32 |
| 1960 | 679 | 307 | 1271 | 606 | 0 | 10 | 4 | 45 | 3 | 34 |
| 1961 | 514 | 348 | 961 | 689 | 0 | 11 | 5 | 52 | 3 | 40 |
| 1962 | 589 | 339 | 1102 | 670 | 0 | 12 | 7 | 52 | 5 | 40 |
| 1963 | 609 | 346 | 1141 | 684 | 0 | 13 | 10 | 55 | 7 | 42 |
| 1964 | 462 | 488 | 864 | 964 | 0 | 13 | 9 | 80 | 6 | 60 |
| 1965 | 718 | 631 | 1345 | 1246 | 0 | 14 | 16 | 106 | 12 | 80 |
| 1966 | 773 | 940 | 1447 | 1858 | 0 | 15 | 20 | 163 | 14 | 123 |
| 1967 | 760 | 1158 | 1423 | 2288 | 0 | 16 | 22 | 205 | 16 | 155 |
| 1968 | 800 | 1274 | 1497 | 2517 | 0 | 16 | 26 | 232 | 19 | 175 |
| 1969 | 843 | 1097 | 1578 | 2167 | 0 | 17 | 30 | 205 | 22 | 155 |
| 1970 | 1047 | 1532 | 1960 | 3027 | 0 | 18 | 41 | 293 | 29 | 221 |
| 1971 | 803 | 1399 | 1504 | 2764 | 0 | 19 | 34 | 274 | 24 | 207 |
| 1972 | 1098 | 1827 | 2054 | 3609 | 0 | 19 | 50 | 366 | 36 | 276 |
| 1973 | 1391 | 2137 | 2603 | 4223 | 0 | 20 | 68 | 438 | 49 | 331 |
| 1974 | 1466 | 2552 | 2745 | 5042 | 0 | 21 | 76 | 569 | 55 | 430 |
| 1975 | 1396 | 2516 | 2613 | 4971 | 0 | 22 | 77 | 428 | 56 | 323 |
| 1976 | 1580 | 1978 | 2957 | 3909 | 0 | 22 | 93 | 635 | 67 | 480 |
| 1977 | 1384 | 1792 | 2590 | 3541 | 0 | 23 | 86 | 492 | 62 | 372 |
| 1978 | 1199 | 1674 | 2245 | 3307 | 0 | 24 | 78 | 514 | 57 | 389 |
| 1979 | 1321 | 2319 | 2472 | 4583 | 0 | 38 | 91 | 562 | 65 | 425 |

Table 9: Number of subsamples (trips) and fish measured for RecFIN length composition data

| Number of subsamples |  |  | Number of fish measured |  |  |  |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: |
|  | N.Cal | S.Cal | Coastwide | N.Cal | S.Cal | Coastwide |
| 1980 | 18 | 32 | 50 | 88 | 303 | 391 |
| 1981 | 6 | 41 | 47 | 90 | 697 | 787 |
| 1982 | 10 | 49 | 59 | 204 | 414 | 618 |
| 1983 | 12 | 33 | 45 | 213 | 433 | 646 |
| 1984 | 41 | 49 | 90 | 675 | 111 | 786 |
| 1985 | 86 | 52 | 138 | 1475 | 537 | 2012 |
| 1986 | 78 | 37 | 115 | 1715 | 383 | 2098 |
| 1987 | 21 | 1 | 22 | 384 | 10 | 394 |
| 1988 | 67 | 5 | 72 | 875 | 53 | 928 |
| 1989 | 20 | 9 | 29 | 658 | 254 | 912 |
| 1994 |  | 5 | 5 |  | 31 | 31 |
| 1995 | 5 |  | 5 | 149 |  | 149 |
| 1996 | 18 | 2 | 20 | 550 | 6 | 556 |
| 1997 | 15 |  | 15 | 590 |  | 590 |
| 1998 | 6 |  | 6 | 263 |  | 263 |
| 1999 | 28 | 19 | 47 | 528 | 53 | 581 |
| 2000 | 9 | 22 | 31 | 194 | 82 | 276 |
| 2001 | 9 | 7 | 16 | 210 | 89 | 299 |
| 2002 | 11 | 7 | 18 | 140 | 85 | 225 |
| 2004 |  | 41 | 41 |  | 233 | 233 |
| 2005 |  | 16 | 16 |  | 53 | 53 |

Table 10: Trawl logbook CPUE time series developed by Ralston et al. (1998) and Ralston (1999)

| Ralston cv year et al. 1998 (assumed) |  |  | catch weighted | SE | CV | area weighted | SE |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1980 | 249 | 0.1 |  |  |  |  |  |
| 1981 | 150 | 0.1 |  |  |  |  |  |
| 1982 | 121 | 0.1 | 132 | 49.8 | 0.38 | 95 | 32.6 |
| 1983 | 116 | 0.1 | 35 | 13.1 | 0.38 | 35 | 11.4 |
| 1984 | 91 | 0.1 | 90 | 27 | 0.30 | 57 | 16.4 |
| 1985 | 88 | 0.1 | 101 | 31.3 | 0.31 | 51 | 13.1 |
| 1986 | 76 | 0.1 | 57 | 17.7 | 0.31 | 35 | 10 |
| 1987 | 116 | 0.1 | 103 | 30.3 | 0.30 | 55 | 14.2 |
| 1988 | 158 | 0.1 | 175 | 59.2 | 0.34 | 77 | 18.6 |
| 1989 | 172 | 0.1 | 92 | 28.4 | 0.31 | 66 | 18 |
| 1990 | 149 | 0.1 | 103 | 31.8 | 0.31 | 74 | 20 |
| 1991 | 146 | 0.1 | 131 | 41.3 | 0.32 | 70 | 17 |
| 1992 | 109 | 0.1 | 120 | 45.8 | 0.38 | 45 | 11.5 |
| 1993 | 80 | 0.1 | 69 | 19 | 0.27 | 45 | 11 |
| 1994 | 112 | 0.1 | 103 | 32.6 | 0.32 | 51 | 13.6 |
| 1995 | 126 | 0.1 | 119 | 34.5 | 0.29 | 59 | 15.6 |
| 1996 | 96 | 0.1 | 95 | 28.1 | 0.29 | 45 | 11.7 |

Table 11: Number of subsamples for length comp data, and numbers of length and age observations by fishery

|  | Subsamples (length) |  |  | Length measurements |  |  | Age measurements |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Trawl | Hk-line | Net | Trawl | Hk-line | Net | trawl | Hk-line | net |
| 1978 | 147 |  |  | 1560 | 4 |  | 559 |  |  |
| 1979 | 110 |  |  | 1860 | 307 |  | 330 |  |  |
| 1980 | 191 | 1 |  | 1590 | 85 |  | 841 | 2 |  |
| 1981 | 125 |  |  | 955 | 109 |  | 701 |  |  |
| 1982 | 195 | 20 |  | 1856 | 227 |  | 1220 |  |  |
| 1983 | 275 | 8 | 24 | 2701 | 79 | 211 | 2305 | 8 | 68 |
| 1984 | 305 | 9 | 68 | 5186 | 94 | 660 | 3574 |  | 42 |
| 1985 | 338 | 14 | 155 | 7153 | 356 | 1090 | 3269 | 100 | 266 |
| 1986 | 219 | 8 | 113 | 4076 | 213 | 824 | 2008 | 173 | 414 |
| 1987 | 211 | 9 | 92 | 4433 | 135 | 700 | 2529 | 36 | 367 |
| 1988 | 199 |  | 70 | 4669 | 122 | 551 | 2428 | 5 | 220 |
| 1989 | 183 | 16 | 82 | 4582 | 284 | 650 | 2524 | 9 | 311 |
| 1990 | 204 | 16 | 99 | 5026 | 80 | 953 | 1692 | 15 | 443 |
| 1991 | 208 | 41 | 35 | 7632 | 1801 | 483 | 1600 | 424 | 96 |
| 1992 | 132 | 84 | 68 | 4208 | 2570 | 946 | 2081 | 745 | 406 |
| 1993 | 126 | 87 | 35 | 4630 | 3584 | 966 | 2001 | 434 | 188 |
| 1994 | 117 | 86 | 47 | 3898 | 3615 | 931 | 742 | 251 | 253 |
| 1995 | 114 | 23 | 32 | 3747 | 841 | 742 | 1306 | 249 | 60 |
| 1996 | 116 | 41 | 21 | 3327 | 1138 | 342 | 803 | 189 | 37 |
| 1997 | 136 | 38 | 14 | 4537 | 1367 | 439 | 1718 | 209 | 63 |
| 1998 | 123 | 38 | 11 | 3109 | 886 | 269 | 2135 | 322 | 93 |
| 1999 | 84 | 11 |  | 3030 | 435 |  | 2091 | 165 |  |
| 2000 | 50 | 9 |  | 1706 | 364 |  | 998 | 161 |  |
| 2001 | 58 | 12 |  | 1996 | 401 |  | 767 | 128 |  |
| 2002 | 54 | 3 |  | 1832 | 64 |  | 1029 | 38 | 1 |
| 2003 | 18 |  |  | 533 | 6 |  | 309 | 3 |  |
| 2004 | 54 |  |  | 1743 |  |  | 949 |  |  |
| 2005 | 20 |  |  | 452 |  |  | 349 |  |  |
| 2006 | 31 | 3 |  | 650 | 70 |  |  |  |  |

Table 12: Number of trips by year and average depth bin for the CPFV observer dataset.

| YEAR | $0-19$ | $20-39$ | $40-59$ | $60-79$ | $80-99$ | $>100$ |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1987 | 1 | 14 | 36 | 21 | 17 | 1 |
| 1988 | 23 | 75 | 62 | 25 | 21 | 4 |
| 1989 | 16 | 77 | 83 | 26 | 25 | 4 |
| 1990 | 3 | 25 | 33 | 8 | 4 | 1 |
| 1991 | 9 | 34 | 32 | 9 |  | 1 |
| 1992 | 28 | 64 | 110 | 22 | 6 |  |
| 1993 | 33 | 93 | 81 | 35 | 5 | 1 |
| 1994 | 35 | 89 | 85 | 25 | 3 |  |
| 1995 | 32 | 89 | 86 | 8 | 3 |  |
| 1996 | 46 | 94 | 76 | 11 | 2 |  |
| 1997 | 54 | 77 | 88 | 20 | 5 |  |
| 1998 | 40 | 72 | 46 | 13 |  |  |

Table 13: Total number of chilipepper caught (by mean depth bin)

|  | $0-19$ | $20-39$ | $40-59$ | $60-79$ | $80-99$ | $>100$ |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1987 |  | 1 | 557 | 1770 | 3573 | 295 |
| 1988 | 3 |  | 493 | 3267 | 2973 | 556 |
| 1989 |  |  | 355 | 2351 | 3004 | 388 |
| 1990 |  |  | 150 | 193 | 442 | 218 |
| 1991 |  | 1 | 60 | 173 | 6 | 8 |
| 1992 |  | 0 | 454 | 852 | 56 |  |
| 1993 |  |  | 181 | 1504 | 457 | 161 |
| 1994 |  | 3 | 186 | 1069 | 111 |  |
| 1995 | 15 | 12 | 45 | 320 | 82 |  |
| 1996 |  | 3 | 33 | 413 | 216 |  |
| 1997 |  |  | 18 | 376 | 91 |  |
| 1998 |  | 3 | 3 | 189 |  |  |

Table 14: AIC scores for the different fixed effect models considered in the recreational observer database CPUE series

| Model | Binomial | Gamma |
| :--- | ---: | ---: |
| Year | 1038 | 442 |
| Depth | 704 | 470 |
| Block | 846 | 436 |
| Year+depth | 696 | 417 |
| Year+block | 834 | 395 |
| Year+depth+block | 656 | 373 |
| Year+depth+block+depth:block | 672 | 379 |
| Null deviance | 1059 | 561 |

Table 15: Triennial trawl survey area-swept biomass estimates by depth and INPFC area. Dashes denote area-strata combinations in which no chilipepper were encountered, zeros denote area-strata combinations in which the total biomass was estimated at less than 0.5 ton, and empty cells denote strata that did not have any survey effort.

|  |  | Columbia |  | Eureka |  | Monterey |  | Conception |  |  | Total <br> CV |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | Depth (m) | Biomass | CV | Biomass | CV | Biomass | CV | Biomass | CV | Biomass |  |
| 1977 | 91-183 | - | - | - | - | 4755 | 0.38 | 94 | 0.76 | 4850 | 0.37 |
|  | 184-366 | - | - | - | - | 4942 | 0.35 | 148 | 0.49 | 5090 | 0.34 |
|  | 367-475 | - | - | - | - | 0 | 0.72 | 1 | 1.00 | 1 | 0.81 |
|  | 91-475 | - | - | - | - | 9697 | 0.26 | 243 | 0.42 | 9940 | 0.25 |
| 1980 | 55-183 | 129 | 0.62 | 901 | 1.00 | 12740 | 0.63 |  |  | 13770 | 0.59 |
|  | 184-366 | 0 | - | 0 | - | 904 | 0.43 |  |  | 904 | 0.43 |
|  | 55-366 | 129 | 0.62 | 901 | 1.00 | 13644 | 0.59 |  |  | 14674 | 0.55 |
| 1983 | 55-183 | 0 | - | 9 | 1.00 | 7113 | 0.62 |  |  | 7123 | 0.61 |
|  | 184-366 | 26 | 0.81 | 19 | 0.07 | 2379 | 0.39 |  |  | 2423 | 0.38 |
|  | 55-366 | 26 | 0.81 | 28 | 0.34 | 9492 | 0.47 |  |  | 9546 | 0.47 |
| 1986 | 55-183 | 0 | - | 2857 | 0.33 | 6596 | 0.32 |  |  | 9453 | 0.33 |
|  | 184-366 | 30 | 1.00 | 228 | 0.63 | 385 | 0.64 |  |  | 643 | 0.61 |
|  | 55-366 | 30 | 1.00 | 3175 | 0.30 | 7135 | 0.30 |  |  | 10340 | 0.30 |
| 1989 | 55-183 | 0 | 1.00 | 221 | 0.98 | 14563 | 0.34 | 1862 | 0.36 | 16646 | 0.30 |
|  | 184-366 | 219 | 0.97 | 67 | 1.00 | 2540 | 0.48 | 643 | 0.42 | 3470 | 0.37 |
|  | 55-366 | 220 | 0.97 | 288 | 0.79 | 17102 | 0.30 | 2505 | 0.29 | 20116 | 0.26 |
| 1992 | 55-183 | 0 | - | 5 | 0.94 | 6661 | 0.51 | 1284 | 0.48 | 7949 | 0.44 |
|  | 184-366 | 0 | - | 18 | 0.37 | 657 | 0.80 | 258 | 0.13 | 933 | 0.57 |
|  | 55-366 | 0 | - | 22 | 0.35 | 7318 | 0.47 | 1542 | 0.40 | 8882 | 0.40 |
| 1995 | 55-183 | 0 | - | 69 | 0.98 | 9640 | 0.31 | 299 | 0.38 | 10009 | 0.30 |
|  | 184-366 | 0 | 1.00 | 33 | 0.61 | 2321 | 0.38 | 1326 | 0.73 | 3681 | 0.37 |
|  | 367-500 | 0 | - | 0 | - | 2 | 0.81 | 2 | 0.66 | 4 | 0.55 |
|  | 55-500 | 0 | 1.00 | 102 | 0.69 | 11963 | 0.26 | 1627 | 0.60 | 13693 | 0.24 |
| 1998 | 55-183 | 0 | 1.00 | 3 | 0.83 | 10991 | 0.47 | 576 | 0.57 | 11570 | 0.45 |
|  | 184-366 | 12 | 0.79 | 235 | 0.83 | 5177 | 0.73 | 126 | 0.32 | 5550 | 0.69 |
|  | 367-500 | 0 | - | 1 | 1.00 | 0 | - | 0 | - | 1 | 1.00 |
|  | 55-500 | 12 | 0.78 | 239 | 0.82 | 16168 | 0.40 | 702 | 0.47 | 17121 | 0.38 |
| 2001 | 55-183 | 0 | - | 15 | 0.72 | 9270 | 0.38 | 13550 | 0.93 | 22835 | 0.58 |
|  | 184-366 | 1 | 0.62 | 60 | 0.99 | 4838 | 0.90 | 107 | 0.50 | 5006 | 0.87 |
|  | 367-500 | 0 | - | 0 | - | 1 | 1.00 | 1 | 1.00 | 3 | 0.71 |
|  | 55-500 | 1 | 0.62 | 76 | 0.80 | 14109 | 0.40 | 13658 | 0.93 | 27844 | 0.50 |
| 2004 | 55-183 | 0 | - | 67 | 0.52 | 31716 | 0.40 | 305 | 0.41 | 32088 | 0.39 |
|  | 184-366 | 4 | 0.88 | 22 | 0.38 | 6916 | 0.44 | 1896 | 0.62 | 8838 | 0.37 |
|  | 367-500 | 0 | - | 0 | - | 0 | - | 0 | - | 0 |  |
|  | 55-500 | 4 | 0.88 | 88 | 0.40 | 38632 | 0.34 | 2202 | 0.54 | 40927 | 0.32 |

Table 16: Comparison of triennial trawl survey indices generated by and core-area swept biomass and GLMM, with associated coefficients of variation.

|  | Core area-swept |  |  | GLMM |
| ---: | ---: | ---: | ---: | ---: |
|  | Biomass | CV | Index | CV |
| 1980 | 14674 | 0.55 | 4093 | 1.73 |
| 1983 | 9546 | 0.47 | 1884 | 2.11 |
| 1986 | 8704 | 0.32 | 1685 | 2.81 |
| 1989 | 17274 | 0.29 | 3313 | 0.86 |
| 1992 | 6774 | 0.5 | 27 | 1.73 |
| 1995 | 11307 | 0.27 | 2034 | 0.98 |
| 1998 | 16007 | 0.4 | 1004 | 0.92 |
| 2001 | 14103 | 0.4 | 964 | 0.79 |
| 2004 | 38444 | 0.34 | 3644 | 1.41 |

Table 17: NWFSC combined survey estimates of area-swept biomass and associated CVs by INPFC area and depth strata, 2003-2006.

| Year | Depth (m) | Conception |  |  | Monterey |  | Eureka | Columbia |  | Total Biomass |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Biomass |  | Biomass |  | mass | CV | mass | CV | Biomass | CV |
| 2003 | 55-183 | 1577 | 0.93 | 106395 | 0.54 | 1741 | 0.68 | 0 |  | 109713 |  |
|  | 184-548 | 12751 | 0.92 | 6510 | 0.46 | 58 | 0.75 | 4 | 1.00 | 19323 |  |
|  | 55-548 | 14329 | 0.82 | 112905 | 0.51 | 1799 | 0.66 | 4 | 1.00 | 129037 | 0.46 |
| 2004 | 55-183 | 238 | 0.39 | 49594 | 0.49 | 4087 | 0.67 | 1747 | 1.00 | 55666 |  |
|  | 184-548 | 2915 | 0.50 | 24704 | 0.57 | 0 |  | 87 | 0.94 | 27705 |  |
|  | 55-548 | 3153 | 0.47 | 74298 | 0.38 | 4087 | 0.67 | 1834 | 0.95 | 83371 | 0.34 |
| 2005 | 55-183 | 1386 | 0.64 | 71694 | 0.73 | 3682 | 0.69 | 216 | 0.78 | 76978 |  |
|  | 184-548 | 4211 | 0.96 | 29388 | 0.40 | 2129 | 0.96 | 0 |  | 35728 |  |
|  | 55-548 | 5597 | 0.74 | 101082 | 0.53 | 5810 | 0.56 | 216 | 0.78 | 112706 | 0.48 |
| 2006 | 55-183 | 1282 | 0.89 | 54131 | 0.55 | 1543 | 0.74 | 13 | 1.00 | 56970 |  |
|  | 184-548 | 356 | 0.54 | 11133 | 0.45 | 56 | 0.92 | 693 | 0.71 | 12239 |  |
|  | 55-548 | 1638 | 0.70 | 65264 | 0.46 | 1600 | 0.71 | 706 | 0.69 | 69209 | 0.43 |

Table 18: Comparison of area-swept and GLMM biomass estimates for the Northwest Fisheries Science Center combined survey

|  | Area-Swept |  |  | GLMM |  |
| ---: | ---: | ---: | ---: | ---: | :---: |
|  | Bio | CV | Bio | CV |  |
| 2003 | 129037 | 0.46 | 3932 | 1.06 |  |
| 2004 | 83371 | 0.34 | 24559 | 2.06 |  |
| 2005 | 112706 | 0.48 | 9540 | 0.77 |  |
| 2006 | 69209 | 0.44 | 7384 | 0.69 |  |

Table 19: Indices of pelagic juvenile (age 0) rockfish abundance

|  | $\begin{array}{r} \text { core } \\ \text { index } \end{array}$ |  | design Index | deltaGLM |  | anova |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | jack.cv |  | CV | index | CV | Index | CV |
| 1983 |  |  |  |  |  |  |  |  |
| 1984 | 7.33 | 0.37 |  |  |  |  |  |  |
| 1985 | 8.12 | 0.46 |  |  |  |  |  |  |
| 1986 | 0.72 | 0.33 |  |  |  |  |  |  |
| 1987 | 13.22 | 0.35 |  |  |  |  |  |  |
| 1988 | 16.38 | 0.39 |  |  |  |  |  |  |
| 1989 | 0.39 | 0.48 |  |  |  |  |  |  |
| 1990 | 0.31 | 0.41 |  |  |  |  |  |  |
| 1991 | 0.98 | 0.34 |  |  |  |  |  |  |
| 1992 | 0.17 | 0.52 |  |  |  |  |  |  |
| 1993 | 10.33 | 0.30 |  |  |  |  |  |  |
| 1994 | 0.02 | 0.81 |  |  |  |  |  |  |
| 1995 | 0.25 | 0.61 |  |  |  |  |  |  |
| 1996 | 0.09 | 0.52 |  |  |  |  |  |  |
| 1997 | 0.13 | 0.74 |  |  |  |  |  |  |
| 1998 |  |  |  |  |  |  |  |  |
| 1999 | 0.21 | 0.43 |  |  |  |  |  |  |
| 2000 | 0.09 | 0.52 |  |  |  |  |  |  |
| 2001 | 0.85 | 0.34 | 1.51 | 0.21 | 0.24 | 0.39 | 1.72 | 0.04 |
| 2002 | 2.29 | 0.32 | 5.61 | 0.25 | 0.76 | 0.38 | 2.76 | 0.05 |
| 2003 | 1.01 | 0.41 | 2.06 | 0.32 | 0.35 | 0.40 | 1.57 | 0.04 |
| 2004 | 1.33 | 0.39 | 5.80 | 0.21 | 0.63 | 0.34 | 2.94 | 0.04 |
| 2005 |  |  | 0.21 | 0.44 | 0.03 | 0.60 | 0.87 | 0.03 |
| 2006 |  |  | 0.02 | 0.44 | 0.01 | 0.59 | 0.75 | 0.03 |

Table 20: Parameter point estimates and standard deviations for the base model (note that both the triennial length selectivity and the recreational CPUE age-selectivity curve parameters were fixed to enable estimation of the Hessian matrix).

| Parameter | value | std | parameter | value | std |
| :---: | :---: | :---: | :---: | :---: | :---: |
| In R0 | 10.45 | 0.05 | 1965 rec dev | -0.50 | 0.72 |
| K (1970-1979) | 0.32 | 0.06 | 1966 rec dev | -0.93 | 0.74 |
| K (1980-1988) | 0.25 | 0.02 | 1967 rec dev | 0.89 | 0.47 |
| K (1989-1991) | 0.23 | 0.04 | 1968 rec dev | 1.05 | 0.39 |
| K (1992-1998) | 0.20 | 0.04 | 1969 rec dev | -0.89 | 0.76 |
| K (1999-2006) | 0.26 | 0.04 | 1970 rec dev | 1.17 | 0.22 |
| Trawl sel inflection | 32.65 | 0.35 | 1971 rec dev | 0.60 | 0.26 |
| Trawl sel width 95\% inflection | 8.46 | 0.36 | 1972 rec dev | -1.66 | 0.62 |
| Hook sel inflection | 37.27 | 0.67 | 1973 rec dev | 1.47 | 0.08 |
| Hook sel width 95\% inflection | 7.20 | 0.60 | 1974 rec dev | -1.04 | 0.48 |
| Setnet sel peak | 59.43 | 3.46 | 1975 rec dev | 1.40 | 0.07 |
| Setnet sel top | -2.19 | 37616 | 1976 rec dev | -0.20 | 0.18 |
| Setnet sel asc-width | 4.99 | 0.18 | 1977 rec dev | -0.27 | 0.13 |
| Setnet sel desc-width | 1.98 | 9359 | 1978 rec dev | -0.42 | 0.14 |
| Setnet sel init | -44.77 | 51789 | 1979 rec dev | 0.87 | 0.06 |
| Setnet sel final | -13.05 | 150010 | 1980 rec dev | -0.38 | 0.12 |
| Rec sel peak | 41.25 | 0.85 | 1981 rec dev | -0.78 | 0.12 |
| Rec sel top | -15.76 | 1149.3 | 1982 rec dev | -1.78 | 0.23 |
| Rec sel asc-width | 4.92 | 0.12 | 1983 rec dev | -1.54 | 0.24 |
| Rec sel desc-width | 2.59 | 1.01 | 1984 rec dev | 1.95 | 0.04 |
| Rec sel init | -8.25 | 3.05 | 1985 rec dev | -0.74 | 0.20 |
| Rec sel final | -0.64 | 0.75 | 1986 rec dev | 0.57 | 0.08 |
| Triennial sel size inflect | 15.70 | fixed | 1987 rec dev | 0.39 | 0.10 |
| width 95\% inflect | 0.00 | fixed | 1988 rec dev | 0.71 | 0.09 |
| Combo sel size inflect | 13.34 | 12.74 | 1989 rec dev | 0.78 | 0.09 |
| Combo sel width 95\% inflect | 12.88 | 22.76 | 1990 rec dev | 0.02 | 0.14 |
| Rec CPUE sel peak | 39.34 | 0.61 | 1991 rec dev | 0.57 | 0.12 |
| Rec CPUE sel top | -6.00 | 0.10 | 1992 rec dev | -0.37 | 0.21 |
| Rec CPUE sel asc-width | 3.76 | 0.09 | 1993 rec dev | 0.97 | 0.12 |
| Rec CPUE sel desc-width | 3.45 | 1.50 | 1994 rec dev | -0.15 | 0.21 |
| Rec CPUEsel init | -7.66 | 0.63 | 1995 rec dev | 0.04 | 0.22 |
| Rec CPUE sel final | -1.32 | 2.32 | 1996 rec dev | -0.78 | 0.38 |
| Rec CPUE age sel peak | 1.11 | fixed | 1997 rec dev | -0.63 | 0.31 |
| Rec CPUE age sel top | -60.00 | fixed | 1998 rec dev | -0.09 | 0.32 |
| Rec CPUE age sel asc-width | -24.80 | fixed | 1999 rec dev | 2.42 | 0.12 |
| Rec CPUE age sel desc-width | -0.12 | fixed | 2000 rec dev | -1.32 | 0.57 |
| Rec CPUE age sel init | -33.55 | fixed | 2001 rec dev | 0.06 | 0.18 |
| Rec CPUE age sel final | -4.11 | fixed | 2002 rec dev | 0.40 | 0.18 |
|  |  |  | 2003 rec dev | -0.23 | 0.17 |
|  |  |  | 2004 rec dev | 0.33 | 0.17 |
|  |  |  | 2005 rec dev | -0.91 | 0.17 |
|  |  |  | 2006 rec dev | -1.07 | 0.17 |

Table 21a: Base model output 1892-1949.

| year | bio-all | bio-smry | SSB | depletion | recruits | total catch | expl. rate |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Unfished | 47214 | 45057 | 33390 | 1.00 | 34490 | 0 | 0.000 |
| 1892 | 47214 | 45057 | 33391 | 1.00 | 34490 | 217 | 0.005 |
| 1893 | 47013 | 44857 | 33200 | 0.99 | 34453 | 205 | 0.005 |
| 1894 | 46841 | 44688 | 33038 | 0.99 | 34421 | 193 | 0.004 |
| 1895 | 46699 | 44547 | 32904 | 0.99 | 34394 | 180 | 0.004 |
| 1896 | 46582 | 44432 | 32795 | 0.98 | 34373 | 171 | 0.004 |
| 1897 | 46486 | 44337 | 32706 | 0.98 | 34355 | 160 | 0.004 |
| 1898 | 46409 | 44261 | 32636 | 0.98 | 34341 | 151 | 0.003 |
| 1899 | 46348 | 44201 | 32582 | 0.98 | 34330 | 140 | 0.003 |
| 1900 | 46303 | 44156 | 32543 | 0.97 | 34322 | 155 | 0.004 |
| 1901 | 46247 | 44101 | 32494 | 0.97 | 34312 | 169 | 0.004 |
| 1902 | 46184 | 44039 | 32437 | 0.97 | 34300 | 185 | 0.004 |
| 1903 | 46112 | 43967 | 32372 | 0.97 | 34287 | 200 | 0.005 |
| 1904 | 46032 | 43889 | 32300 | 0.97 | 34272 | 215 | 0.005 |
| 1905 | 45946 | 43803 | 32222 | 0.97 | 34256 | 229 | 0.005 |
| 1906 | 45855 | 43713 | 32139 | 0.96 | 34239 | 244 | 0.006 |
| 1907 | 45759 | 43618 | 32051 | 0.96 | 34221 | 259 | 0.006 |
| 1908 | 45658 | 43518 | 31959 | 0.96 | 34201 | 274 | 0.006 |
| 1909 | 45552 | 43414 | 31862 | 0.95 | 34181 | 307 | 0.007 |
| 1910 | 45426 | 43289 | 31747 | 0.95 | 34157 | 342 | 0.008 |
| 1911 | 45279 | 43144 | 31611 | 0.95 | 34128 | 377 | 0.009 |
| 1912 | 45113 | 42980 | 31459 | 0.94 | 34095 | 411 | 0.010 |
| 1913 | 44931 | 42800 | 31292 | 0.94 | 34059 | 445 | 0.010 |
| 1914 | 44735 | 42606 | 31111 | 0.93 | 34020 | 479 | 0.011 |
| 1915 | 44525 | 42399 | 30919 | 0.93 | 33978 | 514 | 0.012 |
| 1916 | 44303 | 42180 | 30715 | 0.92 | 33933 | 666 | 0.016 |
| 1917 | 43960 | 41840 | 30397 | 0.91 | 33861 | 819 | 0.020 |
| 1918 | 43506 | 41391 | 29977 | 0.90 | 33765 | 973 | 0.024 |
| 1919 | 42950 | 40843 | 29462 | 0.88 | 33644 | 637 | 0.016 |
| 1920 | 42758 | 40656 | 29292 | 0.88 | 33604 | 664 | 0.016 |
| 1921 | 42560 | 40460 | 29118 | 0.87 | 33562 | 562 | 0.014 |
| 1922 | 42474 | 40376 | 29051 | 0.87 | 33545 | 508 | 0.013 |
| 1923 | 42445 | 40347 | 29037 | 0.87 | 33542 | 601 | 0.015 |
| 1924 | 42330 | 40233 | 28942 | 0.87 | 33519 | 560 | 0.014 |
| 1925 | 42260 | 40165 | 28888 | 0.87 | 33505 | 647 | 0.016 |
| 1926 | 42115 | 40021 | 28762 | 0.86 | 33474 | 889 | 0.022 |
| 1927 | 41757 | 39666 | 28434 | 0.85 | 33393 | 754 | 0.019 |
| 1928 | 41555 | 39468 | 28254 | 0.85 | 33347 | 739 | 0.019 |
| 1929 | 41386 | 39302 | 28105 | 0.84 | 33309 | 659 | 0.017 |
| 1930 | 41306 | 39223 | 28040 | 0.84 | 33292 | 822 | 0.021 |
| 1931 | 41081 | 39001 | 27839 | 0.83 | 33240 | 830 | 0.021 |
| 1932 | 40867 | 38790 | 27648 | 0.83 | 33190 | 656 | 0.017 |
| 1933 | 40834 | 38758 | 27627 | 0.83 | 33185 | 568 | 0.015 |
| 1934 | 40885 | 38809 | 27685 | 0.83 | 33200 | 531 | 0.014 |
| 1935 | 40965 | 38888 | 27770 | 0.83 | 33222 | 571 | 0.015 |
| 1936 | 40999 | 38921 | 27810 | 0.83 | 33233 | 583 | 0.015 |
| 1937 | 41017 | 38939 | 27833 | 0.83 | 33239 | 537 | 0.014 |
| 1938 | 41076 | 38997 | 27893 | 0.84 | 33254 | 394 | 0.010 |
| 1939 | 41262 | 39181 | 28071 | 0.84 | 33300 | 318 | 0.008 |
| 1940 | 41502 | 39418 | 28300 | 0.85 | 33359 | 386 | 0.010 |
| 1941 | 41658 | 39570 | 28447 | 0.85 | 33396 | 360 | 0.009 |
| 1942 | 41822 | 39732 | 28604 | 0.86 | 33435 | 107 | 0.003 |
| 1943 | 42206 | 40112 | 28965 | 0.87 | 33524 | 155 | 0.004 |
| 1944 | 42511 | 40412 | 29254 | 0.88 | 33594 | 219 | 0.005 |
| 1945 | 42725 | 40623 | 29460 | 0.88 | 33644 | 434 | 0.011 |
| 1946 | 42715 | 40611 | 29464 | 0.88 | 33645 | 380 | 0.009 |
| 1947 | 42754 | 40650 | 29506 | 0.88 | 33655 | 347 | 0.009 |
| 1948 | 42822 | 40716 | 29569 | 0.89 | 33670 | 347 | 0.009 |
| 1949 | 42883 | 40777 | 29627 | 0.89 | 33683 | 368 | 0.009 |

Table 21b: Base model output 1950-2007.

| Year | bio-all | bio-smry | SSB | depletion | rec | total catch | expl. rate |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1950 | 42920 | 40813 | 29662 | 0.89 | 33691 | 576 | 0.014 |
| 1951 | 42758 | 40652 | 29519 | 0.88 | 33658 | 870 | 0.021 |
| 1952 | 42330 | 40228 | 29141 | 0.87 | 33567 | 1055 | 0.026 |
| 1953 | 41761 | 39666 | 28637 | 0.86 | 33443 | 1207 | 0.030 |
| 1954 | 41096 | 39010 | 28048 | 0.84 | 33294 | 1215 | 0.031 |
| 1955 | 40479 | 38401 | 27505 | 0.82 | 33152 | 1381 | 0.036 |
| 1956 | 39756 | 37688 | 26875 | 0.80 | 32982 | 1643 | 0.044 |
| 1957 | 38842 | 36787 | 26079 | 0.78 | 32758 | 1687 | 0.046 |
| 1958 | 37961 | 35920 | 25314 | 0.76 | 32533 | 1889 | 0.053 |
| 1959 | 36963 | 34937 | 24442 | 0.73 | 32263 | 1593 | 0.046 |
| 1960 | 36325 | 34313 | 23892 | 0.72 | 32085 | 1443 | 0.042 |
| 1961 | 35879 | 33876 | 23524 | 0.70 | 31962 | 1146 | 0.034 |
| 1962 | 35748 | 33750 | 23431 | 0.70 | 31931 | 1118 | 0.033 |
| 1963 | 35652 | 33656 | 23370 | 0.70 | 31910 | 1077 | 0.032 |
| 1964 | 35596 | 33601 | 23347 | 0.70 | 31902 | 884 | 0.026 |
| 1965 | 35086 | 33727 | 23478 | 0.70 | 11737 | 993 | 0.029 |
| 1966 | 34339 | 33735 | 23473 | 0.70 | 7623 | 2182 | 0.065 |
| 1967 | 33633 | 31923 | 22447 | 0.67 | 46692 | 2796 | 0.088 |
| 1968 | 32115 | 28980 | 20755 | 0.62 | 53478 | 1775 | 0.061 |
| 1969 | 29870 | 27973 | 19569 | 0.59 | 7602 | 1090 | 0.039 |
| 1970 | 30621 | 28520 | 19029 | 0.57 | 59113 | 1273 | 0.045 |
| 1971 | 33863 | 30943 | 21323 | 0.64 | 34502 | 1253 | 0.040 |
| 1972 | 34608 | 33423 | 23118 | 0.69 | 3682 | 1899 | 0.057 |
| 1973 | 37977 | 35174 | 24162 | 0.72 | 85193 | 3644 | 0.104 |
| 1974 | 36701 | 33844 | 24005 | 0.72 | 6905 | 3960 | 0.117 |
| 1975 | 35964 | 33305 | 22406 | 0.67 | 77489 | 3228 | 0.097 |
| 1976 | 36092 | 33196 | 22459 | 0.67 | 15714 | 3092 | 0.093 |
| 1977 | 35209 | 34259 | 22631 | 0.68 | 14693 | 2091 | 0.061 |
| 1978 | 36770 | 35912 | 24114 | 0.72 | 12750 | 1934 | 0.054 |
| 1979 | 38241 | 36360 | 25500 | 0.76 | 47094 | 2725 | 0.075 |
| 1980 | 36490 | 34605 | 24919 | 0.75 | 13496 | 3255 | 0.094 |
| 1981 | 31887 | 31194 | 22019 | 0.66 | 8719 | 2776 | 0.089 |
| 1982 | 28876 | 28508 | 19682 | 0.59 | 3130 | 2492 | 0.087 |
| 1983 | 26269 | 26051 | 18125 | 0.54 | 3862 | 2465 | 0.095 |
| 1984 | 27234 | 23240 | 16495 | 0.49 | 122750 | 2923 | 0.126 |
| 1985 | 23721 | 19667 | 14284 | 0.43 | 7999 | 3182 | 0.162 |
| 1986 | 20941 | 19835 | 11548 | 0.35 | 27210 | 3147 | 0.159 |
| 1987 | 21602 | 20057 | 10969 | 0.33 | 22256 | 2059 | 0.103 |
| 1988 | 23163 | 21448 | 12593 | 0.38 | 32477 | 2691 | 0.125 |
| 1989 | 23808 | 21682 | 13242 | 0.40 | 35464 | 3395 | 0.157 |
| 1990 | 22382 | 20771 | 12573 | 0.38 | 16270 | 3110 | 0.150 |
| 1991 | 21653 | 20279 | 11919 | 0.36 | 27574 | 3311 | 0.163 |
| 1992 | 20340 | 19153 | 11258 | 0.34 | 10565 | 2753 | 0.144 |
| 1993 | 19649 | 18087 | 10540 | 0.32 | 39139 | 2393 | 0.132 |
| 1994 | 18583 | 16975 | 10036 | 0.30 | 12526 | 1877 | 0.111 |
| 1995 | 17872 | 17008 | 9812 | 0.29 | 15080 | 2021 | 0.119 |
| 1996 | 17127 | 16453 | 9589 | 0.29 | 6555 | 1870 | 0.114 |
| 1997 | 16307 | 15865 | 9489 | 0.28 | 7584 | 2110 | 0.133 |
| 1998 | 15209 | 14578 | 8968 | 0.27 | 12569 | 1430 | 0.098 |
| 1999 | 18866 | 13635 | 8666 | 0.26 | 153415 | 977 | 0.072 |
| 2000 | 18442 | 13573 | 9029 | 0.27 | 3708 | 499 | 0.037 |
| 2001 | 19149 | 18556 | 9536 | 0.29 | 15148 | 517 | 0.028 |
| 2002 | 24397 | 23175 | 12671 | 0.38 | 23831 | 329 | 0.014 |
| 2003 | 28205 | 27023 | 17040 | 0.51 | 14082 | 21 | 0.001 |
| 2004 | 31275 | 30022 | 20229 | 0.61 | 25895 | 236 | 0.008 |
| 2005 | 32553 | 31509 | 22146 | 0.66 | 7647 | 192 | 0.006 |
| 2006 | 32852 | 32405 | 23224 | 0.70 | 6645 | 127 | 0.004 |
| 2007 | 33619 | 32401 | 23827 | 0.71 | 32063 | n/a | n/a |

Table 22: Reference Points

|  |  | -95\% Confidence Limits |  |
| ---: | ---: | ---: | ---: |
| Unfished Stock | Estimate | Lower | Upper |
| Summary (1+) Biomass | 45057 |  |  |
| Spawning Biomass (SSB) | 33390 | 30138 | 36642 |
| Equilibrium recruitment | 34490 | 31131 | 37849 |


|  | SPR proxy MSY | SB $_{40 \%}$ | Estimated MSY |
| ---: | ---: | ---: | ---: |
| SPR | 0.50 | 0.45 | 0.43 |
| Fmult (2006) | 25.2 | 29.9 | 33.0 |
| Exploitation rate | 0.088 | 0.102 | 0.112 |
| Yield | 2099 | 2155 | 2164 |
| SSB at Equilibrium | 15482 | 21034 | 12126 |
| SSB/SSB ${ }_{0}$ | 0.46 | 0.40 | 0.36 |

Table 23: Decision table with 10 year forecast


Table 24: Likelihood values and reference points for the base model and 13 "jittered" base models

|  |  | BASE | Jittered | d models-> |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| SSB0 |  | 33390 | 33576 | 33756 | 31924 | 33483 | 32076 | 33390 | 33427 | 33776 | 32543 | 33845 | 32221 | 32268 | 33416 |
| R0 |  | 34490 | 34682 | 34868 | 32975 | 34586 | 33133 | 34490 | 34528 | 34888 | 33615 | 34960 | 33282 | 33331 | 34516 |
| Maximum gradient |  | 0.00057 | 0.00072 | 0.00006 | 0.00072 | 0.00062 | 0.00055 | 0.00085 | 0.00098 | 0.00037 | 0.00052 | 0.00050 | 0.00084 | 0.00079 | 0.00090 |
| Total Likelihood |  | 1972.2 | 1973.8 | 1978.5 | 2010.5 | 1978.2 | 2006.6 | 1972.2 | 1974.7 | 1974.3 | 2014.8 | 1975.8 | 2008.0 | 2013.7 | 1972.4 |
| Likelihood components |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| indices |  | 43.6 | 43.8 | 44.1 | 67.6 | 43.4 | 65.5 | 43.6 | 43.4 | 43.7 | 67.8 | 43.8 | 65.5 | 67.8 | 43.6 |
| length_comps |  | 430.1 | 431.0 | 436.2 | 453.6 | 435.5 | 450.6 | 430.1 | 432.3 | 428.2 | 457.1 | 433.0 | 451.8 | 457.8 | 430.2 |
| age_comps |  | 1479.0 | 1479.5 | 1478.8 | 1470.2 | 1479.7 | 1471.6 | 1479.0 | 1479.4 | 1482.7 | 1470.9 | 1479.6 | 1471.8 | 1468.9 | 1479.0 |
| Recruitment |  | 19.5 | 19.5 | 19.3 | 19.1 | 19.6 | 19.0 | 19.5 | 19.6 | 19.7 | 19.0 | 19.4 | 19.0 | 19.2 | 19.5 |
| Indices |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Fleet | lambda |  | surv_like |  |  |  |  |  |  |  |  |  |  |  |  |
| trawl | 1 | 9.9 | 9.8 | 9.8 | 9.9 | 9.9 | 9.9 | 9.9 | 10.0 | 10.1 | 9.9 | 9.8 | 9.9 | 9.9 | 9.9 |
| triennial | 1 | 8.7 | 8.7 | 8.7 | 8.2 | 8.9 | 7.9 | 8.7 | 8.7 | 8.6 | 8.2 | 8.8 | 8.0 | 8.3 | 8.7 |
| combined | 1 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 |
| coast juvenile | 1 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 |
| recreational CPUE | 1 | 23.8 | 24.0 | 24.4 | 48.2 | 23.4 | 46.3 | 23.8 | 23.5 | 23.7 | 48.4 | 24.0 | 46.4 | 48.3 | 23.8 |
| Length composition | lambda |  | ngth_like |  |  |  |  |  |  |  |  |  |  |  |  |
| trawl | 0.1 | 468.9 | 469.4 | 471.7 | 470.7 | 468.9 | 472.4 | 468.9 | 468.3 | 473.7 | 472.4 | 471.4 | 472.7 | 471.1 | 468.9 |
| hook | 0.1 | 171.9 | 171.9 | 173.1 | 189.2 | 170.1 | 188.5 | 171.9 | 170.7 | 169.1 | 188.8 | 171.8 | 188.4 | 189.3 | 171.9 |
| setnet | 0.1 | 228.7 | 228.6 | 225.7 | 235.9 | 228.0 | 235.3 | 228.7 | 229.6 | 188.1 | 230.6 | 225.8 | 233.4 | 234.8 | 228.6 |
| recreational | 1 | 126.1 | 126.8 | 127.9 | 126.2 | 126.5 | 126.0 | 126.1 | 125.8 | 126.5 | 129.1 | 128.1 | 127.1 | 126.4 | 126.1 |
| triennial | 1 | 146.4 | 146.3 | 147.4 | 146.9 | 146.6 | 146.8 | 146.4 | 146.3 | 146.9 | 147.4 | 147.1 | 146.8 | 146.8 | 146.4 |
| combined | 0.1 | 33.6 | 33.6 | 33.6 | 35.6 | 35.0 | 33.7 | 33.6 | 33.6 | 33.9 | 33.7 | 33.6 | 33.7 | 33.6 | 33.6 |
| recreational CPUE | 1 | 67.4 | 67.5 | 70.5 | 87.3 | 72.2 | 84.8 | 67.4 | 70.0 | 68.3 | 88.1 | 67.6 | 85.1 | 91.6 | 67.4 |
| Age composition | lambda |  | age_like |  |  |  |  |  |  |  |  |  |  |  |  |
| trawl | 1 | 672.7 | 673.3 | 672.9 | 664.6 | 673.4 | 666.3 | 672.7 | 672.9 | 671.6 | 665.7 | 673.6 | 666.7 | 663.9 | 672.7 |
| hook | 1 | 266.1 | 266.4 | 266.4 | 261.1 | 267.0 | 261.2 | 266.1 | 266.5 | 265.5 | 261.5 | 266.7 | 261.4 | 261.3 | 266.2 |
| setnet | 1 | 531.9 | 531.6 | 531.4 | 536.6 | 531.1 | 536.3 | 531.9 | 531.8 | 537.3 | 535.8 | 531.1 | 535.9 | 535.8 | 531.9 |
| combined | 1 | 8.2 | 8.2 | 8.2 | 7.9 | 8.2 | 7.9 | 8.2 | 8.2 | 8.3 | 7.9 | 8.2 | 7.9 | 7.9 | 8.2 |

Table 25: Select run results and likelihood components from profiles on alternative steepness and natural mortality values.

| Parameter Value (h and M) |  | $\mathrm{h}=0.21$ | $\mathrm{h}=0.34$ | $\mathrm{h}=0.57$ | $\mathrm{h}=0.81$ | $\mathrm{h}=0.99$ | $\mathrm{M}=0.12$ | $\mathrm{M}=0.14$ | $\mathrm{M}=0.16$ | $\mathrm{M}=0.18$ | $\mathrm{M}=0.2$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| SSB0 | SSB0 | 54233 | 40274 | 33390 | 30718 | 29667 | 34235 | 33933 | 33390 | 32606 | 32182 |
| R0 | R0 | 56019 | 41600 | 34490 | 31730 | 30645 | 20621 | 27096 | 34490 | 42718 | 52617 |
| Total Likelihood |  | 2009.5 | 1980.0 | 1972.2 | 1971.1 | 1970.9 | 2018.6 | 1983.8 | 1972.2 | 1977.8 | 1994.1 |
| Likelihood components |  |  |  |  |  |  |  |  |  |  |  |
| indices |  | 40.4 | 41.3 | 43.6 | 44.9 | 45.4 | 44.1 | 44.0 | 43.6 | 43.1 | 42.7 |
| length_comps |  | 442.9 | 434.1 | 430.1 | 428.8 | 428.5 | 444.0 | 434.7 | 430.1 | 428.1 | 429.1 |
| age_comps |  | 1481.3 | 1478.9 | 1479.0 | 1479.1 | 1479.0 | 1500.9 | 1482.3 | 1479.0 | 1488.3 | 1503.9 |
| Recruitment |  | 44.9 | 25.6 | 19.5 | 18.4 | 18.1 | 29.7 | 22.9 | 19.5 | 18.3 | 18.4 |
| Fleet | lambda | surv_like |  |  |  |  |  |  |  |  |  |
| trawl | 1 | 10.6 | 10.4 | 9.9 | 9.6 | 9.5 | 8.7 | 9.3 | 9.9 | 10.6 | 11.6 |
| triennial | 1 | 7.2 | 7.5 | 8.7 | 9.3 | 9.6 | 9.2 | 9.0 | 8.7 | 8.3 | 7.9 |
| combined | 1 | 1.1 | 1.1 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 | 1.1 |
| coast juvenile | 1 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.3 |
| recreational CPUE | 1 | 21.3 | 22.1 | 23.8 | 24.7 | 25.1 | 24.9 | 24.5 | 23.8 | 23.0 | 21.9 |
| Length composition lambda length_like |  |  |  |  |  |  |  |  |  |  |  |
| trawl | 0.1 | 474.5 | 470.7 | 468.9 | 468.0 | 467.7 | 476.7 | 469.2 | 468.9 | 474.6 | 489.4 |
| hook | 0.1 | 176.2 | 173.8 | 171.9 | 171.1 | 170.9 | 181.1 | 176.0 | 171.9 | 168.4 | 165.1 |
| setnet | 0.1 | 227.0 | 228.2 | 228.7 | 228.8 | 229.0 | 233.7 | 231.3 | 228.7 | 219.2 | 190.9 |
| recreational | 1 | 131.0 | 127.5 | 126.1 | 125.6 | 125.6 | 132.1 | 128.2 | 126.1 | 124.5 | 124.5 |
| Triennial | 1 | 152.5 | 148.0 | 146.4 | 146.0 | 145.8 | 151.9 | 148.3 | 146.4 | 146.2 | 147.3 |
| combined | 0.1 | 28.8 | 31.9 | 33.6 | 34.0 | 34.1 | 29.2 | 31.8 | 33.6 | 35.2 | 36.9 |
| recreational CPUE | 1 | 68.8 | 68.1 | 67.4 | 67.0 | 66.9 | 67.9 | 67.4 | 67.4 | 67.7 | 69.1 |
| Age composition lambda |  | age_like |  |  |  |  |  |  |  |  |  |
| Trawl |  | 669.1 | 670.7 | 672.7 | 673.3 | 673.6 | 695.4 | 677.8 | 672.7 | 676.8 | 686.4 |
| Hook |  | 265.8 | 266.2 | 266.1 | 265.9 | 265.8 | 273.5 | 269.2 | 266.1 | 263.7 | 262.3 |
| Setnet combined | 1 | 535.2 | 533.2 | 531.9 | 531.7 | 531.5 | 521.5 | 526.3 | 531.9 | 540.1 | 547.6 |
|  | 1 | 11.1 | 8.7 | 8.2 | 8.2 | 8.2 | 10.5 | 9.0 | 8.2 | 7.8 | 7.6 |

Table 26: Model sensitivity runs, sequentially remove data or alter total catches.

|  |  | BASE | no trawl cpue | no <br> triennial index, LFs | $\begin{array}{r} \text { no } \\ \text { combo } \\ \text { index, } \\ \text { LFs, AF } \\ \hline \end{array}$ | no juv survey | no rec cpue, LF's | no trawl сриe, LFs, AfsL | no hook LFs, AFsL | $\begin{array}{r} \text { no net } \\ \text { LFs, AFs } \\ \hline \end{array}$ | net sel. asymp. | K timeinvariant | $\begin{gathered} 2 x \\ \text { pre- } 1970 \text { r } \\ \text { catches } \end{gathered}$ | $0.5 x$ <br> pre-1970 catches |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| SSB0 |  | 33390 | 32958 | 32919 | 32273 | 33698 | 35285 | 33886 | 31160 | 35126 | 33510 | 39879 | 48079 | 25097 |
| R0 |  | 34490 | 34044 | 34003 | 33336 | 34808 | 36447 | 35003 | 32186 | 36284 | 34614 | 41193 | 49662 | 25924 |
| Maximum gradient |  | 0.00057 | 0.00046 | 0.00073 | 0.00059 | 0.00054 | 0.00080 | 0.00074 | 0.00093 | 0.00071 | 0.00095 | 0.00060 | 0.00079 | 0.00081 |
| Total Likelihood |  | 1972.2 | 1964.4 | 1851.6 | 2001.6 | 1961.9 | 1863.9 | 1179.7 | 1718.4 | 1394.8 | 1989.2 | 2067.1 | 2023.6 | 1981.1 |
| Likelihood compone indices |  | 43.6 | 31.7 | 58.3 | 66.3 | 43.1 | 21.4 | 17.3 | 61.1 | 45.5 | 45.9 | 54.2 | 75.5 | 41.2 |
| length_comps |  | 430.1 | 433.1 | 311.6 | 456.7 | 420.0 | 365.2 | 362.1 | 432.7 | 400.7 | 437.8 | 509.8 | 454.5 | 433.3 |
| age_comps |  | 1479.0 | 1480.2 | 1463.1 | 1459.8 | 1479.7 | 1456.8 | 782.0 | 1205.4 | 930.6 | 1486.2 | 1484.4 | 1475.8 | 1483.2 |
| Recruitment |  | 19.5 | 19.4 | 18.7 | 18.9 | 19.0 | 20.5 | 18.2 | 19.2 | 18.1 | 19.4 | 18.7 | 17.8 | 23.4 |
| Indices |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Fleet | lambda | surv_like |  |  |  |  |  |  |  |  |  |  |  |  |
| trawl | 1 | 9.9 | 0.0 | 9.1 | 9.7 | 10.0 | 8.9 | 0.0 | 12.0 | 9.9 | 10.2 | 9.2 | 8.7 | 10.8 |
| triennial | 1 | 8.7 | 8.3 | 0.0 | 8.1 | 8.5 | 11.2 | 7.2 | 7.0 | 8.9 | 8.5 | 8.7 | 11.1 | 7.5 |
| combined | 1 | 1.0 | 1.0 | 1.1 | 0.0 | 1.0 | 1.1 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 | 1.1 |
| coast juvenile | 1 | 0.2 | 0.2 | 0.0 | 0.2 | 0.0 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.3 | 0.2 |
| recreational CPUE | 1 | 23.8 | 22.1 | 48.1 | 48.3 | 23.5 | 0.0 | 8.9 | 41.0 | 25.5 | 25.9 | 35.1 | 54.5 | 21.7 |
| Length composition | lambda | ength_like |  |  |  |  |  |  |  |  |  |  |  |  |
| trawl | 0.1 | 468.9 | 467.7 | 482.6 | 470.9 | 468.8 | 485.3 | 0.0 | 473.4 | 453.4 | 470.3 | 679.5 | 467.1 | 472.9 |
| hook | 0.1 | 171.9 | 170.0 | 193.8 | 189.1 | 171.9 | 187.4 | 166.1 | 0.0 | 177.5 | 173.2 | 170.4 | 186.4 | 173.1 |
| setnet | 0.1 | 228.7 | 229.5 | 223.8 | 234.6 | 228.8 | 213.3 | 211.3 | 230.6 | 0.0 | 198.9 | 173.8 | 236.6 | 225.2 |
| recreational | 1 | 126.1 | 127.1 | 118.8 | 125.9 | 126.3 | 122.5 | 116.2 | 125.3 | 125.9 | 130.7 | 111.9 | 126.3 | 125.9 |
| triennial | 1 | 146.4 | 145.7 | 0.0 | 148.1 | 135.8 | 150.7 | 141.5 | 142.4 | 143.0 | 146.3 | 186.2 | 144.8 | 148.5 |
| combined | 0.1 | 33.6 | 33.5 | 42.9 | 0.0 | 35.5 | 33.9 | 32.1 | 34.0 | 33.9 | 33.6 | 59.2 | 35.6 | 32.7 |
| recreational CPUE | 1 | 67.4 | 70.2 | 98.4 | 93.2 | 67.4 | 0.0 | 63.4 | 91.2 | 65.2 | 73.1 | 103.4 | 90.9 | 68.6 |
| Age composition | lambda | age_like |  |  |  |  |  |  |  |  |  |  |  |  |
| trawl | 1 | 672.7 | 673.9 | 660.1 | 662.4 | 672.3 | 656.0 | 0.0 | 663.6 | 658.9 | 670.9 | 677.0 | 676.7 | 669.9 |
| hook | 1 | 266.1 | 266.3 | 259.5 | 261.2 | 266.2 | 259.0 | 276.8 | 0.0 | 263.6 | 265.8 | 272.6 | 260.4 | 266.5 |
| setnet | 1 | 531.9 | 531.9 | 534.9 | 536.2 | 532.0 | 533.6 | 498.6 | 534.2 | 0.0 | 541.2 | 526.1 | 530.9 | 538.1 |
| combined | 1 | 8.2 | 8.2 | 8.6 | 0.0 | 9.2 | 8.2 | 6.7 | 7.5 | 8.1 | 8.2 | 8.7 | 7.8 | 8.6 |




Figure 1 a (top) and 1 b (bottom): Externally fitted growth curves and size at age data for female and male chilipepper rockfish.


Figure 2 a (top) and 2 b (bottom): Average size at age over time for three representative ages of chilipepper rockfish (trawl fishery only).



Figure 3a (top) and 3b (bottom): Female and male weight/length relationship.


Figure 4: Maturity curve for chilipepper rockfish


Figure 5: Observed and predicted natural mortality rates ( $1 / \mathrm{M}$ ) based on age at $50 \%$ maturity for West Coast groundfish and Gulf of Alaska rockfish.


Figure 6: Observed and predicted natural mortality rates $(1 / M)$ based on age at $50 \%$ maturity for West Coast groundfish and Gulf of Alaska rockfish.


Figures 7a (top) and 7b (bottom): Total California rockfish landings by CDF\&G region, 1928-2002.


Figure 8 a and 8 b : Total estimated commercial chilipepper rockfish landings by CDF\&G region, 1928-2002.


Figure 9: Comparison of base (reconstructed \#1) versus an alternative (reconstructed \#2) catch history for the period between 1953 and 1977.


Figures 10a-10e: Records of the fraction of landings by gear type from 19301978 reported by district.




Figure 11a-11d: Comparison of catch estimates from Ralston 1998 with catch estimates used in this model.


Figure 12: Percentage of total rockfish catch (in 1000s) estimated to be chilipepper by RecFIN (modes CPFV and private only) and from CPFV observer data.


Figure 13: Total estimated recreational rockfish catches in northern and southern California as reported by RecFIN and CPFV logbook data, with reconstructed catches (in numbers) to 1928.


Figure 14: Estimated historical recreational catches of chilipepper rockfish in northern and southern California (tons) based on RecFIN data (1980-2006) and reconstructions based on historical sampling and CPFV logbook data (1928-1979).


Figure 15: Total estimated chilipepper rockfish landings by fishery, 1880-2006.


Figure 16: Trawl logbook CPUE time series developed in the last assessment by Ralston et al. (1998) and Ralston (1999).

Females


Age

Males


Age

Figure 17: Age composition data from trawl fisheries, 1978-2005

## Females



Age

Males


Age

Figure 18: Age composition data from hook and line fisheries, 1985-2002 with no data for some years


Figure 19: Age composition data from net fisheries, 1983-1998.

Females
Males
length bin

length bin

Figure 20: Length composition data from trawl fisheries, 1978-2006

Females


Figure 21: Length composition data from hook and line fisheries, 1980-2006 (with many years with no data)

Females
length bin


Figure 22: Length composition data from net fisheries, 1983-1998



Figure 23: Length composition data for Southern and Northern California from RecFIN database

## California



Figure 24: Coastwide length composition data from RecFIN database


Figure 25: Species coefficients for CDFG observer data using the Stephens/MacCall method.




Figure 26 (top): CPUE time series from the CDF\&G recreational observer data, with error estimated with a jackknife routine. Figure 27 (center) is block effects for the Rec CPUE model, Figure 28 (bottom) shows the depth bin effects.


Figure 29: Length frequency information (sex unknown) for the CDF\&G observer program recreational CPUE time series.


Figure 30: Chilipepper CPUE from triennial trawl survey across latitude and depth, 1977-2004; orange dots represent hauls in which no chilipepper were caught.


Figure 31 (top): Triennial survey core area-swept biomass index with estimated CV, and 32 (bottom) GLMM biomass point estimates with standard error.


Figure 33: Size composition of chilipepper rockfish from the triennial trawl survey.


Figure 34: Shift in size composition of chilipepper rockfish by depth (from raw triennial trawl survey catches, all years).


Figure 35: Chilipepper CPUE from NWFSC Combined survey, 2003-2006; orange dots reflect hauls in which no chilipepper were encoutnered.


Figure 36: NWFSC Combined survey abundance indices for Chilipepper rockfish


Figure 37: NWC Combined survey length compositions.


Figure 38: Juvenile (age 0) indices for core area (1984-2004) and coastwide (2001-2006) juvenile rockfish surveys


Figure 39 (top): Catches of chilipepper rockfish larvae from CalCOFI surveys, 1951-1969. Figure 40 (bottom), zones for estimating distance from shore in 25 km bins.


Figures 41a-c: Latitude (top), month (middle), and distance from shore (bottom) effects for the CalCOFI larval abundance index.


Figure 43: CalCOFI index point estimates, with error estimated from a jackknife. As two positive tows are necessary to run the jackknife, many years with a single positive tow $(1984,1985,1991,2000)$ are not included.

~95\% Asymptotic confidence interval


Figure 44-45: Base model output estimates of total biomass (top) and of spawning biomass with $\sim 95 \%$ asymptotic confidence intervals (bottom).


Figure 46-47: Base model output estimates of relative depletion (top) and projections of estimated depletion through 2018 with $\sim 95 \%$ asymptotic confidence intervals (bottom).

~95\% Asymptotic confidence interval


Figure 48-49: Model estimate recruitments (top) and observed recruitments with $\sim 95 \%$ asymptotic confidence intervals (bottom).



Figure 50-51: Model estimated recruitment deviation parameters (top) and recruitment deviance variance check (bottom).


Figure 52-53: Harvest rates for each of the four fisheries (top) and model estimated spawner recruit relationship (bottom).


Figure 54-55: Base model output estimates of Spawning potential ratio (SPR) relative to the $50 \%$ level (top) and phase plot of the same information relative to SPR and SSB targets (bottom).

Female ending year selectivity for fleet 1


Female ending year selectivity for fleet 2


Figure 56-57: Selectivity curves (double-normal form) for trawl (top) and hook and line (bottom) fisheries.

Female ending year selectivity for fleet 3


Figure 58-59: Selectivity curves (double-normal form) for setnet (top) and recreational (bottom) fisheries.


Female ending year selectivity for fleet 6


Figure 60-61: Selectivity curves (logistic form) for triennial bottom trawl survey (top) and NWC combined survey (bottom).


Female ending year selectivity for fleet 10


Figure 62-63: Selectivity curves (logistic form) for triennial bottom trawl survey (top) and NWC combined survey (bottom).

Expected numbers of females at age in thousands (max=76987.6)


Expected numbers of males at age in thousands (max=76987.6)


Figure 64-65: Model estimated numbers at age over time for females (top) and males (bottom).


Figure 66-67: Mean age of females (top) and males (bottom) in the population over time.



Figure 68-69: Fits to the trawl CPUE time series



Figure 70-71: Fits to the trawl CPUE time series in log space



Figure 72-73: Fits to the triennial survey core area swept index.



Figure 74-75: Fits to the triennial survey core area swept index in log space.


Index fleet 6


Figure 76-77: Fits to the NWC Combined survey.



Figure 78-79: Fits to the NWC Combined survey in log space.



Figure 80-81: Fits to the Coastwide juvenile survey.



Figure 82-83: Fits to the Coastwide juvenile survey in log space.


Index fleet 10


Figure 84-85: Fits to the Recreational CPUE index.



Figure 86-87: Fits to the recreational CPUE index in log space.


Male time-varying growth


Figures 88-89: Size at age contours for female (top) and male (bottom) chilipepper rockfish over time under time-varying growth assumptions.


Figure 90: Estimates of time-varying growth coefficient (K), with mean annual winter PDO and a running three year mean of the winter PDO.

Female whole catch length fits for fleet 1


Figure 91: Observed and predicted catch at length for female chilipepper in the trawl fishery.

Male whole catch length fits for fleet 1


Figure 92: Observed and predicted catch at length for male chilipepper in the trawl fishery.


Male whole catch Pearson residuals for fleet 1 (max=9.83)


Figure 93-94: Residuals to the length composition data in the trawl fishery


Sample size for male whole catch lengths for fleet 1


Figure 95-96: Observed and effective sample sizes for length composition data from the bottom trawl fishery.

Female whole catch length fits for fleet 2


Figure 97-98: Observed and predicted length composition data for females in the hook and line fishery.

Male whole catch length fits for fleet 2


Figure 99: Observed and predicted length composition data for males in the hook and line fishery.


Sample size for male whole catch lengths for fleet 2


Figure 100-101: Observed and effective sample sizes for length composition data from the hook and line fishery.


Male whole catch Pearson residuals for fleet 2 (max=9.18)


Figure 102-103: Residuals to the length composition data in the hook and line fishery

Female whole catch length fits for fleet 3


Male whole catch length fits for fleet 3


Figure 104-105: Observed and predicted length composition data for females (top) and males (bottom) in the setnet fishery.


Sample size for male whole catch lengths for fleet 3


Figure 106-107: Observed and effective sample sizes for length composition data from the setnet fishery.


Male whole catch Pearson residuals for fleet 3 (max=8.91)


Figure 108-109: Residuals to the length composition data in the setnet fishery

Combined sex whole catch length fits for fleet 4


Figure 110-111: Observed and predicted length composition data for combined sexes in the recreational fishery.


Combined sex whole catch Pearson residuals for fleet 4 (max=7.52)


Figure 112-113: Residuals (top) to the length composition data in the recreational fishery and (bottom) observed and effective sample sizes.

Female whole catch length fits for fleet 5


Male whole catch length fits for fleet 5


Figure 114-115: Observed and predicted length composition data for females (top) and males (bottom) in the triennial trawl survey.


Sample size for male whole catch lengths for fleet 5


Figure 116-117: Observed and effective sample sizes for length composition data from the triennial trawl survey.


Male whole catch Pearson residuals for fleet 5 (max=7.79)


Figure 118-119: Residuals to the length composition data in the triennial trawl survey.

Female whole catch length fits for fleet 6


Male whole catch length fits for fleet 6


Figure 120-121: Observed and predicted length composition data for females (top) and males (bottom) in the NWC combined survey.

Sample size for female whole catch lengths for fleet 6


Sample size for male whole catch lengths for fleet 6


Figure 122-123: Observed and effective sample sizes for length composition data from the NWC combined survey.


Male whole catch Pearson residuals for fleet 6 (max=3.84)


Figure 124-125: Residuals to the length composition data in the NWC combined survey

Combined sex whole catch length fits for fleet 10


Figure 126: Observed and predicted length composition data for mixed sexes in the recreational observer data associated with the CPUE index.


Combined sex whole catch Pearson residuals for fleet 10 (max=2.74)


Figure 127 (top): Observed and effective sample sizes for length composition data from the recreational CPUE index, and Figure 128 (bottom): residuals to the length composition data in the recreational CPUE index.

Female whole catch age fits for fleet 1


Figure 129: Observed and predicted catch at age data for females in the bottom trawl fishery.

Male whole catch age fits for fleet 1


Figure 130: Observed and predicted catch at age data for males in the bottom trawl fishery.



Figure 131-132: Observed and effective sample sizes for age composition data from the bottom trawl fishery.


Male whole catch Pearson residuals for age comps from fleet 1 (max=6.16)


Figure 133-134: Residuals to the age composition data in the bottom trawl fishery

Female whole catch age fits for fleet 2


Male whole catch age fits for fleet 2


Figure 135-136: Observed and predicted catch at age data for females (top) and males (bottom) in the hook and line fishery.


Sample size for male whole catch ages for fleet 2


Figure 137-138: Observed and effective sample sizes for age composition data from the hook and line fishery.
'emale whole catch Pearson residuals for age comps from fleet 2 (max=8.07


Male whole catch Pearson residuals for age comps from fleet 2 (max=7.25)


Figure 139-140: Residuals to the age composition data in the hook and line fishery

Female whole catch age fits for fleet 3


Male whole catch age fits for fleet 3


Figure 141-142: Observed and predicted catch at age data for females (top) and males (bottom) in the setnet fishery.


Sample size for male whole catch ages for fleet 3


Figure 143-144: Observed and effective sample sizes for age composition data from the setnet fishery.
'emale whole catch Pearson residuals for age comps from fleet 3 (max=11.8


Vale whole catch Pearson residuals for age comps from fleet 3 (max=19.28)


Figure 145-146: Residuals to the age composition data in the setnet fishery

Female whole catch age fits for fleet 6


Male whole catch age fits for fleet 6


Figure 147-148: Observed and predicted catch at age data for females (top) and males (bottom) for the year 2004 in the NWC Combined survey.

Female whole catch Pearson residuals for age comps from fleet 6 (max=1)


Male whole catch Pearson residuals for age comps from fleet 6 (max=1.34)


Figure 149-150: Residuals to the age composition data in the NWC combined survey

Observed and predicted female chilipepper mean size at age


Observed and predicted male chilipepper mean size at age


Figure 151-152: Observed (from commercial fisheries) and predicted (with time-varying k parameter) size at age for chilipepper rockfish females (top) and males (bottom).


Figure 153: Estimates of equilibrium recruitment (R0) plotted against likelihood values for twelve "jittered" base model runs.



Figure 154-155: Likelihood profiles for steepness (top) and female natural mortality in which the male offset is constant (bottom).


Figure 156: Likelihood surface plot for steepness against female natural mortality (in which the male offset is constant).


Figure 157-158: Estimated spawning biomass and recruitment trajectories when steepness is set to 0.34 (top, solid black lines) relative to the base model (grey, dashed lines) and when steepness is set to 0.81 (bottom).


Figure 159-160: Estimated spawning biomass and recruitment trajectories when female natural mortality is set to 0.12 (top, solid black lines) relative to the base model (grey, dashed lines) and when steepness is set to 0.20 (bottom).


Figure 161-162: Estimated SSB and recruitment trajectories when the trawl fishery CPUE time series is excluded (top) relative to the base model and when the triennial survey index and length frequency data are excluded (bottom).


Figure 163-164: Estimated SSB and recruitment trajectories when the NWC combined survey data are excluded (top) and when the coastwide juvenile survey index is excluded (including forecast, bottom).


Figure 165-166: Estimated spawning biomass and recruitment trajectories when the recreational CPUE data are excluded (top), and when all trawl fishery data (CPUE, length composition, age composition) are excluded.


Figure 167-168: Estimated spawning biomass and recruitment trajectories when hook and line age and length data are excluded (top, solid black lines) and when the setnet fishery data are excluded (bottom).


Figure 169-170: Estimated spawning biomass and recruitment trajectories with asymptotic selectivity estimated for the setnet fishery (top) and with timeinvariant growth (bottom).


Figure 171-172: Estimated spawning biomass and recruitment trajectories when historical (pre-1970) catches are doubles (top) or halved (bottom)


Figure 173-174: Comparison of the base model results with the results of the 1998 assessment for spawning biomass (top) and recruitment (bottom).


Figure 175-176: Comparison of the base model results with the results of the 2004 retrospective (top) and 2002 retrospective (bottom).


Figure 177-178: Comparison of the base model results with the results of the 1998 assessment for spawning biomass (top) and recruitment (bottom).

## Appendix A: Detailed history of regulations affecting the harvest of chilipepper rockfish

| Year | Period | Sector <br> (s) | Cum. Limit | Area(s) | RCA Configuration |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  | Jan. 1 - June 27 |  |  |  |  |
| 1983 | $\begin{gathered} \text { June } 28-\text { Sep. } 9 \\ \hline \text { Sep. } 10-\text { Dec. } \\ 31 \end{gathered}$ | All comm. | 40,000 lbs Sebastes/trip ${ }^{1}$ | Coastwide | NA |
| 1984 | Jan. 1 - Dec. 31 |  | 40,000 lbs Sebastes/trip ${ }^{2}$ | Eur., Mon., Concep. |  |
| 1985 | Jan. 1 - Dec. 31 | All comm. | 40,000 lbs Sebastes/trip | Eur., Mon., Concep. | NA |
| 1986 | Jan. 1 - Dec. 31 | All comm. | 40,000 lbs Sebastes/trip | Eur., Mon., Concep. | NA |
| 1987 | Jan. 1 - Dec. 31 | All comm. | 40,000 lbs Sebastes/trip | Eur., Mon., Concep. | NA |
| 1988 | Jan. 1 - Dec. 31 | All comm. | 40,000 lbs Sebastes/trip | Eur., Mon., Concep. | NA |
| 1989 | Jan. 1 - Dec. 31 | All comm. | 40,000 lbs Sebastes/trip | Eur., Mon., Concep. | NA |
| 1990 | Jan. 1 - Dec. 31 | All comm. | 40,000 lbs Sebastes/trip | Eur., Mon., Concep. | NA |
| 1991 | Jan. 1 - Dec. 31 | All comm. | 25,000 lbs Sebastes/trip of which no more than 5,000 lbs may be bocaccio | Eur., Mon., Concep. | NA |
| 1992 | Jan. 1 - Dec. 31 | All comm. | $50,000 \mathrm{lbs}$ Sebastes $/ 2$ weeks of which no more than $8,000 \mathrm{lbs}$ may be yellowtail (north of Cape Lookout, OR), no more than 10,000 lbs may be bocaccio (south of Cape Mendocino at $40^{\circ} 30^{\prime}$ N lat.) ${ }^{3}$ | Coastwide | NA |
| 1993 | Jan. 1 - Dec. 31 | All comm. | $50,000 \mathrm{lbs}$ Sebastes/2 weeks of which no more than $8,000 \mathrm{lbs}$ may be yellowtail (north of Coos Bay, OR), no more than 10,000 lbs may be bocaccio (south of Cape Mendocino at $40^{\circ} 30$ ' $N$ lat. $)^{4}$ | Coastwide | NA |
| 1994 | Jan. 1 - Dec. 31 | All comm. ${ }^{5}$ | $80,000 \mathrm{lbs}$ Sebastes/month of which no more than $14,000 \mathrm{lbs}$ may be yellowtail (north of Cape Lookout, OR), no more than 30,000 lbs may be yellowtail (south of Cape Lookout, OR), no more than 30,000 lbs may be bocaccio (south of Cape Mendocino at $40^{\circ} 30^{\prime} \mathrm{N}$ lat.) | Coastwide | NA |
|  | May 1 - Dec. 31 | Setnet | 40,000 lbs Sebastes/month | Off California |  |
|  | Sept. 1 - Dec. 31 | LE | 100,000 lbs Sebastes/month | South of Cape Mendocino at $40^{\circ} 30^{\prime}$ N lat. |  |
| 1995 |  |  | 50,000 lbs Sebastes/month of which no more than 30,000 lbs may be yellowtail and no more than $6,000 \mathrm{lbs}$ may be canary (coastwide) | Cape Lookout, OR Cape Mendocino at $40^{\circ} 30^{\prime} \mathrm{N}$ lat. | NA |
|  |  |  | $100,000 \mathrm{lbs}$ Sebastes/month of which no more than 30,000 lbs may be bocaccio and no more than $6,000 \mathrm{lbs}$ may be canary (coastwide) | South of Cape Mendocino at $40^{\circ} 30^{\prime}$ N lat. |  |
|  |  | OA | 35,000 lbs Sebastes/month | North of Cape Lookout, OR |  |
|  |  |  | 40,000 lbs Sebastes/month | South of Cape Lookout, OR |  |
|  |  | OA: hook-and-line and pot gears only | 10,000 lbs Sebastes/trip | Coastwide |  |


| 1996 | Jan. 1 - Oct. 31 |  | $100,000 \mathrm{lbs}$ Sebastes/2 months of which no more than 70,000 lbs may be yellowtail and no more than $18,000 \mathrm{lbs}$ may be canary (coastwide) | Cape Lookout, OR Cape Mendocino at $40^{\circ} 30^{\prime} \mathrm{N}$ lat. | NA |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  | Jan. 1 - Dec. 31 |  | 200,000 lbs Sebastes/2 months of which no more than 60,000 lbs may be bocaccio and no more than 18,000 lbs may be canary (coastwide) | South of Cape Mendocino at $40^{\circ} 30^{\prime}$ N lat. |  |
|  | Nov. 1 - Dec. 31 |  | $50,000 \mathrm{lbs}$ Sebastes/month of which no more than $35,000 \mathrm{lbs}$ may be yellowtail and no more than 9,000 lbs may be canary (coastwide) | Cape Lookout, OR Cape Mendocino at $40^{\circ} 30^{\prime} \mathrm{N}$ lat. |  |
|  |  | OA | 40,000 lbs Sebastes/month | South of Cape Lookout, OR |  |
|  |  | OA: hook-and-line and pot gears only | 10,000 lbs Sebastes/trip | Coastwide |  |
| 1997 | Jan. 1 - Apr. 30 |  | 150,000 lbs Sebastes/2 months of which no more than 12,000 lbs may be bocaccio and no more than 14,000 lbs may be canary (coastwide) | South of Cape Mendocino at 40오́ N lat. | NA |
|  | May 1 - Sept. 30 |  | 150,000 lbs Sebastes/2 months of which no more than 10,000 lbs may be bocaccio and no more than 14,000 lbs may be canary (coastwide) |  |  |
|  | Oct. 1 - Dec. 31 |  | $75,000 \mathrm{lbs}$ Sebastes/month of which no more than $5,000 \mathrm{lbs}$ may be bocaccio and no more than 10,000 lbs may be canary (coastwide) |  |  |
|  |  | OA ${ }^{6}$ | 40,000 lbs Sebastes/month | Coastwide |  |
|  | Jan. 1 - Dec. 31 | OA: hook-and-line and pot gears only ${ }^{7}$ | 10,000 lbs Sebastes/trip | Coastwide |  |
| 1998 | Jan. 1 - June 30 |  | 150,000 lbs Sebastes/ 2 months of which no more than 2,000 lbs may be bocaccio and no more than 15,000 lbs may be canary (coastwide) | South of Cape Mendocino at $40^{\circ} 30^{\prime}$ N lat. | NA |
|  | July 1 - Aug. 31 |  | 40,000 lbs Sebastes/2 months of which no more than 10,000 lbs may be bocaccio and no more than 14,000 lbs may be canary (coastwide) |  |  |
|  | Sept. 1-30 |  | $40,000 \mathrm{lbs}$ Sebastes/month of which no more than 10,000 lbs may be bocaccio and no more than 14,000 lbs may be canary (coastwide) |  |  |
|  | Oct. 1 - Dec. 31 |  | $15,000 \mathrm{lbs}$ Sebastes/month of which no more than $10,000 \mathrm{lbs}$ may be bocaccio and no more than 500 lbs may be canary (coastwide) |  |  |
|  | Jan. 1 - Dec. 31 | OA ${ }^{8}$ | 40,000 lbs Sebastes/month | Coastwide |  |
|  | Oct. 1 - Dec. 31 |  | Canary closed |  |  |
|  | Jan. 1 - Dec. 31 | OA: hook-and-line and pot gears only ${ }^{9}$ | 10,000 lbs Sebastes/trip |  |  |


| 1999 | Jan. 1 - March <br> 31 (phase 1) | LE ${ }^{10}$ | 45,000 lbs chilipepper/3 months | South of Cape Mendocino at $40^{\circ} 30^{\prime}$ N lat. | NA |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  | Apr. 1 - Sept. 30 (phase 2) |  | 25,000 lbs chilipepper/2 months |  |  |
|  | Oct. 1 - Dec. 31 (phase 3) |  | 5,000 lbs chilipepper/month |  |  |
|  | Jan. 1 - Dec. 31 | OA | 6,000 lbs chilipepper/month |  |  |
| 2000 | Jan. 1 - Dec. 31 | LE Trawl | MW trawls: 25,000 lbs chilipepper/2 months; Sm. FR trawls: 7,500 lbs chilipepper/2 months | South of Cape Mendocino at 40ㅇ́́ N lat. | NA |
|  | Jan. 1 - Feb. 29 | LE FG | 2,000 lbs chilipepper/month | $36^{\circ}-40^{\circ} 10^{\prime} \mathrm{N}$ lat. |  |
|  | Mar. 1 - Apr. 30 |  | Closed |  |  |
|  | May 1 - Dec. 31 |  | 2,000 lbs chilipepper/month |  |  |
|  | Jan. 1 - Feb. 29 |  | Closed | South of $36^{\circ} \mathrm{N}$ lat. |  |
|  | Mar. 1 - Dec. 31 |  | 2,000 lbs chilipepper/month |  |  |
|  | Jan. 1 - Feb. 29 | OA | 2,000 lbs chilipepper/month | $36^{\circ}-40^{\circ} 10^{\prime} \mathrm{N}$ lat. |  |
|  | Mar. 1 - Apr. 30 |  | Closed |  |  |
|  | May 1 - Dec. 31 |  | 2,000 lbs chilipepper/month |  |  |
|  | Jan. 1 - Feb. 29 |  | Closed | South of $36^{\circ} \mathrm{N}$ lat. |  |
|  | Mar. 1 - Dec. 31 |  | 2,000 lbs chilipepper/month |  |  |
| 2001 | Jan. 1 - Oct. 31 | LE Trawl | MW trawls: 25,000 Ibs chilipepper/2 months; trawls: 7,500 lbs chilipepper/2 months | South of Cape Mendocino at $40^{\circ} 10^{\prime}$ N lat. | Cowcod Conservation Areas implemented. |
|  | Nov. 1 - Dec. 31 |  | MW trawls: 25,000 lbs chilipepper/2 months; Sm. FR trawls: 5,000 lbs chilipepper/2 months |  |  |
|  | Jan. 1 - Feb. 29 | LE FG | 2,000 lbs chilipepper/month | $36^{\circ}-40^{\circ} 10^{\prime} \mathrm{N}$ lat. |  |
|  | Mar. 1 - Apr. 30 |  | Closed |  |  |
|  | May 1 - Dec. 31 |  | 2,000 lbs chilipepper/month |  |  |
|  | Jan. 1 - Feb. 29 |  | Closed | South of $36^{\circ} \mathrm{N}$ lat. |  |
|  | Mar. 1 - Dec. 31 |  | 2,000 lbs chilipepper/month |  |  |
|  | Jan. 1 - Feb. 29 | OA | 2,000 lbs chilipepper/month | $36^{\circ}-40^{\circ} 10^{\prime} \mathrm{N}$ lat. |  |
|  | Mar. 1 - Apr. 30 |  | Closed |  |  |
|  | May 1 - Dec. 31 |  | 2,000 lbs chilipepper/month |  |  |
|  | Jan. 1 - Feb. 29 |  | Closed | South of $36^{\circ} \mathrm{N}$ lat. |  |
|  | Mar. 1 - Dec. 31 |  | 2,000 lbs chilipepper/month |  |  |
| 2002 |  | LE Trawl |  | South of Cape Mendocino at $40^{\circ} 10^{\prime}$ N lat. | NA |
|  | Jan. 1- Apr. 30 |  | MW trawls: 25,000 lbs chilipepper/2 months; <br> Sm. FR trawls: 7,500 lbs chilipepper/2 months; Lg. FR trawls: 500 lbs chilipepper/trip not to exceed the sm. FR cumulative limit |  |  |
|  | May 1 - June 30 |  | MW trawls: 25,000 lbs chilipepper/2 months; <br> Sm. FR trawls: 4,000 lbs chilipepper/2 months; Lg. FR trawls: 500 lbs chilipepper/trip not to exceed the sm. FR cumulative limit |  |  |
|  | July 1 - Dec. 31 |  | Closed |  |  |


|  | Jan. 1 - Feb. 28 | LE FG | 500 lbs chilipepper/month | $34^{\circ} 27^{\prime}-40^{\circ} 10^{\prime} \mathrm{N}$ lat. |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  | Mar. 1 - Dec. 31 |  | Closed |  |  |
|  | Jan. 1 - Feb. 28 |  | Closed | South of $34^{\circ} 27^{\prime} \mathrm{N}$ lat. |  |
|  | Mar. 1 - June 30 |  | 2,000 lbs chilipepper/month |  |  |
|  | July 1 - Dec. 31 |  | Closed |  |  |
|  | Jan. 1 - Feb. 28 | OA | 500 lbs chilipepper/month | $34^{\circ} 27^{\prime}-40^{\circ} 10^{\prime} \mathrm{N}$ lat. |  |
|  | Mar. 1 - Dec. 31 |  | Closed |  |  |
|  | Jan. 1 - Feb. 28 |  | Closed | South of $34^{\circ} 27^{\prime} \mathrm{N}$ lat. |  |
|  | Mar. 1 - June 30 |  | 2,000 lbs chilipepper/month |  |  |
|  | July 1 - Dec. 31 |  | Closed |  |  |
| 2003 | Jan. 1 - Feb. 28 | LE Trawl | MW and sm. FR trawls: 300 lbs of minor shelf rockfish, widow, and chilipepper/month | $38^{\circ}-40^{\circ} 10^{\prime} \mathrm{N}$ lat. | 50-250 fm w/ petrale areas |
|  | Mar. 1 - Apr. 30 |  |  |  | 60-250 fm |
|  | May 1 - Oct. 31 |  |  |  | 60-200 fm |
|  | Nov. 1 - Dec. 31 |  |  |  | shoreline - $200 \mathrm{fm} \mathrm{w} /$ petrale areas |
|  | Jan. 1 - Feb. 28 |  |  | $34^{\circ} 27^{\prime}-38^{\circ} \mathrm{N}$ lat. | 50-150 fm |
|  | Mar. 1 - Apr. 30 |  |  |  | 60-150 fm |
|  | May 1 - Oct. 31 |  |  |  | 60-200 fm |
|  | Nov. 1 - Dec. 31 |  |  |  | shoreline - 200 fm w/ petrale areas |
|  | Jan. 1 - Apr. 30 |  |  | South of $34^{\circ} 27^{\prime} \mathrm{N}$ lat. | 100-150 fm along mainland coast; shoreline - 150 fm around islands |
|  | May 1 - Oct. 31 |  |  |  | 100-200 fm along mainland coast; shoreline - 200 fm around islands |
|  | Nov. 1 - Dec. 31 |  |  |  | shoreline - 200 fm along mainland coast and around islands w/ petrale areas |
|  | Jan. 1 - Feb. 28 | LE FG and OA | 100 Ibs minor shelf rockfish, widow, chilipepper, and yellowtai//2 months | $34^{\circ} 27^{\prime}-40^{\circ} 10^{\prime} \mathrm{N}$ lat | 20-150 fm |
|  | Mar. 1 - Apr. 30 |  | Closed |  |  |
|  | May 1 - June 30 |  | 200 Ibs minor shelf rockfish, widow, chilipepper, and yellowtail/2 months |  |  |
|  | July 1 - Aug. 31 |  | 250 lbs minor shelf rockfish, widow, chilipepper, and yellowtail/2 months |  |  |
|  | Sept. 1 - Oct. 31 |  | 200 lbs minor shelf rockfish, widow, chilipepper, and yellowtail/2 months |  |  |
|  | Nov. 1 - Dec. 31 |  | 100 lbs minor shelf rockfish, widow, chilipepper, and yellowtail/2 months |  | shoreline - 150 fm |
|  | Jan. 1 - Feb. 28 |  | 100 lbs minor shelf rockfish, widow, chilipepper, and yellowtail/2 months | South of $34^{\circ} 27^{\prime} \mathrm{N}$ lat. | 20-150 fm along mainland coast and around islands |
|  | Mar. 1 - Apr. 30 |  | Closed |  |  |
|  | May 1 - June 30 |  | 200 Ibs minor shelf rockfish, widow, chilipepper, and yellowtail/2 months |  |  |


|  | July 1 - Aug. 31 |  | 250 lbs minor shelf rockfish, widow, chilipepper, and yellowtail/2 months |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  | Sept. 1 - Oct. 31 |  | 200 lbs minor shelf rockfish, widow, chilipepper, and yellowtail/2 months |  | 30-150 fm along mainland coast and around islands |
|  | Nov. 1 - Dec. 31 |  | 100 lbs minor shelf rockfish, widow, chilipepper, and yellowtail/2 months |  | shoreline - 150 fm along mainland coast and around islands |
| 2004 | Jan. 1 - Apr. 30 | LE Trawl | MW and Ig. FR trawls: 2,000 lbs of chilipepper/2 months; sm FR trawls: 300 lbs of minor shelf rockfish, widow, and chilipepper/month | $38^{\circ}-40^{\circ} 10^{\prime} \mathrm{N}$ lat. | 75-150 fm; shoreline - 10 fm around Farallon Islands |
|  | May 1 - Aug. 31 |  | MW and Ig. FR trawls: 12,000 lbs of chilipepper/2 months; sm <br> FR trawls: 300 lbs of minor shelf rockfish, widow, and chilipepper/month through June 30, then $1,000 \mathrm{lbs}$ of minor shelf rockfish, widow and chilipepper/month no more than 200 lbs of which may be minor shelf and widow rockfish through Sept. 30 |  | 100-150 fm; shoreline - 10 fm around Farallon Islands |
|  | Sept. 1-30 |  | MW and Ig. FR trawls: 8,000 lbs of chilipepper/2 months through Dec. 31; sm FR trawls: 1,000 lbs of minor shelf rockfish, widow and chilipepper/month no more than 200 lbs of which may be minor shelf and widow rockfish through Sept. 30 |  | 75-150 fm; shoreline - 10 fm around Farallon Islands |
|  | Oct. 1 - Dec. 31 |  | MW, Ig. FR, and sm. FR trawls: 8,000 Ibs of chilipepper/2 months |  | shoreline - 250 fm |
|  | Jan. 1 - Apr. 30 |  | MW and Ig. FR trawls: 2,000 lbs of chilipepper/2 months; sm FR trawls: 300 lbs of minor shelf rockfish, widow, and chilipepper/month | $36^{\circ}-38^{\circ} \mathrm{N}$ lat. | 75-150 fm; shoreline - 10 fm around Farallon Islands |
|  | May 1 - Aug. 31 |  | MW and Ig. FR trawls: 12,000 lbs of chilipepper/2 months; sm <br> FR trawls: 300 lbs of minor shelf rockfish, widow, and chilipepper/month through June 30, then $1,000 \mathrm{lbs}$ of minor shelf rockfish, widow and chilipepper/month no more than 200 lbs of which may be minor shelf and widow rockfish through Sept. 30 |  | 100-150 fm; shoreline - 10 fm around Farallon Islands |
|  | Sept. 1-30 |  | MW and Ig. FR trawls: 8,000 Ibs of chilipepper/2 months through Dec. 31; sm FR trawls: 1,000 lbs of minor shelf rockfish, widow and chilipepper/month no more than 200 lbs of which may be minor shelf and widow rockfish through Sept. $30$ |  | $75-150 \mathrm{fm}$ shoreline - 10 fm around Farallon Islands |
|  | Oct. 1 - Dec. 31 |  | MW, Ig. FR, and sm. FR trawls: 8,000 Ibs of chilipepper/2 months |  | shoreline - 200 fm |
|  | Jan. 1 - Apr. 30 |  | MW and Ig. FR trawls: 2,000 lbs of chilipepper/2 months; sm FR trawls: 300 lbs of minor shelf rockfish, widow, and chilipepper/month | $34^{\circ} 27^{\prime}-36^{\circ} \mathrm{N}$ lat. | 75-150 fm |
|  | May 1 - Aug. 31 |  | MW and Ig. FR trawls: 12,000 Ibs of chilipepper/2 months; sm <br> FR trawls: 300 lbs of minor shelf rockfish, widow, and chilipepper/month through June 30, then $1,000 \mathrm{lbs}$ of minor shelf rockfish, widow and chilipepper/month no more than 200 lbs of which may be minor shelf and widow rockfish through Sept. 30 |  | 100-150 fm |
|  | Sept. 1-30 |  | MW and Ig. FR trawls: 8,000 lbs of chilipepper/2 months through Dec. 31; sm FR trawls: 1,000 lbs of minor shelf rockfish, widow and chilipepper/month no more than 200 lbs of which may be minor shelf and widow rockfish through Sept. 30 |  | 75-150 fm |
|  | Oct. 1 - Dec. 31 |  | MW, Ig. FR, and sm. FR trawls: 8,000 Ibs of chilipepper/2 months |  | shoreline - 150 fm |
|  | Jan. 1 - Apr. 30 |  | MW and Ig. FR trawls: 2,000 lbs of chilipepper/2 months; sm FR trawls: 300 lbs of minor shelf rockfish, widow, and chilipepper/month | South of $34{ }^{\circ} 27{ }^{\prime} \mathrm{N}$ lat. | 75-150 fm along mainland coast; shoreline - 150 fm around islands |



| Sept. 1-30 |  | MW and Ig. FR trawls: 8,000 lbs of chilipepper/2 months; sm FR trawls: 300 lbs of minor shelf rockfish, shortbelly, widow, yelloweye, and chilipepper/month |  |  |
| :---: | :---: | :---: | :---: | :---: |
| Oct. 1 - Dec. 31 |  |  |  | shoreline - 200 fm |
| Jan. 1 - Feb. 28 |  | MW and Ig. FR trawls: 2,000 lbs of chilipepper/2 months; sm FR trawls: 300 lbs of minor shelf rockfish, shortbelly, widow, yelloweye, and chilipepper/month | $34^{\circ} 27^{\prime}-36^{\circ} \mathrm{N}$ lat. | 75-150fm |
| Mar. 1 - Apr. 30 |  |  |  |  |
| May 1 - Aug. 31 |  | MW and Ig. FR trawls: 12,000 lbs of chilipepper/2 months; sm FR trawls: 300 lbs of minor shelf rockfish, shortbelly, widow, yelloweye, and chilipepper/month |  | 100-150 m |
| Sept. 1-30 |  | MW and Ig. FR trawls: 8,000 lbs of chilipepper/2 months; sm FR trawls: 300 lbs of minor shelf rockfish, shortbelly, widow, yelloweye, and chilipepper/month |  | 50-200 fm |
| Oct. 1 - Dec. 31 |  |  |  |  |
| Jan. 1 - Feb. 28 |  | MW and Ig. FR trawls: 2,000 lbs of chilipepper/2 months; sm FR trawls: 300 lbs of minor shelf rockfish, shortbelly, widow, yelloweye, and chilipepper/month | South of $34^{\circ} 27^{\prime} \mathrm{N}$ lat. | 75-150 fm along mainland coast; shoreline - 150 fm around islands |
| Mar. 1 - Apr. 30 |  |  |  |  |
| May 1 - Aug. 31 |  | MW and Ig. FR trawls: 12,000 lbs of chilipepper/2 months; sm FR trawls: 300 lbs of minor shelf rockfish, shortbelly, widow, yelloweye, and chilipepper/month |  | mainland coast; shoreline - 150 fm around islands |
| Sept. 1-30 |  | MW and Ig. FR trawls: 8,000 Ibs of chilipepper/2 months; sm FR trawls: 300 lbs of minor shelf rockfish, shortbelly, widow, yelloweye, and chilipepper/month |  |  |
| Oct. 1 - Dec. 31 |  |  |  | 50-200 fm along mainland coast; shoreline - 200 fm around islands |
| Jan. 1- Apr. 30 | LE FG | 2,000 lbs of chilipepper/2 months (opportunity only available seaward of the non-trawl RCA) | $34^{\circ} 27^{\prime}-40^{\circ} 10^{\prime} \mathrm{N}$ lat | 30-150 fm; shoreline - 10 fm around Farallon Islands |
| May 1 - Aug. 31 |  |  |  | 20-150 fm; shoreline - 10 fm around Farallon Islands |
| Sept. 1 - Dec. 31 |  |  |  | 30-150 fm; shoreline - 10 fm around Farallon Islands |
| Jan. 1 - Dec. 31 |  |  | South of $34^{\circ} 27^{\prime} \mathrm{N}$ lat. | 60-150 fm along mainland coast and around islands |
| Jan. 1 - Feb. 28 | OA | 300 lbs of minor shelf rockfish, shortbelly, widow, and chilipepper/2 months | $34^{\circ} 27^{\prime}-40^{\circ} 10^{\prime} \mathrm{N}$ lat | $30-150 \mathrm{fm} ;$ shoreline - 10 fm around Farallon Islands |
| Mar. 1 - Apr. 30 |  | Closed |  |  |
| May 1 - Aug. 31 |  | 300 lbs of minor shelf rockfish, shortbelly, widow, and chilipepper/2 months |  | 20-150 fm; shoreline - 10 fm around Farallon Islands |
| Sept. 1 - Dec. 31 |  | 300 lbs of minor shelf rockfish, shortbelly, widow, and chilipepper/2 months |  | $30-150 \mathrm{fm} ;$ shoreline - 10 fm around Farallon Islands |
| Jan. 1 - Feb. 28 |  | 500 lbs of minor shelf rockfish, shortbelly, widow, and chilipepper/2 months | South of $34^{\circ} 27^{\prime} \mathrm{N}$ lat. | 60-150 fm along mainland coast and |



| Mar. 1 - Apr. 30 |
| :---: | :---: | :---: | :---: | :---: |

(1) From Jan. 1 to June 27, Van. \& Col. Sebastes $\mathrm{HG}=14,000 \mathrm{mt}$, from June 28-Sept. 9, Van. \& Col. Sebastes $\mathrm{HG}=$ $18,500 \mathrm{mt}$, $1 \mathrm{trip} /$ week, from Sept. 10-Dec. 31, Van. \& Col. 3,000 lbs/trip, no weekly trip limit.
(2) From 1984-1991, no weekly trip limits
(3) Sebastes harvest guideline north of Cape Lookout, $\mathrm{OR}=8,000 \mathrm{mt}$; min. mesh size for trawl codends increased from 3 to 4.5 inches effective May 9, 1992.
(4) Sebastes harvest guideline north of Coos Bay, OR $=11,200 \mathrm{mt}$
(5) Groundfish fishery separated into limited entry and open access sectors w/ LE gear endorsements for trawl, longline, and pot/trap gears
(6) Setnets only legal south of $38^{\circ} \mathrm{N}$ lat.; setnets limited to $4,000 \mathrm{lbs}$ bocaccio/month.
(7) Limits include 300 lbs bocaccio/trip, not to exceed $2,000 \mathrm{lbs} /$ month south of Cape Mendocino (Jan. 1 - Apr. 30 ); 250 lbs bocaccio/trip not to exceed $2,000 \mathrm{lbs} /$ month south of Cape Mendocino (May 1 - Dec. 31).
(8) Setnets only legal south of $38^{\circ} \mathrm{N}$ lat.; setnets limited to $2,000 \mathrm{lbs}$ bocaccio/month.
(9) 250 lbs bocaccio/trip not to exceed $1,000 \mathrm{lbs} /$ month south of Cape Mendocino (Jan. 1 - Dec. 31).
(10) First year of limits specifically for chilipepper rockfish. For limited entry fishery, a new three-phase cumulative limit period system is introduced: phase 1 is a single 3-month cum. limit period from Jan. 1 - March 31, phase 2 has three separate 2-month cum. limit periods (Apr. 1 - May 31, June 1 - July 31, and Aug. 1 - Sept. 30, and phase 3 has three separate 1-month cum. limit periods (Oct. 1-31, Nov. 1-30 and Dec. 1-31); only POP and bocaccio have monthly limits within a cum. limit period.
(11) Cumulative landing limit periods redefined to encompass six 2-month periods through the year (Jan-Feb, Mar-Apr, May-June, July-Aug, Sept-Oct, and Nov-Dec). Chilipepper rockfish required to be sorted south of $40^{\circ} 10^{\prime} \mathrm{N}$ lat. Small footrope trawls required to land chilipepper rockfish in the LE trawl sector.
(12) Small footrope trawls required to land chilipepper rockfish in the LE trawl sector.

## Appendix B: Data (.dat) and Control (.ctl) files for chilipepper rockfish model



| 32 | 615 | 0 | 0 | $\#$ | 1925 |
| :--- | :--- | :--- | :--- | :--- | :--- |
| 44 | 845 | 0 | 0 | $\#$ | 1926 |
| 38 | 716 | 0 | 0 | $\#$ | 1927 |
| 37.05 | 701.45 | 0 | 0 | $\#$ | 1928 |
| 33.28 | 626.11 | 0 | 0.02 | $\#$ | 1929 |
| 41.41 | 780.81 | 0 | 0.11 | $\#$ | 1930 |
| 41.63 | 788.44 | 0 | 0.26 | $\#$ | 1931 |
| 32.87 | 622.52 | 0 | 0.46 | $\#$ | 1932 |
| 28.42 | 539.33 | 0 | 0.72 | $\#$ | 1933 |
| 26.63 | 503.03 | 0 | 1.04 | $\#$ | 1934 |
| 28.68 | 541 | 0 | 1.41 | $\#$ | 1935 |
| 29.29 | 552.03 | 0 | 1.84 | $\#$ | 1936 |
| 26.9 | 508 | 0 | 2.46 | $\#$ | 1937 |
| 20.24 | 371.34 | 0 | 2.69 | $\#$ | 1938 |
| 16.69 | 298.89 | 0 | 2.59 | $\#$ | 1939 |
| 19.81 | 362.24 | 0 | 4.07 | $\#$ | 1940 |
| 92.13 | 267.63 | 0 | 0 | $\#$ | 1941 |
| 55.41 | 51.91 | 0 | 0 | $\#$ | 1942 |
| 122.97 | 32.15 | 0 | 0 | $\#$ | 1943 |
| 210.21 | 9.15 | 0 | 0 | $\#$ | 1944 |
| 417.86 | 16.31 | 0 | 0 | $\#$ | 1945 |
| 362.4 | 17.56 | 0 | 0 | $\#$ | 1946 |
| 321.63 | 21.59 | 0 | 3.42 | $\#$ | 1947 |
| 312.78 | 25.71 | 0 | 8.83 | $\#$ | 1948 |
| 324.8 | 28.86 | 0 | 14.79 | $\#$ | 1949 |
| 510.48 | 47.9 | 0 | 17.61 | $\#$ | 1950 |
| 777.91 | 74.8 | 0 | 16.79 | $\#$ | 1951 |
| 935.3 | 97.74 | 0 | 21.66 | $\#$ | 1952 |
| 1068.63 | 111.16 | 0 | 27.36 | $\#$ | 1953 |
| 1036.67 | 117.59 | 0 | 60.75 | $\#$ | 1954 |
| 1149.08 | 122.25 | 0 | 109.39 | $\#$ | 1955 |
| 1344.04 | 163.3 | 0 | 135.95 | $\#$ | 1956 |
| 1433.55 | 173.86 | 0 | 79.32 | $\#$ | 1957 |
| 1504.31 | 326.47 | 0 | 57.85 | $\#$ | 1958 |
| 1286.21 | 271.22 | 0 | 35.8 | $\#$ | 1959 |
| 1258.21 | 148.56 | 0 | 36.69 | $\#$ | 1960 |
| 956.33 | 146.41 | 0 | 42.99 | $\#$ | 1961 |
| 917.45 | 155.6 | 0 | 45.01 | $\#$ | 1962 |
| 917.46 | 111.18 | 0 | 48.64 | $\#$ | 1963 |
| 711 | 105.72 | 0 | 66.79 | $\#$ | 1964 |
| 765.36 | 136.09 | 0 | 91.87 | $\#$ | 1965 |
| 1904.92 | 140.17 | 0 | 137.25 | $\#$ | 1966 |
| 2497.6 | 127.21 | 0 | 171.21 | $\#$ | 1967 |
| 1468.36 | 112.75 | 0 | 193.89 | $\#$ | 1968 |
| 810.32 | 103.79 | 0 | 176.31 | $\#$ | 1969 |
| 907.76 | 114.21 | 0 | 250.66 | $\#$ | 1970 |
| 866.94 | 154.71 | 0 | 231.32 | $\#$ | 1971 |
| 1371.84 | 215.02 | 0 | 312.43 | $\#$ | 1972 |
| 2893.25 | 371.42 | 0 | 379.74 | $\#$ | 1973 |
| 3192.94 | 282.37 | 0 | 485.07 | $\#$ | 1974 |
| 2588.29 | 260.32 | 0 | 379.17 | $\#$ | 1975 |
| 2334.62 | 210.45 | 0 | 546.82 | $\#$ | 1976 |
| 1490.73 | 166.5 | 0 | 433.94 | $\#$ | 1977 |
| 1293.23 | 169.16 | 25.83 | 445.32 | $\#$ | 1978 |
| 2003.71 | 176.6 | 54.19 | 490.43 | $\#$ | 1979 |
| 2720.86 | 95.87 | 45.38 | 392.91 | $\#$ | 1980 |
|  |  |  |  |  |  |


| 2294.63 | 139.13 | 71.28 | 271.32 | $\#$ | 1981 |
| :--- | :--- | :--- | :--- | :--- | :--- |
| 1680.73 | 356.35 | 85.42 | 369.44 | $\#$ | 1982 |
| 1879.45 | 80.23 | 345.21 | 159.78 | $\#$ | 1983 |
| 2447.65 | 98.1 | 231.04 | 145.75 | $\#$ | 1984 |
| 1807.06 | 278.99 | 738.69 | 357.66 | $\#$ | 1985 |
| 1269.14 | 330.88 | 1161.46 | 385.97 | $\#$ | 1986 |
| 1313.85 | 172.61 | 461.11 | 111.75 | $\#$ | 1987 |
| 1777.91 | 333.47 | 289.36 | 290.01 | $\#$ | 1988 |
| 2363.3 | 425.58 | 361.37 | 245.15 | $\#$ | 1989 |
| 2317.2 | 232.12 | 372.77 | 188.11 | $\#$ | 1990 |
| 2229.02 | 618.32 | 332.08 | 131.08 | $\#$ | 1991 |
| 1329.79 | 1052.67 | 296.72 | 74.04 | $\#$ | 1992 |
| 1282.12 | 860.86 | 232.91 | 17 | $\#$ | 1993 |
| 1267.12 | 484.99 | 107.71 | 17.16 | $\#$ | 1994 |
| 1594.58 | 324.9 | 94.05 | 7.17 | $\#$ | 1995 |
| 1528.08 | 254.23 | 57.67 | 30.31 | $\#$ | 1996 |
| 1613.97 | 339.29 | 82.97 | 73.47 | $\#$ | 1997 |
| 1137.97 | 208.84 | 77.62 | 5.39 | $\#$ | 1998 |
| 838.61 | 104.18 | 9.67 | 24.29 | $\#$ | 1999 |
| 403.38 | 50.6 | 6.11 | 38.92 | $\#$ | 2000 |
| 435.57 | 25.18 | 4.9 | 51.74 | $\#$ | 2001 |
| 300.03 | 6.22 | 0.42 | 22.25 | $\#$ | 2002 data from 2002 onward include |
| 20.33 | 0.25 | 0.05 | 0 | $\#$ | 2003 WCGOP estimates of discard |
| 203.1 | 10.43 | 2.86 | 19.43 | $\#$ | 2004 |
| 171.97 | 9.77 | 0.14 | 10.17 | $\#$ | 2005 |
| 104.74 | 17.62 | 0.45 | 3.85 | $\#$ | 2006 |

\# Abundance indices

| 94 | \# number of observations |  |  |  |
| :--- | :--- | :--- | :--- | :--- |
| \#year | season | type | value | SD |
| 1980 | 1 | 1 | 249 | 0.25 |
| 1981 | 1 | 1 | 150 | 0.25 |
| 1982 | 1 | 1 | 121 | 0.25 |
| 1983 | 1 | 1 | 116 | 0.25 |
| 1984 | 1 | 1 | 91 | 0.25 |
| 1985 | 1 | 1 | 88 | 0.25 |
| 1986 | 1 | 1 | 76 | 0.25 |
| 1987 | 1 | 1 | 116 | 0.25 |
| 1988 | 1 | 1 | 158 | 0.25 |
| 1989 | 1 | 1 | 172 | 0.25 |
| 1990 | 1 | 1 | 149 | 0.25 |
| 1991 | 1 | 1 | 146 | 0.25 |
| 1992 | 1 | 1 | 109 | 0.25 |
| 1993 | 1 | 1 | 80 | 0.25 |
| 1994 | 1 | 1 | 112 | 0.25 |
| 1995 | 1 | 1 | 126 | 0.25 |
| 1996 | 1 | 1 | 96 | 0.25 |

\#

| \# triennial GLM tuned |  |  |  |
| :---: | :---: | :---: | :---: |
| 1980 | 1 | 5 | 3954.371 .625 |
| 1983 | 1 | 5 | 1994.420 .613 |
| 1986 | 1 | 5 | 1166.331 .213 |
| 1989 | 1 | 5 | 2400.580 .300 |
| 1992 | 1 | 5 | 368.770 .581 |
| 1995 | 1 | 5 | 1545.100 .264 |
| 1998 | 1 | 5 | 945.460 .341 |


| 2001 | 1 | 5 | 806.63 | 0.285 |
| :--- | :--- | :--- | :--- | :--- |
| 2004 | 1 | 5 | 2157.54 | 0.254 |

\#NWC combo survey glm tuned

| 2003 | 1 | 6 | 3932 | 0.61654 |
| :--- | :--- | :--- | :--- | :--- |
| 2004 | 1 | 6 | 24559 | 1.19248 |
| 2005 | 1 | 6 | 9540 | 0.4466 |
| 2006 | 1 | 6 | 7384 | 0.40252 |


| \# juvenile survey- FED <br> \#year |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- |
| season | type | value | SD |  |
| 1984 | 1 | 7 | 7.3254 | 0.37012 |
| 1985 | 1 | 7 | 8.1232 | 0.4589 |
| 1986 | 1 | 7 | 0.7227 | 0.3300 |
| 1987 | 1 | 7 | 13.2204 | 0.3468 |
| 1988 | 1 | 7 | 16.3753 | 0.3859 |
| 1989 | 1 | 7 | 0.3869 | 0.4811 |
| 1990 | 1 | 7 | 0.3093 | 0.4094 |
| 1991 | 1 | 7 | 0.9761 | 0.3383 |
| 1992 | 1 | 7 | 0.1687 | 0.5192 |
| 1993 | 1 | 7 | 10.3256 | 0.2972 |
| 1994 | 1 | 7 | 0.0235 | 0.8093 |
| 1995 | 1 | 7 | 0.2455 | 0.6069 |
| 1996 | 1 | 7 | 0.0909 | 0.5163 |
| 1997 | 1 | 7 | 0.1310 | 0.7428 |
| 1999 | 1 | 7 | 0.2059 | 0.4342 |
| 2000 | 1 | 7 | 0.0888 | 0.5242 |
| 2001 | 1 | 7 | 0.8528 | 0.3412 |
| 2002 | 1 | 7 | 2.2921 | 0.3228 |
| 2003 | 1 | 7 | 1.0052 | 0.4103 |
| 2004 | 1 | 7 | 1.3333 | 0.3902 |
| $\#$ |  |  |  |  |
| 2001 | 1 | 8 | 1.7161 | 0.0401 |
| 2002 | 1 | 8 | 2.7629 | 0.0451 |
| 2003 | 1 | 8 | 1.5719 | 0.0367 |
| 2004 | 1 | 8 | 2.9379 | 0.0360 |
| 2005 | 1 | 8 | 0.8658 | 0.0346 |
| 2006 | 1 | 8 | 0.7523 | 0.0301 |

\#
\#
\# CalCOFI survey

| \#year | season | type | Index CV <br> \#year | season |
| :--- | :--- | :--- | :--- | :--- |
| type | Index CV |  |  |  |
| 1951 | 1 | 9 | 0.14183053 | 0.8414901 |
| 1953 | 1 | 9 | 0.16864622 | 0.4698166 |
| 1954 | 1 | 9 | 0.21885162 | 0.3547108 |
| 1955 | 1 | 9 | 0.2545118 | 0.4020231 |
| 1956 | 1 | 9 | 0.12075705 | 0.6590477 |
| 1957 | 1 | 9 | 0.30887709 | 0.522799 |
| 1958 | 1 | 9 | 0.39454343 | 0.3479359 |
| 1959 | 1 | 9 | 0.08842933 | 0.4466416 |
| 1960 | 1 | 9 | 0.18220879 | 0.3299083 |
| 1961 | 1 | 9 | 0.08775916 | 0.5532203 |
| 1962 | 1 | 9 | 0.068755 | 0.6127899 |
| 1963 | 1 | 9 | 0.19684699 | 0.4639924 |
| 1964 | 1 | 9 | 0.0631976 | 0.5157418 |
| 1965 | 1 | 9 | 0.14914866 | 0.3859004 |


| 1966 | 1 | 9 | 0.24731002 | 0.3842774 |
| :--- | :--- | :--- | :--- | :--- |
| 1967 | 1 | 9 | 0.34379234 | 0.540158 |
| 1968 | 1 | 9 | 0.63368278 | 0.5381044 |
| 1969 | 1 | 9 | 0.55183877 | 0.3579827 |
| 1970 | 1 | 9 | 0.27392882 | 0.5389176 |
| 1975 | 1 | 9 | 0.02550871 | 0.6909198 |
| 1992 | 1 | 9 | 0.12549796 | 0.5956311 |
| 2002 | 1 | 9 | 0.04308614 | 0.6761029 |
| 2003 | 1 | 9 | 0.08688551 | 0.4902213 |
| 2004 | 1 | 9 | 0.1717815 | 0.4136779 |
| 2005 | 1 | 9 | 0.01187012 | 0.7130089 |
| 2006 | 1 | 9 | 0.03316714 | 0.7720739 |
| $\#$ rec cpue |  |  |  |  |
| $\#$ year | season | type | index | jack.cv |
| 1987 | 1 | 10 | 0.166856206 | 0.1631351 |
| 1988 | 1 | 10 | 0.083010716 | 0.1794928 |
| 1989 | 1 | 10 | 0.054122438 | 0.1633441 |
| 1990 | 1 | 10 | 0.031462634 | 0.4267126 |
| 1991 | 1 | 10 | 0.040173333 | 0.3545357 |
| 1992 | 1 | 10 | 0.064866103 | 0.5545214 |
| 1993 | 1 | 10 | 0.026517113 | 0.2333201 |
| 1994 | 1 | 10 | 0.023850668 | 0.2796596 |
| 1995 | 1 | 10 | 0.024610012 | 0.4197283 |
| 1996 | 1 | 10 | 0.015093027 | 0.4449115 |
| 1997 | 1 | 10 | 0.008328447 | 0.3430329 |
| 1998 | 1 | 10 | 0.006612019 | 0.421573 |

\# Discard section- currently I have no discard data
2 \# Discard biomass (1=biomass, 2=fraction)
0 \# number of observations
\# mean body weight (in kg )
0 \# number of observations
\# length composition
-1 \# compress tails of composition (negative turns off)
0.0001 \# constant added to observed and expected proportions at age

19 \# number of length bins

| 16 | 18 | 20 | 22 | 24 | 26 | 28 | 30 | 32 | 34 | 36 | 38 | 40 | 42 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |

112 \# number of length observations-
\# length composition
\#
\# Trawl fishery Females first, then males females
males

| \#year | season | type | gender | partition \# samples | 16 | 18 | 20 | 22 | 24 | 26 | 28 |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
|  | 30 | 32 | 34 | 36 | 38 | 40 | 42 | 44 | 46 | 48 | 50 | 52 | 16 |
|  | 18 | 20 | 22 | 24 | 26 | 28 | 30 | 32 | 34 | 36 | 38 | 40 | 42 |
|  | 44 | 46 | 48 | 50 | 52 |  |  |  |  |  |  |  |  |
| 1978 | 1 | 1 | 3 | 0 | 147 | 0.00022 | 0 | 0 | 0.01818 | 0.00388 | 0.00229 | 0.00744 |  |
|  | 0.01194 | 0.04564 | 0.05786 | 0.04806 | 0.05182 | 0.07637 | 0.10655 | 0.05257 | 0.04429 | 0.07482 | 0.01717 | 0.01018 | 0 |
|  | 0 | 0 | 0.00021 | 0.00069 | 0.00102 | 0.01447 | 0.05906 | 0.18275 | 0.04776 | 0.04849 | 0.01021 | 0.00039 | 0 |
|  | 0.00018 | 0.00121 | 0 | 0.00429 | 0 |  |  |  |  |  |  |  |  |
| 1979 | 1 | 1 | 3 | 0 | 110 | 0 | 0 | 0.00049 | 0 | 0.00004 | 0.00132 | 0.02087 | 0.0092 |
|  | 0.01246 | 0.04269 | 0.03287 | 0.03745 | 0.1193 | 0.066 | 0.17126 | 0.10614 | 0.08089 | 0.00735 | 0.00528 | 0 | 0 |


|  | 0 0 | 0 | 0.00041 | 0.00095 | 0.00821 | 0.04017 | 0.0724 | 0.067510 .059740 .035850 .000110 .000010 .0008 |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 0 | 0.00008 | 0.00017 |  |  |  |  |  |  |  |  |  |  |
| 1980 | 1 | 1 | 3 | 0 | 191 | 0 | 0 | 0.00039 | 0 | 0 | 0.00349 | 0.00287 | 0.0041 |
|  | 0.02768 | 0.05072 | 0.06043 | 0.1232 | 0.09582 | 0.10987 | 0.08439 | 0.07823 | 0.03707 | 0.0149 | 0.00063 |  | 0 |
|  | 0 | 0 | 0.00342 | 0.00256 | 0.00799 | 0.03147 | 0.08474 | 0.09921 | 0.04584 | 0.01837 | 0.00273 | 0.00223 |  |
|  | 0.00025 | 0.00042 | 0.0066 | 0.00008 | 0.0003 |  |  |  |  |  |  |  |  |
| 1981 | 1 | 1 | 3 | 0 | 125 | 0 | 0 | 0 | 0 | 0 | 0.00088 | 0.00667 |  |
|  | 0.00529 | 0.01266 | 0.01064 | 0.09861 | 0.2005 | 0.09316 | 0.10213 | 0.0487 | 0.07159 | 0.04917 | 0.00273 | 0.00009 |  |
|  | 0 | 0 | 0 | 0 | 0.00064 | 0.00026 | . 0487 | 0.11222 | 0.12205 | 0.0119 | 0.00084 | 0.00005 |  |
|  | 0.00046 |  | 0.00002 | 0 | 0 | 0 |  |  |  |  |  |  |  |
| 1982 | 1 | 1 | 3 | 0 | 195 | 0 | 0 | 0 | 0.00035 | 0.00022 | 0.00067 | 0.00525 |  |
|  | 0.01354 | 0.01678 | 0.0125 | 0.06505 | 0.08043 | 0.13048 | 0.18373 | 0.15391 | 0.076 | 0.03757 | 0.01085 | 0.00174 |  |
|  | 0 | 0 | 0.00078 | 0.00005 | 0.00359 | 0.00727 | 0.02841 | 0.07633 | 0.06915 | 0.02099 | 0.00408 | 0.00023 |  |
|  | 0.00006 |  | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |
| 1983 | 1 | 1 | 3 | 0 | 275 | 0 | 0 | 0 | , | 0.0002 | 0.00113 | 0.00338 |  |
|  | 0.01176 | 0.01812 | 0.01728 | 0.02633 | 0.03683 | 0.13454 | 0.20614 | 0.14642 | 0.11552 | 0.07491 | 0.02504 | 0.00759 |  |
|  | 0 | 0 | 0.00004 | 0.0001 | 0.00066 | 0.00736 | 0.03449 | 0.03921 | 0.05539 | 0.02184 | 0.00391 | 0.00018 |  |
|  | 0.00244 | 0.00191 | 0.00005 | 0.00001 | 0.00007 | 0.00715 |  |  |  |  |  |  |  |
| 1984 | . | 1 | 3 | 0 | 305 | 0 | 0 | 0 | 0.00003 | 0.000 | 0.0036 | 0.00333 |  |
|  | 0.01501 | 0.05746 | 0.08824 | 0.16352 | 0.06524 | 0.10441 | 0.07823 | 0.06725 | 0.04769 | 0.02093 | 0.00477 | 0.0017 |  |
|  | 0.000020 |  | 0 | 0 | 0.00009 | 0.00102 | 0.02879 | 0.0387 | 0.0771 | 0.06447 | 0.05422 | 0.00792 |  |
|  | 0.00032 | 0.00166 | 0.00061 | 0.00242 | 0.00049 | 0.00052 | 0.00002 |  |  |  |  |  |  |
| 1985 | 1 | 1 | 3 | 0 | 338 | 0 | 0 | 0 | 0.001 | 0.00035 | 0.00128 | 0.00832 |  |
|  | 0.02207 | 0.04019 | 0.06271 | 0.08883 | 0.11605 | 0.06376 | 0.05989 | 0.07079 | 0.04972 | 0.02535 | 0.00534 | 0.00193 |  |
|  | 0 | 0 | 0.00009 | 0.00011 | 0.00232 | 0.01902 | 0.06599 | 0.10678 | 0.1175 | 0.04632 | 0.01314 | 0.00603 |  |
|  | 0.00042 | 0.00045 | 0.00138 | 0.0015 | 0.00138 | 0 |  |  |  |  |  |  |  |
| 1986 | 1 | 1 |  |  | 219 | 0.00044 | 0.0001 | , | 0.00022 | 0.00009 | 0.00458 | 0.00832 |  |
|  | 0.02425 | 0.0379 | 0.0594 | 0.07245 | 0.09209 | 0.07529 | 0.05696 | 0.07571 | 0.06683 | 0.03424 | 0.03705 | 0.00078 |  |
|  | 0.00004 |  | 0.00093 | 0.0034 | 0.00564 | 0.01592 | 0.09321 | 0.10176 | 0.06953 | 0.03448 | 0.01659 | 0.00662 |  |
|  | 0.00095 |  | 0.0018 | 0.00244 | 0 | 0 |  |  |  |  |  |  |  |
| 1987 | 1 | 1 | 3 | 0 | 211 | 0.00016 |  | 0.00012 | 0.00003 | 0.00 | 0.015 | 0.07235 |  |
|  | 0.16683 | 0.09549 | 0.04457 | 0.03733 | 0.04516 | 0.04761 | 0.04209 | 0.0179 | 0.00896 | 0.005 | 0.00057 | 0.00056 |  |
|  | 0 | 0 | 0 | 0.00112 | 0.04064 | 0.1188 | 0.06182 | 0.08213 | 0.06136 | 0.022 | 0.00782 | 0.00086 |  |
|  | 0.00019 | 0.00001 | 0.00001 | 0 | 0 | 0 |  |  |  |  |  |  |  |
| 1988 | 1 | 1 | 3 | 0 | 199 | 0 | 0 | 0 | 0 | 0.00003 | 0.01118 | 0.03265 |  |
|  | 0.08052 | 0.0893 | 0.10642 | 0.08444 | 0.01661 | 0.03359 | 0.05067 | 0.02813 | 0.01291 | 0.00676 | 0.00425 | 0.0009 | 0 |
|  | 0 | 0 | 0.00003 | 0.00014 | 0.04746 | 0.12885 | 0.10265 | 0.08427 | 0.0428 | 0.03387 | 0.00139 |  |  |
|  | 0.00016 | 0.00001 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |
| 1989 | 1 | 1 | 3 | 0 | 183 | 0.00007 | 0 | 0 | 0 | 0.00207 | 0.00491 | 0.0133 |  |
|  | 0.01524 | 0.05436 | 0.09059 | 0.13372 | 0.17294 | 0.02935 | 0.01437 | 0.01396 | 0.00704 | 0.00758 | 0.00131 |  | 0 |
|  | 0 | 0 | 0.00096 | 0.00612 | 0.00994 | 0.0414 | 0.15366 | 0.12776 | 0.06141 | 0.03496 | 0.00173 | 0.00017 |  |
|  | 0.00098 | 0 | 0.00009 | 0 | 0 | 0 |  |  |  |  |  |  |  |
| 1990 | 1 | 1 | 3 | 0 | 204 | 0.00001 |  | 0.00006 |  | 0.00355 | 0.00738 | 0.03629 |  |
|  | 0.04755 | 0.04567 | 0.04607 | 0.06876 | 0.14846 | 0.10491 | 0.043 | 0.03709 | 0.00822 | 0.00432 | 0.00119 | 0.00018 |  |
|  | 0 | 0 | 0 | 0.00195 | 0.02245 | 0.05403 | 0.08982 | 0.12547 | 0.04891 | 0.04953 | 0.004 | 0.00087 |  |
|  | 0.00021 |  | 0.00002 | 0.00005 |  |  |  |  |  |  |  |  |  |
| 1991 | 1 | 1 | 3 | 0 | 208 | 0.00017 |  | 0.0005 | 0.00091 | 0.00456 | 0.01515 | 0.02599 |  |
|  | 0.05384 | 0.08291 | 0.06996 | 0.06904 | 0.07213 | 0.07997 | 0.04056 | 0.03088 | 0.01192 | 0.0107 | 0.00363 | 0.00104 |  |
|  | 0 | 0.00015 | 0.00013 | 0.00662 | 0.01265 | 0.05956 | 0.10457 | 0.13979 | 0.06707 | 0.02766 | 0.00608 | 0.00157 |  |
|  | 0.00009 |  | 0.0002 | 0 | 0 |  |  |  |  |  |  |  |  |
| 1992 | 1 | 1 | 3 | 0 | 132 | 0 | 0 | 0 | 0.00005 | 0.00405 | 0.0288 | 0.05881 |  |
|  | 0.09328 | 0.08427 | 0.06824 | 0.04726 | 0.07089 | 0.06935 | 0.07266 | 0.04536 | 0.03254 | 0.02026 | 0.00379 |  | 0 |
|  | 0 | 0.00001 | 0.00008 | 0.00384 | 0.02468 | 0.03734 | 0.0624 | 0.08162 | 0.05922 | 0.01503 | 0.00609 | 0.00293 |  |
|  | 0.00213 | 0.00284 | 0.00075 | 0.00142 |  | 0 |  |  |  |  |  |  |  |
| 1993 | 1 | 1 | 3 | 0 | 126 | 0 | 0.00012 | 0.00001 | 0.00064 | 0.00864 | 0.01402 | 0.05882 |  |
|  | 0.16809 | 0.08456 | 0.08385 | 0.08023 | 0.05142 | 0.04641 | 0.0406 | 0.02042 | 0.007 | 0.005 | 0.00094 |  | 0 |


|  | 0 | 0 | 0.00203 | 0.00957 | 0.06125 | 0.11245 | 0.07924 | 0.04639 | 0.01194 | 0.00498 | 0.000060 | 0 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 0 | 0 | 0 | 0.0006 | 0 |  |  |  |  |  |  |  |
| 1994 | 1 | 1 | 3 | 0 | 117 | 0 | 0 | 0 | 0 | 0.00167 | 0.0112 | 0.02259 |
|  | 0.02581 | 0.04153 | 0.06489 | 0.1126 | 0.06874 | 0.07034 | 0.05595 | 0.05194 | 0.02649 | 0.01075 | 0.00073 | 0.0009 |
|  | 0 | 0 | 0 | 0.00184 | 0.04468 | 0.08946 | 0.12132 | 0.0972 | 0.06042 | 0.01519 | 0.0029 | 0.00021 |
|  | 0.00068 |  | 0 | 0 | 0 | 0 |  |  |  |  |  |  |
| 1995 | 1 | 1 | 3 | 0 | 114 | 0 | 0 | 0 | 0.00035 | 0.000 | 0.001110 | 0.00893 |
|  | 0.03026 | 0.05741 | 0.05007 | 0.08525 | 0.12008 | 0.09374 | 0.06827 | 0.0388 | 0.02381 | 0.00884 | 0.00242 | 0.00119 |
|  | 0.00175 |  | 0 | 0.00205 | 0 | 0.01412 | 0.03783 | 0.08782 | 0.14094 | 0.0774 | 0.03078 | 0.00468 |
|  | 0.00073 | 0.00171 | 0.00223 | 0.0049 | 0 | 0 | 0.00175 |  |  |  |  |  |
| 1996 | 1 | 1 | 3 | 0 | 116 | 0 | 0 | 0 | 0.00033 | 0.00445 | 0.03196 | 0.08891 |
|  | 0.08369 | 0.0443 | 0.04167 | 0.05217 | 0.04535 | 0.06299 | 0.06357 | 0.01947 | 0.01333 | 0.00335 | 0.00023 | 0.000190 |
|  | 0 | 0 | 0.00168 | 0.01966 | 0.10183 | 0.10599 | 0.06959 | 0.07843 | 0.0509 | 0.01033 | 0.00186 | 0.001940 .0005 |
|  | 0.00132 |  | 0 | 0 | 0 |  |  |  |  |  |  |  |
| 1997 | 1 | 1 | 3 | 0 | 136 | O | 0 | 0 | 0.00077 | 0.00202 | 0.00216 | 0.02881 |
|  | 0.12925 | 0.10512 | 0.03317 | 0.02917 | 0.05403 | 0.05664 | 0.04962 | 0.04472 | 0.01526 | 0.00855 | 0.0007 | 0.000010 |
|  | 0 | 0 | 0.0033 | 0.00045 | 0.06268 | 0.14975 | 0.09977 | 0.06919 | 0.02845 | 0.01467 | 0.00857 | 0.0001 |
|  | 0.00137 | 0.00127 | 0.00042 | 0 | 0 | 0 |  |  |  |  |  |  |
| 1998 | 1 | 1 | , | 0 | 123 | 0 | 0 | 0 | 0 | 0.00397 | 0.01444 | 0.0224 |
|  | 0.03925 | 0.06226 | 0.09141 | 0.0686 | 0.06555 | 0.07515 | 0.05957 | 0.04919 | 0.03089 | 0.00886 | 0.00108 | $0.0018 \quad 0$ |
|  | 0 | 0 | 0 | 0.04411 | 0.01694 | 0.06933 | 0.12133 | 0.08988 | 0.0328 | 0.02736 | 0.00183 | 0.000420 .0005 |
|  | 0.00085 | 0.00014 | 0.00003 | 0.00001 | 0 |  |  |  |  |  |  |  |
| 1999 | 1 | 1 | 3 | 0 | 84 | 0.00047 | 0.00112 | 0 | 0 | 0.00036 | 0.00233 | 0.03304 |
|  | 0.08849 | 0.0807 | 0.03665 | 0.06671 | 0.08052 | 0.05581 | 0.07201 | 0.05503 | 0.04537 | 0.01173 | 0.00715 | 0.000160 |
|  | 0 | 0 | 0 | 0.00011 | 0.03147 | 0.08443 | 0.10657 | 0.07571 | 0.04674 | 0.01023 | 0.006730 |  |
|  | 0.00002 | 0.00035 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |
| 2000 | . | 1 |  | 0 | 50 | 0 | 0 | 0 | 0.00228 | 0.00019 | 0.00019 | 0.00928 |
|  | 0.01157 | 0.02875 | 0.05166 | 0.05578 | 0.11252 | 0.10642 | 0.09753 | 0.11272 | 0.08519 | 0.03014 | 0.00908 | 0.00308 |
|  | 0.00002 |  | 0 | 0.00031 | 0 | 0.01031 | 0.02243 | 0.0715 | 0.0666 | 0.07021 | 0.0207 | 0.017190 .0016 |
|  | 0.00051 | 0.00101 | 0.00089 | 0.00033 | 0 | 0 |  |  |  |  |  |  |
| 2001 | 1 | 1 | . | 0 | 58 | 0 | 0 | 0 | 0.0083 | 0.01993 | 0.00771 | 0.01187 |
|  | 0.01642 | 0.03758 | 0.0536 | 0.05483 | 0.06074 | 0.05892 | 0.10988 | 0.03332 | 0.05608 | 0.0312 | 0.0132 | 0.056630 |
|  | 0 | 0 | 0.01426 | 0.02615 | 0.01599 | 0.02994 | 0.0876 | 0.10742 | 0.0699 | 0.01551 | 0.0022 | 0.000320 |
|  | 0.0004 | 0 | 0 | 0 | 0.00011 |  |  |  |  |  |  |  |
| 2002 | 1 | 1 | 3 | 0 | 54 | 0 | 0.00586 | 0.00114 | 0.00864 | 0.03363 | 0.071920 | 0.090170 .0404 |
|  | 0.02739 | 0.0244 | 0.01947 | 0.05204 | 0.05112 | 0.08519 | 0.0902 | 0.07081 | 0.04005 | 0.00877 | 0.00706 | 0.00113 |
|  | 0.00452 | 0.00124 | 0.0041 | 0.02706 | 0.07152 | 0.02883 | 0.03737 | 0.03884 | 0.03246 | 0.01081 | 0.00224 | 0.00322 |
|  | 0.00246 | 0.00284 | 0 | 0 | 0.00083 | 0.0023 |  |  |  |  |  |  |
| 2003 | , | 1 | 3 | 0 | 18 | 0 | 0 | 0 | 0.00218 | 0.00084 | 0.00031 | 0.00632 |
|  | 0.19441 | 0.31227 | 0.10404 | 0.01206 | 0.00536 | 0.00727 | 0.01577 | 0.01604 | 0.00329 | 0.00214 |  | 0.000960 |
|  | 0.00023 | 0.00011 | 0.00084 | 0.00011 | 0.07587 | 0.12785 | 0.0586 | 0.02396 | 0.02086 | 0.00712 | 0.001190 | 00 |
|  | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |
| 2004 | 1 | 1 | 3 | 0 | 54 | 0 | 0 | 0 | 0.00012 | 0.00048 | 0.00063 | 0.00095 |
|  | 0.00524 | 0.02633 | 0.21118 | 0.27406 | 0.05632 | 0.01742 | 0.03838 | 0.05902 | 0.04136 | 0.02919 | 0.0043 | 0 |
|  | 0 | 0 | 0.00023 | 0.00058 | 0.00026 | 0.02585 | 0.10078 | 0.07134 | 0.02827 | 0.00561 | 0.00212 | 00 |
|  | 0 | 0 | 0 | 0 | $0$ |  |  |  |  |  |  |  |
| 2005 | 1 | 1 | 3 | 0 | 20 | 0 | 0 | 0 | 0.00095 |  | 0 | 0 |
|  | 0.01986 | 0.0208 | 0.00037 | 0.06466 | 0.3323 | 0.18004 | 0.04388 | 0.04495 | 0.02574 | 0.01096 | 0 | 0 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0.06488 | 0.12996 | 0.03707 | 0.00865 | 0.00543 | 0 | 0.009490 |
|  | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |
| 2006 | 1 | 1 | 3 | 0 | 31 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0.00112 | 0.01377 | 0.00514 | 0.02027 | 0.08864 | 0.3692 | 0.25929 | 0.03989 | 0.06281 | 0.0263 | 0.00508 | 0.000530 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0.01525 | 0.01022 | 0.04 | 0.04166 | 0 | 0.000830 |
|  | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |


| \# Hook and line fishery |  |  |  |  |  |  |  | females |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | males |  |  |  |  |  |  |  |  |  |  |  |  |
| \#year | season | type | gender | partition \# | \# sample |  | 16 | 18 | 20 | 22 | 24 | 26 | 28 |
|  | 30 | 32 | 34 | 36 | 38 | 40 | 42 | 44 | 46 | 48 | 50 | 52 | 16 |
|  | 18 | 20 | 22 | 24 | 26 | 28 | 30 | 32 | 34 | 36 | 38 | 40 | 42 |
|  | 44 | 46 | 48 | 50 | 52 |  |  |  |  |  |  |  |  |
| 1980 | 1 | 2 | 3 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0.05346 | 0.0004 | 0.0002 | 0.10731 | 0.21581 | 0.62144 | 0.0004 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.0002 | 0 | 0 | 0 | 0 |
|  | 0.0002 | 0.0004 | 0 | 0 |  |  |  |  |  |  |  |  |  |
| 1982 | 1 | 2 | 3 | 0 | 20 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0.02656 | 0.07327 | 0.14654 | 0.35618 | 0.19872 | 0.17263 | 0.02609 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |
| 1983 | 1 | 2 | 3 | 0 | 8 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0.01666 | 0.14961 | 0.06663 | 0.09964 | 0.26559 | 0.38521 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0.01666 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |
| 1984 | 1 | 2 | 3 | 0 | 9 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0.05882 | 0.11765 | 0.17647 | 0.23529 | 0.17647 | 0.17647 | 0 | 0.05882 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |
| 1985 | 1 | 2 | 3 | 0 | 14 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0.00023 | 0.0222 | 0.10922 | 0.15438 | 0.09717 | 0.3143 | 0.15556 | 0.0774 | 0.01025 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0.01315 | 0.02107 | 0.0246 | 0 | 0 | 0 |  |
|  | 0.00047 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |
| 1986 | 1 | 2 | 3 | 0 | 8 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
|  | 0.00138 | 0 | 0.00204 | 0.00836 | 0.02555 | 0.14258 | 0.10739 | 0.35049 | 0.17396 | 0.11928 | 0.04642 | 0.0002 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0.00003 | 0 | 0 | 0.01824 | 0.0004 | 0 | 0 |  |
|  | 0.00191 | 0 | 0.00178 | 0 | 0 | 0 |  |  |  |  |  |  |  |
| 1987 | $1$ | 2 | 3 | $0$ | 9 | 0 | 0 | $0$ | 0 | 0 | 0 | $0.00657$ |  |
|  | $0.02064$ | 0.0066 | 0 | 0.05516 | 0.17066 | 0.23488 | 0.1451 | $0.10775$ | 0.05923 | 0.1022 | $0.00734$ | $0.00004$ |  |
|  | 0 | 0 | 0 | 0 | 0 | 0.00319 | 0.00657 | 0.00657 | 0.00319 | 0 | 0.06432 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |
| 1989 | 1 | 2 | 3 | 0 | 16 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0.03538 | 0.08849 | 0.08298 | 0.02435 | 0.0592 | 0.01779 | 0.01218 | 0.01826 | 0.02435 |  | 0 | $0$ | 0 |
|  | 0 | 0 | 0 | 0 | 0.01769 | 0.08846 | 0.05308 | 0.33615 | 0.12388 | 0.01769 | 0 | 0.00007 |  |
|  | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |
| 1990 | 1 | 2 | 3 | 0 | 16 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
|  | 0.00205 | 0 | 0.05716 | 0.16326 | 0.58683 | 0.16725 | 0 | 0.0032 | 0.00326 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0.00483 | 0 | 0.00526 | 0.00689 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |
| 1991 | 1 | 2 | 3 | 0 | 41 | 0 | 0.00143 | 0 | 0 | 0.00003 | 0.01129 | 0.00118 |  |
|  | 0.01025 | 0.06023 | 0.08648 | 0.19366 | 0.08308 | 0.15067 | 0.07261 | 0.05628 | 0.01759 | 0.00397 | 0.00164 |  | 0 |
|  | 0 | 0 | 0 | 0.00003 | 0.00045 | 0.02487 | 0.04852 | 0.09975 | 0.06582 | 0.00883 | 0.00088 | 0.00025 |  |
|  | 0.00019 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |
| 1992 | 1 | 2 | 3 | 0 | 84 | 0 | 0 | 0 | 0 | 0 | 0.00081 | 0.00155 |  |
|  | 0.03048 | 0.03815 | 0.08563 | 0.08881 | 0.1549 | 0.11131 | 0.13644 | 0.08134 | 0.03369 | 0.01247 | 0.00425 |  | 0 |
|  | 0 | 0 | 0 | 0 | 0.00315 | 0.01819 | 0.07305 | 0.05973 | 0.05016 | 0.01027 | 0.00158 | 0.00079 |  |
|  | 0.00311 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |
| 1993 | 1 | 2 | 3 | 0 | 87 | 0 | 0 | 0.00036 | 0 | 0 | 0.0251 | 0.10349 |  |
|  | 0.25814 | 0.18048 | 0.14098 | 0.08223 | 0.05605 | 0.00957 | 0.0072 | 0.0021 | 0.001 | 0.00086 | 0 | 0 | 0 |


|  | 0 | 0 | 0.00036 | 0.01122 | 0.02667 | 0.02754 | 0.02959 | 0.03582 | 0.00116 | 0.00007 |  | 0 | 0 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |
| 1994 | 1 | 2 | 3 | 0 | 86 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
|  | 0.00284 | 0.01322 | 0.04427 | 0.08209 | 0.16641 | 0.19531 | 0.21998 | 0.08578 | 0.03136 | 0.03328 | 0.00023 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0.03582 | 0.05304 | 0.02098 | 0.00407 | 0.0113 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |
| 1995 | 1 | 2 | 3 | 0 | 23 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
|  | 0.02018 | 0.02427 | 0.02279 | 0.10374 | 0.2622 | 0.10859 | 0.0662 | 0.02693 | 0.0042 | 0.00013 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0.01229 | 0.03623 | 0.0747 | 0.04455 | 0.06782 | 0.05856 | 0.03752 |  |
|  | 0.00387 | 0.01682 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |
| 1996 | 1 | 2 | 3 | 0 | 41 | 0 | 0 | 0 | 0 | 0 | 0 | 0.01667 | 0.0016 |
|  | 0.01394 | 0.08846 | 0.1179 | 0.22555 | 0.21468 | 0.07447 | 0.04815 | 0.03936 | 0.00221 | 0.00204 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0.01948 | 0.05499 | 0.06521 | 0.00247 | 0.01121 | 0 | 0.0016 | 0 |
|  | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |
| 1997 | 1 | 2 | 3 | 0 | 38 | 0 | 0 | 0 | 0 | 0.00215 | 0.00078 | 0 |  |
|  | 0.01598 | 0.08748 | 0.09409 | 0.08517 | 0.14414 | 0.19467 | 0.10841 | 0.07685 | 0.04188 | 0.01266 | 0.00378 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0.00303 | 0.03014 | 0.04673 | 0.02531 | 0.02327 | 0.00078 | 0.00239 | 0.00003 |  |
|  | 0.00027 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |
| 1998 | 1 | 2 | 3 | 0 | 38 | 0.00326 | 0 | 0 | 0 | 0 | 0 | 0.00563 | 0.0064 |
|  | 0.03196 | 0.13658 | 0.09991 | 0.06159 | 0.11968 | 0.13457 | 0.07747 | 0.04899 | 0.00844 | 0.00774 | 0.00391 | 0 | 0 |
|  | 0 | 0 | 0.00461 | 0.00326 | 0.00226 | 0.06047 | 0.09318 | 0.07127 | 0.01461 | 0.00047 | 0 | 0.00372 | 0 |
|  | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |
| 1999 | 1 | 2 | 3 | 0 | 11 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
|  | 0.02659 | 0.06492 | 0.07368 | 0.17232 | 0.24041 | 0.09193 | 0.11931 | 0.06458 | 0.02409 | 0.00238 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0.00467 | 0.00517 | 0.02843 | 0.04026 | 0.02993 | 0.01134 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |
| 2000 | 1 | 2 | 3 | 0 | 9 | 0 | 0 | 0 | 0 | 0 | 0.00031 | 0.00031 |  |
|  | 0.01411 | 0.02543 | 0.13084 | 0.25728 | 0.12122 | 0.16961 | 0.077 | 0.05276 | 0.0226 | 0.02131 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0.00031 | 0.01034 | 0.01534 | 0.04837 | 0.02074 | 0.00626 | 0 | 0 |  |
|  | 0.00587 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |
| 2001 | 1 | 2 | 3 | 0 | 12 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
|  | 0.00132 | 0 | 0.01175 | 0.03414 | 0.0829 | 0.11837 | 0.1749 | 0.12195 | 0.05119 | 0.02052 | 0.01335 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0.01026 | 0.06216 | 0.17562 | 0.10756 | 0.01241 | 0 | 0 | 0.0016 |
|  | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |
| 2002 | 1 | 2 | 3 | 0 | 3 | 0 | 0 | 0 | 0 | 0 | 0.02632 | 0.10526 | 0 |
|  | 0 | 0 | 0 | 0.02632 | 0 | 0 | 0.05263 | 0.02632 | 0.02632 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0.02632 | 0.02632 | 0 | 0.15789 | 0.39474 | 0.13158 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |
| 2006 | 1 | 2 | 3 | 0 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0.01272 | 0 | 0.16185 | 0.23815 | 0.25318 | 0.10867 | 0.05549 | 0.10636 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0.02543 | 0 | 0 | 0 | 0.02543 | 0.01272 | 0 |
|  | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |
| \# |  |  |  |  |  |  |  |  |  |  |  |  |  |

\#Net fishery
females

| \#year | season | type | gender | partition \# samples | 16 | 18 | 20 | 22 | 24 | 26 | 28 |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
|  | 30 | 32 | 34 | 36 | 38 | 40 | 42 | 44 | 46 | 48 | 50 | 52 | 16 |
|  | 18 | 20 | 22 | 24 | 26 | 28 | 30 | 32 | 34 | 36 | 38 | 40 | 42 |
|  | 44 | 46 | 48 | 50 | 52 |  |  |  |  |  |  |  |  |
| 1983 | 1 | 3 | 3 | 0 | 24 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0.01248 | 0.06211 | 0.14868 | 0.19754 | 0.332 | 0.13685 | 0.02443 | 0 | 0.00307 | 0 | 0 |


|  | 0 | 0 | 0 | 0 | 0 | 0.01248 | 0.03545 | 0.02297 |  | 0.01195 |  | 0 | 0 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |
| 198 | 1 | 3 | 3 | 0 | 68 | 0 | 0.01047 |  | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0.16667 | 0.29147 | 0.32045 | 0.10306 | 0.09742 | 0.01047 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |
| 1985 | 1 | 3 | 3 | 0 | 155 | 0 | 0 | 0 | 0 | 0 | 0 | 0.00122 |  |
|  | 0.00021 | 0.00467 | 0.02343 | 0.07395 | 0.09334 | 0.15591 | 0.24592 | 0.23791 | 0.06391 | 0.00509 | 0.00302 |  | 0 |
|  | 0 | 0 | 0 | 0 | 0.00015 | 0.00273 | 0.02204 | 0.03686 | 0.01733 | 0.01211 |  | 0.0002 | 0 |
|  | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |
| 1986 | 1 | 3 | 3 | 0 | 113 | 0 | 0 | 0 | 0 | 0 | 0 | 0.00023 | 0.0004 |
|  | 0.00057 | 0.00026 | 0.01582 | 0.06056 | 0.18991 | 0.18421 | 0.21071 | 0.20903 | 0.05679 | 0.00621 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0.00011 | 0.00566 | 0.02964 | 0.00568 | 0.00403 | 0.00343 | 0.00667 | 0 |
|  | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |
| 1987 | 1 | 3 | 3 | 0 | 92 | 0 | 0 | 0 | 0 | 0 | 0 | 0.00079 |  |
|  | 0.00162 | 0.00036 | 0.00232 | 0.00897 | 0.01165 | 0.19355 | 0.2855 | 0.17057 | 0.1123 | 0.0467 | 0.01564 | 0.00089 |  |
|  | 0 | 0 | 0 | 0 | 0 | 0.00347 | 0.04653 | 0.01944 | 0.01772 | 0.01386 | 0.04378 | 0.00194 |  |
|  | 0.00186 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |
| 1988 | 1 | 3 | 3 | 0 | 70 | 0 | 0 | 0 | 0 | 0 | 0.00041 | 0.00044 |  |
|  | 0.00117 | 0.0638 | 0.12296 | 0.00271 | 0.00163 | 0.00385 | 0.31123 | 0.257 | 0.09212 | 0.01448 | 0.00127 |  | 0 |
|  | 0 | 0 | 0 | 0.00006 | 0.00015 | 0.00097 | 0.11848 | 0.00267 | 0.00138 | 0.00279 | 0.00013 | 0.00005 | 0 |
|  | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |
| 1989 | 1 | 3 | 3 | 0 | 82 | 0 | 0 | 0 | 0 | 0 | 0 | 0.01848 |  |
|  | 0.01832 | 0.03839 | 0.12987 | 0.14382 | 0.11016 | 0.07334 | 0.12715 | 0.10056 | 0.13359 | 0.01859 | 0.01313 | 0.01893 |  |
|  | 0 | 0 | 0 | 0 | 0 | 0.0123 | 0.01375 | 0.01428 | 0.00822 | 0.00655 | 0.00043 | 0.00014 |  |
|  | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |
| 1990 | 1 | 3 | 3 | 0 | 99 | 0 | 0 | 0.00078 | 0 | 0 | 0.00057 | 0.0025 |  |
|  | 0.00785 | 0.01569 | 0.01327 | 0.0751 | 0.1624 | 0.13408 | 0.04108 | 0.2186 | 0.08537 | 0.05356 | 0.00613 | 0.00021 |  |
|  | 0 | 0 | 0 | 0 | 0 | 0.00171 | 0.0388 | 0.04572 | 0.02568 | 0.01163 | 0.04536 | 0.00371 |  |
|  | 0.0102 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |
| 1991 | 1 | 3 | 3 | 0 | 35 | 0 | 0 | 0 | 0 | 0 | 0.00144 | 0.00352 |  |
|  | 0.00863 | 0.0187 | 0.03612 | 0.08646 | 0.16717 | 0.23046 | 0.13553 | 0.04859 | 0.03628 | 0.00927 |  | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0.00016 | 0.02781 | 0.06585 | 0.05945 | 0.04155 | 0.00943 | 0.00767 | 0 |  |
|  | 0.00591 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |

\# 1992 length comps had several large males from Morro Bay area - probably mis-ID'd sex or species- thus sample size turned to negative 1

| 1992 | 1 | 3 | 3 | 0 | -1 | 0 | 0 | 0 | 0 | 0 | 0 | 0.00216 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 0.01539 | 0.00683 | 0.04506 | 0.07463 | 0.09314 | 0.14088 | 0.16453 | 0.10951 | 0.10248 | 0.06281 | 0.00667 | $0 \quad 0$ |
|  | 0 | 0 | 0 | 0 | 0.00139 | 0.01445 | 0.02481 | 0.08037 | 0.03203 | 0.01596 | 0.00178 | 0.00095 |
|  | 0.00059 | 0.00027 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |
| 1993 | 1 | 3 | 3 | 0 | 35 | 0 | 0 | 0 | 0 | 0 | 0.00102 | 0.00848 |
|  | 0.01798 | 0.0186 | 0.03445 | 0.10195 | 0.15712 | 0.24255 | 0.15447 | 0.09174 | 0.01546 | 0 | 0 | $0 \quad 0$ |
|  | 0 | 0 | 0 | 0.00473 | 0.00358 | 0.04126 | 0.06158 | 0.02809 | 0.01171 | 0.00428 | 0 | 0.000970 |
|  | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |
| 1994 | 1 | 3 | 3 | 0 | 47 | 0 | 0 | 0 | 0 | 0 | 0 | $0 \quad 0$ |
|  | 0.00085 | 0.01046 | 0.03534 | 0.05834 | 0.11516 | 0.34256 | 0.15397 | 0.0921 | 0.05238 | 0.00712 | 0 | $0 \quad 0$ |
|  | 0 | 0 | 0 | 0 | 0.00085 | 0.02841 | 0.03954 | 0.0351 | 0.0278 | 0 | 0 | $0 \quad 0$ |
|  | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |
| 1995 | 1 | 3 | 3 | 0 | 32 | 0 | 0 | 0 | 0 | 0 | 0 | $0 \quad 0$ |
|  | 0 | 0.00906 | 0 | 0.0436 | 0.08736 | 0.31989 | 0.22707 | 0.20206 | 0.07282 | 0.02 | 0 | $0 \quad 0$ |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.01813 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |
| 1996 | 1 | 3 | 3 | 0 | 21 | 0 | 0 | 0 | 0 | 0 | 0 | $0 \quad 0$ |
|  | 0 | 0.01626 | 0.03252 | 0.0813 | 0.1626 | 0.26016 | 0.25203 | 0.09756 | 0.07317 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0.01626 | 0 | 0 | 0 | 0 | 0 |


| 1997 | 1 | 3 | 3 | 0 | 14 | 0 | 0 | 0 | 0 | 0 | 00 | 0 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 0.01361 | 0.00537 | 0.00956 | 0.05249 | 0.15283 | 0.29519 | 0.25541 | 0.11019 | 0.01381 | 0.01074 | 00 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0.00517 | 0.01829 | 0.03229 | 0.02504 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |
| 1998 | 1 | 3 | 3 | 0 | 11 | 0 | 0 | 0 | 0 | 0 | 00 | 0 |
|  | 0 | 0.01304 | 0.0087 | 0.01739 | 0.14783 | 0.27391 | 0.33913 | 0.07826 | 0.02609 | 0 | $0 \quad 0$ | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0.02174 |  | 0.04783 | 0.01304 | 0 | 0.013040 | 0 |
|  | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |
| \# |  |  |  |  |  |  |  |  |  |  |  |  |
| \# Rec | length c | comps | Coastwi | de ( N and |  |  |  |  |  |  |  |  |


| \#year | season | type | gender | part | Nsamp | 16 | 18 | 20 | 22 | 24 | 26 | 28 | 30 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
|  | 32 | 34 | 36 | 38 | 40 | 42 | 44 | 46 | 48 | 50 | 52 | 16 | 18 |
|  | 20 | 22 | 24 | 26 | 28 | 30 | 32 | 34 | 36 | 38 | 40 | 42 | 44 |
|  | 46 | 48 | 50 | 52 |  |  |  |  |  |  |  |  |  |
| 1980 | 1 | 4 | 0 | 0 | 50 | 0.00255 | 0 | 0.01278 | 0.0358 | 0.07928 | 0.07672 | 0.13554 |  |
|  | 0.11253 | 0.11253 | 0.09718 | 0.07161 | 0.08439 | 0.07161 | 0.04092 | 0.02813 | 0.02301 | 0.01278 | 0 | 0.00255 |  |
|  | 0.00255 | 0 | 0.01278 | 0.0358 | 0.07928 | 0.07672 | 0.13554 | 0.11253 | 0.11253 | 0.09718 | 0.07161 | 0.08439 |  |
|  | 0.07161 | 0.04092 | 0.02813 | 0.02301 | 0.01278 | 0 | 0.0025 |  |  |  |  |  |  |
| 1981 | 1 | 4 | 0 | 0 | 47 | 0.00127 | 0 | 0 | 0.00508 | 0.02033 | 0.0343 | 0.06607 |  |
|  | 0.14485 | 0.11689 | 0.13214 | 0.10673 | 0.1385 | 0.08767 | 0.04447 | 0.04066 | 0.02668 | 0.02033 | 0.0127 | 0.00127 |  |
|  | 0.00127 | 0 | 0 | 0.00508 | 0.02033 | 0.0343 | 0.06607 | 0.14485 | 0.11689 | 0.13214 | 0.10673 | 0.1385 |  |
|  | 0.08767 | 0.04447 | 0.04066 | 0.02668 | 0.02033 | 0.0127 | 0.00127 |  |  |  |  |  |  |
| 1982 | 1 |  | 0 | 0 | 50 | 0 | 0 | 0 | 0.02427 | 0.0566 | 0.07605 |  |  |

 0.082520 .090610 .067960 .085760 .126210 .137540 .114880 .055010 .050160 .024270 .006470 .001610 $0 \quad 0 \quad 0 \quad 0.024270 .056630 .076050 .082520 .090610 .067960 .085760 .126210 .13754$ 0.114880 .055010 .050160 .024270 .006470 .00161
 0.156340 .128480 .078940 .054170 .071200 .092870 .077390 .044890 .043340 .023210 .003090 .001540 $\begin{array}{llllllllllllllllllllllllll}0 & 0.00464 & 0.01547 & 0.02321 & 0.07739 & 0.10371 & 0.15634 & 0.12848 & 0.07894 & 0.05417 & 0.0712 & 0.09287\end{array}$ 0.077390 .044890 .043340 .023210 .003090 .00154
$\begin{array}{lllllllllllllll}1984 & 1 & 4 & 0 & 0 & 90 & 0 & 0 & 0.00254 & 0.00636 & 0.01908 & 0.03053 & 0.0547 & 0.0916\end{array}$ $\begin{array}{lllllllllll}0.15267 & 0.20101 & 0.13613 & 0.07506 & 0.10432 & 0.07633 & 0.0318 & 0.01653 & 0.00127 & 0 & 0\end{array} 00$ $\begin{array}{lllllllllllllllllllll}0.00254 & 0.00636 & 0.01908 & 0.03053 & 0.0547 & 0.0916 & 0.15267 & 0.20101 & 0.13613 & 0.07506 & 0.10432 & 0.07633 & 0.0318\end{array}$ 0.016530 .0012700
$1985 \quad 1 \quad 4 \quad 0 \quad 0 \quad 138 \quad 0.000990 .000490 .001980 .005960 .009940 .018380 .03628$ $0.090450 .13320 .121760 .125240 .140150 .112820 .086970 .0656 \quad 0.029320 .013910 .005460 .00099$ 0.000990 .000490 .001980 .005960 .009940 .018380 .036280 .090450 .13320 .121760 .125240 .14015 $0.112820 .086970 .0656 \quad 0.029320 .013910 .005460 .00099$
1986 $\begin{array}{lllllllllllllll}0.04718 & 0.07769 & 0.1101 & 0.0958 & 0.10247 & 0.13203 & 0.09103 & 0.04385 & 0.0305 & 0.01096 & 0.00238 & 0.00047 & 0\end{array}$ $0.000950 .003810 .018580 .074350 .107240 .050520 .047180 .077690 .110100 .0958 \quad 0.102470 .13203$ $0.091030 .043850 .0305 \quad 0.010960 .002380 .00047$
$1987 \quad 1 \quad 4 \quad 0 \quad 0 \quad 22 \quad 0 \quad 0 \quad 0.007610 .017760 .045680 .083750 .12436$ 0.116750 .116750 .106590 .045680 .050760 .032990 .068520 .076140 .043140 .017760 .02030 .025380 $0 \quad 0.007610 .017760 .045680 .083750 .124360 .116750 .116750 .106590 .045680 .050760 .03299$ 0.068520 .076140 .043140 .017760 .02030 .02538

1988 1 $4 \begin{array}{lllllllllllllll}0 & 0 & 0 & 72 & 0 & 0 & 0 & 0.00323 & 0.02047 & 0.04956 & 0.12931\end{array}$
 $0 \quad 0 \quad 0.003230 .020470 .049560 .129310 .204740 .239220 .160560 .026930 .017240 .02693$ $0.061420 .039870 .011850 .006460 \quad 0.00215$
$\begin{array}{llllllllllll}1989 & 1 & 4 & 0 & 0 & 29 & 0 & 0 & 0 & 0.00219 & 0.0307 & 0.04495\end{array} 0.0921$ $\begin{array}{llllllllll}0.14692 & 0.1546 & 0.21052 & 0.21052 & 0.06469 & 0.02083 & 0.00986 & 0.00877 & 0.00328 & 0\end{array} 0$

|  | 0 0 | 0 | 0.00219 | 0.0307 | 0.04495 | 0.0921 | 0.14692 | 0.1546 | 0.210520 .210520 .064690 .02083 |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 0.00986 | 0.00877 | 0.00328 |  | 0 | 0 |  |  |  |  |  |  |  |
| 1994 | 1 | 4 | 0 | 0 | 5 | 0 | 0 | 0 | 0 | 0.09677 | 0.06451 | 0.16129 |  |
|  | 0.16129 | 0.2258 | 0.16129 | 0.09677 | 0.03225 | 0 | 0 | 0 | 0 | , | 0 |  | 0 |
|  | 0 | 0 | 0 | 0.09677 | 0.06451 | 0.16129 | 0.16129 | 0.2258 | 0.16129 | 0.09677 | 0.03225 |  | 0 |
|  | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |
| 1995 | 1 | 4 | 0 | 0 | 5 | 0 | 0 | 0 | 0.08053 | 0.05369 | 0.22147 | 0.26174 |  |
|  | 0.20134 | 0.12751 | 0.02684 | 0.02013 | 0 | 0 | 0 | 0 | , | 0 | 0 | 0.00671 |  |
|  | 0 | 0 | 0.08053 | 0.05369 | 0.22147 | 0.26174 | 0.201 | 0.12751 | 0.02684 | 0.020 | 0 | . | 0 |
|  | 0 | 0 | 0 | 0 | 0.00671 |  |  |  |  |  |  |  |  |
| 1996 | 1 | 4 | 0 | 0 | 20 | 0 | 0 | 0.00359 | 0.05215 | 0.07553 | 0.14928 | 0.19064 |  |
|  | 0.09892 | 0.07553 | 0.10431 | 0.07913 | 0.05935 | 0.05575 | 0.04136 | 0.01258 | 0.00179 |  | 0 |  | 0 |
|  | 0 | 0.00359 | 0.05215 | 0.07553 | 0.14928 | 0.19064 | 0.09892 | 0.07553 | 0.10431 | 0.07913 | 0.05935 | 0.05575 |  |
|  | 0.04136 | 0.01258 | 0.00179 | 0 | 0 | 0 |  |  |  |  |  |  |  |
| 1997 | 1 | 4 | 0 | 0 | 15 | 0 | 0 | 0 | 0.00338 | 0.0305 | 0.08305 | 0.05254 |  |
|  | 0.07627 | 0.05423 | 0.05423 | 0.07796 | 0.18474 | 0.17288 | 0.12542 | 0.05254 | 0.02203 | 0.00677 | 0.00338 |  | 0 |
|  | 0 | 0 | 0.00338 | 0.0305 | 0.08305 | 0.05254 | 0.07627 | 0.05423 | 0.05423 | 0.07796 | 0.18474 | 0.17288 |  |
|  | 0.12542 | 0.05254 | 0.02203 | 0.00677 | 0.00338 | 0 |  |  |  |  |  |  |  |
| 1998 | 1 | 4 | 0 | 0 | 6 | 0 | 0 | 0 | 0 | 0.0114 | 0.01901 | 0.06083 |  |
|  | 0.19771 | 0.13307 | 0.12167 | 0.08365 | 0.06463 | 0.11026 | 0.08745 | 0.07604 | 0.01901 | 0.0152 | 0 |  | 0 |
|  | 0 | 0 | 0 | 0.0114 | 0.01901 | 0.06083 | 0.19771 | 0.13307 | 0.12167 | 0.08365 | 0.06463 | 0.11026 |  |
|  | 0.08745 | 0.07604 | 0.01901 | 0.0152 | 0 | 0 |  |  |  |  |  |  |  |
| 1999 |  | 4 | 0 | 0 | 47 | 0 | 0.00516 | 0.01204 | 0.02065 | 0.02925 | 0.07056 | 0.07917 |  |
|  | 0.09294 | 0.06196 | 0.07228 | 0.06196 | 0.0981 | 0.11187 | 0.16179 | 0.09122 | 0.02409 | 0.00516 |  | 0.00172 |  |
|  | 0.00516 | 0.01204 | 0.02065 | 0.02925 | 0.07056 | 0.07917 | 0.09294 | 0.06196 | 0.07228 | 0.06196 | 0.0981 | 0.11187 |  |
|  | 0.16179 | 0.09122 | 0.02409 | 0.00516 | 0 | 0.00172 |  |  |  |  |  |  |  |
| 2000 | , | 4 | 0 | 0 | 31 | 0 | 0.01086 | 0.08695 | 0.06521 | 0.02898 | 0.07246 | 0.07608 | 0.0942 |
|  | 0.06521 | 0.0471 | 0.02173 | 0.05797 | 0.0942 | 0.09057 | 0.08695 | 0.08695 | 0.01086 | 0.00362 |  | 0 |  |
|  | 0.01086 | 0.08695 | 0.06521 | 0.02898 | 0.07246 | 0.07608 | 0.0942 | 0.06521 | 0.0471 | 0.02173 | 0.05797 | 0.0942 |  |
|  | 0.09057 | 0.08695 | 0.08695 | 0.01086 | 0.00362 | 0 |  |  |  |  |  |  |  |
| 2001 | 1 | 4 | 0 | 0 | 16 | 0 | 0 | 0.02675 | 0.09698 | 0.1806 | 0.0903 | 0.05685 |  |
|  | 0.05016 | 0.07692 | 0.05351 | 0.03678 | 0.05351 | 0.08361 | 0.07023 | 0.07023 | 0.04013 | 0.01337 |  |  | 0 |
|  | 0 | 0.02675 | 0.09698 | 0.1806 | 0.0903 | 0.05685 | 0.05016 | 0.07692 | 0.0535 | 0.03678 | 0.05351 | 0.0836 |  |
|  | 0.07023 | 0.07023 | 0.04013 | 0.01337 |  | 0 |  |  |  |  |  |  |  |
| 2002 | 1 | 4 | 0 | 0 | 18 | 0 | 0 | 0 | 0.00888 | 0.13777 | 0.14666 | 0.14666 |  |
|  | 0.07111 | 0.01333 | 0.02666 | 0.04888 | 0.00888 | 0.05333 | 0.07555 | 0.12 | 0.11111 | 0.02666 | 0.00444 |  | 0 |
|  | 0 | 0 | 0.00888 | 0.13777 | 0.14666 | 0.14666 | 0.07111 | 0.01333 | 0.02666 | 0.04888 | 0.00888 | 0.05333 |  |
|  | 0.07555 | 0.12 | 0.11111 | 0.02666 | 0.00444 | 0 |  |  |  |  |  |  |  |
| \#2004 | 1 | 4 | 0 | 0 | 41 | 0.00429 | 0.01716 | 0.01287 | 0.03433 | 0.11587 | 0.21459 | 0.13304 |  |
|  | 0.09442 | 0.1545 | 0.11158 | 0.07296 | 0.02575 | 0.00429 |  | 0 | 0.00429 |  | 0 | 0 |  |
|  | 0.00429 | 0.01716 | 0.01287 | 0.03433 | 0.11587 | 0.21459 | 0.13304 | 0.09442 | 0.1545 | 0.11158 | 0.07296 | 0.02575 |  |
|  | 0.00429 | 0 | 0 | 0.00429 | 0 | 0 | 0 |  |  |  |  |  |  |
| \#2005 | 1 | 4 | 0 | 0 | 16 | 0 | 0.07547 | 0.30188 | 0.09433 | 0.01886 | 0.07547 | 0.0566 |  |
|  | 0.09433 | 0.03773 | 0.01886 | 0.13207 | 0.0566 | 0.03773 |  | 0 | 0 | 0 | . |  | 0 |
|  | 0.07547 | 0.30188 | 0.09433 | 0.01886 | 0.07547 | 0.0566 | 0.09433 | 0.03773 | 0.01886 | 0.13207 | 0.0566 | 0.03773 |  |
|  | 0 |  | 0 | 0 | 0 |  |  |  |  |  |  |  |  |
| \# <br> \# Triennial survey length data- |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| $\begin{aligned} & \text { \# Trien } \\ & 1977 \end{aligned}$ | 1 | 5 | 3 | 0 | 56 | 0.00132 | 0.0028 | 0.01864 | 0.04554 | 0.02555 | 0.01866 | 0.01316 |  |
|  | 0.01863 | 0.04304 | 0.08371 | 0.05878 | 0.02463 | 0.03757 | 0.05619 | 0.05998 | 0.05109 | 0.04681 | 0.02098 | 0.00456 |  |
|  | 0.00157 | 0.0026 | 0.01833 | 0.04147 | 0.01525 | 0.01458 | 0.01431 | 0.06889 | 0.08181 | 0.06158 | 0.03506 | 0.00853 |  |
|  | 0.00065 | 0.00107 | 0.00148 | 0.00043 | 0.00057 | 0 | 0 |  |  |  |  |  |  |
| 1980 | 1 | 5 | 3 | 0 | 17 | 0 | 0 | 0 | 0 | 0 | 0.00102 | 0.00022 |  |
|  | 0.00442 | 0.03417 | 0.0489 | 0.06656 | 0.04987 | 0.08431 | 0.09185 | 0.06391 | 0.0378 | 0.0108 | 0.01103 | 0.00138 |  |
|  | 0 | 0.00092 | 0.00123 | 0.00056 | 0.00021 | 0.01013 | 0.06132 | 0.15277 | 0.18459 | 0.06082 | 0.00831 | 0.00208 |  |
|  | 0.00842 | 0.00156 | 0.00056 | 0.00014 |  | . |  |  |  |  |  |  |  |


| 1983 | 1 | 5 | $3 \quad 0$ | 0 | 17 | 0.00147 | 0.00236 | 0.00222 | 0.00237 | 0.01546 | 0.031550 .05519 |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 0.09165 | 0.11927 | 0.04888 | 0.01741 | 0.01022 | 0.02294 | 0.02131 | 0.01335 | 0.01473 | 0.01341 | 0.00281 | 0.00054 |  |
|  | 0.00129 | 0.00236 | 0.00082 | 0.00187 | 0.01964 | 0.04507 | 0.13632 | 0.1805 | 0.0633 | 0.03084 | 0.02869 | 0.00197 |  |
|  | 0 | 0.00003 | 0 | 0 | 0 | O |  |  |  |  |  |  |  |
| 1986 | 1 | 5 | 3 | 0 | 14 | 0.00021 | 0.00021 | 0.054 | 0.09675 | 0.10531 | 0.03826 | 0.00166 |  |
|  | 0.00191 | 0.00319 | 0.01658 | 0.03826 | 0.06103 | 0.04773 | 0.04995 | 0.01422 | 0.00968 | 0.00458 | 0.00138 |  | 0 |
|  | 0.00214 | 0.042 | 0.0741 | 0.12401 | 0.01268 | 0.01143 | 0.06192 | 0.07889 | 0.03768 | 0.0074 | 0.00226 | 0.00044 |  |
|  | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |
| 1989 | 1 | 5 | 3 | 0 | 91 | 0.14115 | 0.08542 | 0.00522 | 0.01077 | 0.0188 | 0.01236 | 0.02578 |  |
|  | 0.03328 | 0.01295 | 0.01263 | 0.03708 | 0.04408 | 0.00765 | 0.01092 | 0.01361 | 0.00611 | 0.00323 | 0.00099 | 0.00065 |  |
|  | 0.15814 | 0.07824 | 0.00423 | 0.01606 | 0.01862 | 0.03192 | 0.05855 | 0.05072 | 0.05481 | 0.02932 | 0.01254 | 0.00347 |  |
|  | 0.00022 | 0.00004 | 0.00005 | 0 | , | 0.00009 | 0.00009 |  |  |  |  |  |  |
| 1992 | 1 | 5 | 3 | 0 | 59 | 0.24397 | 0.02135 | 0.01956 | 0.025 | 0.00991 | 0.0186 | 0.04261 |  |
|  | 0.03886 | 0.01397 | 0.00795 | 0.00448 | 0.00373 | 0.00244 | 0.00253 | 0.00212 | 0.00026 | 0.00065 | 0.00006 |  | 0.2715 |
|  | 0.01878 | 0.02134 | 0.02997 | 0.01546 | 0.0718 | 0.06547 | 0.0214 | 0.01717 | 0.00594 | 0.00245 | 0.00024 | 0.0000 | 0 |
|  | 0 | 0 | 0.00012 | 0.00006 | 0 |  |  |  |  |  |  |  |  |
| 1995 | 1 | 5 | 3 | 0 | 79 | 0.07182 | 0.0105 | 0.02365 | 0.03701 | 0.03052 | 0.00774 | 0.01664 |  |
|  | 0.03555 | 0.02933 | 0.02137 | 0.02177 | 0.04439 | 0.03114 | 0.02686 | 0.02366 | 0.01874 | 0.00794 | 0.00212 | 0.00033 |  |
|  | 0.08029 | 0.0065 | 0.02289 | 0.03343 | 0.02708 | 0.04323 | 0.06932 | 0.08634 | 0.09242 | 0.05937 | 0.01576 | 0.00175 |  |
|  | 0.00006 | 0.00016 | 0.00008 | 0.00008 | 0 |  | 0 |  |  |  |  |  |  |
| 1998 | 1 | 5 | 3 | 0 | 81 | 0.01317 | 0.03329 | 0.02219 | 0.01371 | 0.05545 | 0.10907 | 0.02906 |  |
|  | 0.01489 | 0.0305 | 0.05614 | 0.00735 | 0.00612 | 0.01038 | 0.01613 | 0.00776 | 0.00386 | 0.00265 | 0.00042 |  |  |
|  | 0.00908 | 0.02868 | 0.02244 | 0.03439 | 0.12487 | 0.07326 | 0.08847 | 0.09834 | 0.06031 | 0.02068 | 0.00673 | 0.00042 |  |
|  | 0 | 0 | 0.00003 |  | 0 | 0 |  |  |  |  |  |  |  |
| 2001 | 1 | 5 | 3 | 0 | 77 | 0.00367 | 0.01002 | 0.05792 | 0.2417 | 0.11619 | 0.00883 | 0.0066 |  |
|  | 0.00424 | 0.00695 | 0.00655 | 0.00921 | 0.00452 | 0.00343 | 0.00301 | 0.00261 | 0.00244 | 0.00065 | 0.00001 |  |  |
|  | 0.00531 | 0.00575 | 0.09168 | 0.27631 | 0.08195 | 0.00664 | 0.01412 | 0.018 | 0.00695 | 0.00373 | 0.00063 | 0.00013 |  |
|  | 0.00001 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |
| 2004 | 1 | 5 | 3 | 0 | 88 | 0.11449 | 0.00173 | 0.00278 | 0.00155 | 0.00074 | 0.0159 | 0.01839 |  |
|  | 0.00552 | 0.01475 | 0.07254 | 0.14576 | 0.06047 | 0.01188 | 0.00359 | 0.00538 | 0.00669 | 0.00589 | 0.00154 | 0.00022 | 0.1552 |
|  | 0.00081 | 0.0029 | 0.0018 | 0.00745 | 0.01609 | 0.05755 | 0.12913 | 0.1032 | 0.02382 | 0.01048 | 0.00153 | 0.00004 |  |
|  | 0 | 0.00004 |  |  | 0 |  |  |  |  |  |  |  |  |
| \# <br> \# NWC combo survey |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| \#year | season | type | gender | part | \#_samp |  | 18 | 20 |  | 24 | 26 | 28 | 30 |
|  | 32 | 34 | 36 | 38 | 40 |  | 44 | 46 | 48 | 50 | 52 | 16 | 18 |
|  | 20 | 22 | 24 | 26 | 28 |  | 32 | 34 | 36 | 38 | 40 | 42 | 44 |
|  | 46 | 48 | 50 | 52 |  |  |  |  |  |  |  |  |  |
| 2003 | 1 | 6 |  | 0 | 91 | 0.00298 | 0.00807 | 0.00688 | 0.00342 | 0.00746 | 0.00424 | 0.0096 |  |
|  | 0.02817 | 0.1095 | 0.18554 | 0.03815 | 0.00738 | 0.00217 | 0.00154 | 0.00099 | 0.00393 | 0.00067 | 0.00251 | 0 |  |
|  | 0.00677 | 0.01157 | 0.0043 | 0.00725 | 0.00539 | 0.01074 | 0.08931 | 0.19781 | 0.11868 | 0.09394 | 0.03074 | 0.00002 |  |
|  | 0.00019 | 0 | 0 | 0.00002 | 0 | 0 | 0 |  |  |  |  |  |  |
| 2004 | 1 | 6 | 3 | 0 | 88 | 0.03914 | 0.01214 | 0.00471 | 0.03843 | 0.0303 | 0.01527 | 0.01859 |  |
|  | 0.01287 | 0.03111 | 0.07962 | 0.14332 | 0.08634 | 0.02108 | 0.0039 | 0.00402 | 0.00361 | 0.00326 | 0.0023 | 0.00012 |  |
|  | 0.03949 | 0.01135 | 0.00811 | 0.02011 | 0.01754 | 0.0103 | 0.02772 | 0.14081 | 0.13563 | 0.03042 | 0.00772 | 0.00057 |  |
|  | 0 | 0 | 0.00008 | 0 | 0 | 0 |  |  |  |  |  |  |  |
| 2005 | 1 | 6 | 3 | 0 | 91 | 0.01717 | 0.00979 | 0.01818 | 0.01461 | 0.00422 | 0.00865 | 0.00481 | 0.0195 |
|  | 0.01542 | 0.03592 | 0.19109 | 0.14109 | 0.04185 | 0.01576 | 0.00738 | 0.00624 | 0.00384 | 0.00164 | 0.0004 | 0.02127 |  |
|  | 0.01078 | 0.01367 | 0.01604 | 0.00897 | 0.00515 | 0.09415 | 0.14629 | 0.08918 | 0.03161 | 0.00381 | 0.00036 | 0.0011 | 0 |
|  | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |
| 2006 | 1 | 6 | 3 | 0 | 70 | 0.00242 | 0.00734 | 0.00929 | 0.01924 | 0.01731 | 0.01448 | 0.01335 |  |
|  | 0.00833 | 0.01775 | 0.01951 | 0.01799 | 0.05114 | 0.10618 | 0.08986 | 0.02131 | 0.02241 | 0.00883 | 0.00433 | 0.00089 |  |
|  | 0.00113 | 0.00712 | 0.00966 | 0.02279 | 0.02103 | 0.02015 | 0.01599 | 0.04448 | 0.15975 | 0.21062 | 0.03326 | 0.00071 |  |
|  | 0.00021 | 0.00113 | 0 | 0 | 0 | 0 | - |  |  |  |  |  |  |

\#Recreational Length data - June 15 fix to TL-> FL conversion!!

| \#year | season | type | gender | part | numsan |  | 16 | 18 | 20 | 22 | 24 | 26 | 28 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 30 | 32 | 34 | 36 | 38 | 40 | 42 | 44 | 46 | 48 | 50 | 52 | 16 |
|  | 18 | 20 | 22 | 24 | 26 | 28 | 30 | 32 | 34 | 36 | 38 | 40 | 42 |
|  | 44 | 46 | 48 | 50 | 52 |  |  |  |  |  |  |  |  |
| 1987 | 1 | 10 | 0 | 0 | 43 | 0.0007 | 0 | 0.0014 | 0.01131 | 0.03182 | 0.13932 | 0.30622 |  |
|  | 0.31046 | 0.13649 | 0.01909 | 0.01202 | 0.01202 | 0.01131 | 0.00353 | 0.00353 | 0.0007 | 0 | 0 | 0 | 0.0007 |
|  | 0 | 0.00141 | 0.01131 | 0.03182 | 0.13932 | 0.30622 | 0.31046 | 0.13649 | 0.01909 | 0.01202 | 0.01202 | 0.01131 |  |
|  | 0.00353 | 0.00353 | 0.0007 | 0 | 0 | 0 |  |  |  |  |  |  |  |
| 1988 | 1 | 10 | 0 | 0 | 44 | 0.0011 | 0.00221 | 0.00832 | 0.03329 | 0.07103 | 0.07047 | 0.12042 |  |
|  | 0.22031 | 0.24028 | 0.15149 | 0.04495 | 0.00832 | 0.00998 | 0.00887 | 0.00277 | 0.00166 | 0.00332 | 0.0011 | 0 | 0.0011 |
|  | 0.00221 | 0.00832 | 0.03329 | 0.07103 | 0.07047 | 0.12042 | 0.22031 | 0.24028 | 0.15149 | 0.04495 | 0.00832 | 0.00998 |  |
|  | 0.00887 | 0.00277 | 0.00166 | 0.00332 | 0.0011 | 0 |  |  |  |  |  |  |  |
| 1989 | 1 | 10 | 0 | 0 | 58 | 0 | 0.00122 | 0.00183 | 0.01102 | 0.02205 | 0.03063 | 0.09803 |  |
|  | 0.19852 | 0.17401 | 0.1734 | 0.17095 | 0.06617 | 0.02205 | 0.0147 | 0.00857 | 0.00428 | 0.00183 | 0 | 0.00061 |  |
|  | 0.00122 | 0.00183 | 0.01102 | 0.02205 | 0.03063 | 0.09803 | 0.19852 | 0.17401 | 0.1734 | 0.17095 | 0.06617 | 0.02205 | 0.0147 |
|  | 0.00857 | 0.00428 | 0.00183 | 0 | 0.00061 |  |  |  |  |  |  |  |  |
| 1990 | 1 | 10 | 0 | 0 | 16 | 0 | 0 | 0 | 0 | 0.00716 | 0.04659 | 0.09318 |  |
|  | 0.15412 | 0.17204 | 0.07526 | 0.10394 | 0.17921 | 0.09318 | 0.04659 | 0.02508 | 0.00358 | 0 | 0 |  | 0 |
|  | 0 | 0 | 0 | 0.00716 | 0.04659 | 0.09318 | 0.15412 | 0.17204 | 0.07526 | 0.10394 | 0.17921 | 0.09318 |  |
|  | 0.04659 | 0.02508 | 0.00358 | 0 | 0 | 0 |  |  |  |  |  |  |  |
| 1991 | 1 | 10 | 0 | 0 | 15 | 0 | 0 | 0.00256 | 0.01794 | 0.04615 | 0.12564 | 0.11794 |  |
|  | 0.14871 | 0.07948 | 0.05128 | 0.04871 | 0.12051 | 0.10769 | 0.06923 | 0.04358 | 0.01794 | 0.00256 | 0 |  | 0 |
|  | 0 | 0.00256 | 0.01794 | 0.04615 | 0.12564 | 0.11794 | 0.14871 | 0.07948 | 0.05128 | 0.04871 | 0.12051 | 0.10769 |  |
|  | 0.06923 | 0.04358 | 0.01794 | 0.00256 | 0 | 0 |  |  |  |  |  |  |  |
| 1992 | 1 | 10 | 0 | 0 | 32 | 0 | 0 | 0.00941 | 0.04143 | 0.05775 | 0.15379 | 0.20966 |  |
|  | 0.17137 | 0.09165 | 0.05963 | 0.03766 | 0.04331 | 0.04959 | 0.05524 | 0.00941 | 0.0069 | 0.00251 | 0.00062 |  | 0 |
|  | 0 | 0.00941 | 0.04143 | 0.05775 | 0.15379 | 0.20966 | 0.17137 | 0.09165 | 0.05963 | 0.03766 | 0.04331 | 0.04959 |  |
|  | 0.05524 | 0.00941 | 0.0069 | 0.00251 | 0.00062 | 0 |  |  |  |  |  |  |  |
| 1993 | 1 | 10 | 0 | 0 | 37 | 0 | 0.00061 | 0.00553 | 0.02642 | 0.0381 | 0.08358 | 0.09649 |  |
|  | 0.13952 | 0.16041 | 0.11124 | 0.07682 | 0.05777 | 0.06883 | 0.06084 | 0.03749 | 0.02274 | 0.01167 | 0.00184 | 0 | 0 |
|  | 0.00061 | 0.00553 | 0.02642 | 0.0381 | 0.08358 | 0.09649 | 0.13952 | 0.16041 | 0.11124 | 0.07682 | 0.05777 | 0.06883 |  |
|  | 0.06084 | 0.03749 | 0.02274 | 0.01167 | 0.00184 | $+0$ |  |  |  |  |  |  |  |
| 1994 | 1 | 10 | 0 | 0 | 26 | 0.0008 | 0.00161 | 0.00726 | 0.03069 | 0.10904 | 0.1155 | 0.1357 | 0.1042 |
|  | 0.10339 | 0.10985 | 0.11227 | 0.07108 | 0.0315 | 0.02827 | 0.02019 | 0.01615 | 0.00242 | 0 | 0 | 0.0008 |  |
|  | 0.00161 | 0.00726 | 0.03069 | 0.10904 | 0.1155 | 0.1357 | 0.1042 | 0.10339 | 0.10985 | 0.11227 | 0.07108 | 0.0315 |  |
|  | 0.02827 | 0.02019 | 0.01615 | 0.00242 | 0 | 0 |  |  |  |  |  |  |  |
| 1995 | 1 | 10 | 0 | 0 | 22 | 0 | 0.00892 | 0.05535 | 0.03928 | 0.06428 | 0.07142 | 0.10535 |  |
|  | 0.10892 | 0.18214 | 0.10892 | 0.08571 | 0.06785 | 0.05357 | 0.02321 | 0.01607 | 0.00714 | 0.00178 | 0 | 0 | 0 |
|  | 0.00892 | 0.05535 | 0.03928 | 0.06428 | 0.07142 | 0.10535 | 0.10892 | 0.18214 | 0.10892 | 0.08571 | 0.06785 | 0.05357 |  |
|  | 0.02321 | 0.01607 | 0.00714 | 0.00178 | 0 | 0 |  |  |  |  |  |  |  |
| 1996 | 1 | 10 | 0 | 0 | 19 | 0 | 0 | 0.01167 | 0.02918 | 0.0642 | 0.11867 | 0.13035 | 0.0642 |
|  | 0.09533 | 0.13424 | 0.09338 | 0.10894 | 0.07782 | 0.05058 | 0.01945 | 0.00194 | 0 | 0 | 0 |  | 0 |
|  | 0.01167 | 0.02918 | 0.0642 | 0.11867 | 0.13035 | 0.0642 | 0.09533 | 0.13424 | 0.09338 | 0.10894 | 0.07782 | 0.05058 |  |
|  | 0.01945 | 0.00194 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |
| 1997 | 1 | 10 | 0 | 0 | 19 | 0 | 0 | 0 | 0.00523 | 0.04712 | 0.12565 | 0.08115 |  |
|  | 0.09162 | 0.04973 | 0.0445 | 0.06806 | 0.1335 | 0.17015 | 0.10471 | 0.04712 | 0.01832 | 0.01047 | 0.00261 | 0 | 0 |
|  | 0 | 0 | 0.00523 | 0.04712 | 0.12565 | 0.08115 | 0.09162 | 0.04973 | 0.0445 | 0.06806 | 0.1335 | 0.17015 |  |
|  | 0.10471 | 0.04712 | 0.01832 | 0.01047 | 0.00261 | 0 |  |  |  |  |  |  |  |
| 1998 | 1 | 10 | 0 | 0 | 9 | 0 | 0 | 0 | 0 | 0.00955 | 0.01592 | 0.0605 |  |
|  | 0.18471 | 0.13057 | 0.10828 | 0.08917 | 0.09554 | 0.12101 | 0.08598 | 0.07006 | 0.01592 | 0.01273 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0.00955 | 0.01592 | 0.0605 | 0.18471 | 0.13057 | 0.10828 | 0.08917 | 0.09554 | 0.12101 |  |
|  | 0.08598 | 0.07006 | 0.01592 | 0.01273 | 0 | 0 |  |  |  |  |  |  |  |
| \# |  |  |  |  |  |  |  |  |  |  |  |  |  |
| \# Age | mposition | n data |  |  |  |  |  |  |  |  |  |  |  |
| 21 | \# number | r of age | bins |  |  |  |  |  |  |  |  |  |  |


| 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |

$\begin{array}{lllllll}15 & 16 & 17 & 18 & 19 & 20 & 21\end{array}$
1 \# number of unique ageing error matrices to generate \# ageing error matrix- no bias, has imprecision (st dev)

| 0.5 | 1.5 | 2.5 | 3.5 | 4.5 | 5.5 | 6.5 | 7.5 | 8.5 | 9.5 | 10.5 | 11.5 | 12.5 | 13.5 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
|  | 14.5 | 15.5 | 16.5 | 17.5 | 18.5 | 19.5 | 20.5 | 21.5 |  |  |  |  |  |
| 0.03 | 0.091 | 0.153 | 0.214 | 0.275 | 0.336 | 0.398 | 0.459 | 0.52 | 0.581 | 0.643 | 0.704 | 0.765 | 0.826 |
|  | 0.888 | 0.949 | 1.01 | 1.072 | 1.133 | 1.194 | 1.255 | 1.317 |  |  |  |  |  |

61 \# number of age observations-
\# this run goes back to traditional age comps-

| \#year | season | type | gender | part | errmat | Lbinlo | LbinHi | \# samp | 1 | 2 | 3 | 4 | 5 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
|  | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 |
|  | 19 | 20 | plus | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 |
|  | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19 | 20 | plus |  |  |
| 1978 | 1 | 1 | 3 | 0 | 1 | 1 | 52 | -1 | 0 | 0 | 0.00378 | 0.00192 |  | $0.051930 .062290 .081030 .112050 .0285 \quad 0.023180 .139500 .041350 .008050 .004510 .011620 .01389$ $0.033250 .019760 .039870 .0299 \quad 0.0635 \quad 0 \quad 0 \quad 0 \quad 0.000860 .000940 .011080 .033270 .03173$ 0.024620 .008720 .002880 .011370 .023570 .021610 .043330 .001170 .001270 .002630 .000190 .001420 .0035 0.00597

$\begin{array}{llllllllllll}1979 & 1 & 1 & 3 & 0 & 1 & 1 & 52 & -1 & 0 & 0 & 0.02289 \\ 0.04417\end{array}$ $0.032560 .120650 .060670 .050470 .1531 \quad 0.090650 .036730 .02620010610 .002850 .027340 .01818$
 $0.024740 .003620 \quad 0.004620 .003350 .019170 .000440 .001410 .057460 .002230 .005310 .00335$ 0.00044
$\begin{array}{lllllllllllll}1980 & 1 & 1 & 3 & 0 & 1 & 1 & 52 & 120 & 0 & 0 & 0.00079 & 0.01116\end{array}$ 0.071180 .035580 .242430 .018480 .040770 .073960 .015130 .011600 .042320 .010380 .002310 .05865 $0.000110 .002440 .00290 .000440 .019730 \quad 0.001020 .004350 .007 \quad 0.057880 .077130 .04955$ 0.006220 .004310 .031010 .004370 .058130 .000710 .002660 .000960 .009180 .000280 .003330 .00621 0.001030 .01431
$\begin{array}{lllllllllllll}1981 & 1 & 1 & 3 & 0 & 1 & 1 & 52 & 80 & 0 & 0 & 0.00121 & 0.00551\end{array}$ 0.157770 .208490 .039430 .156070 .012130 .003780 .004980 .008350 .00390 .057090 .001820 .00056 $0.002450 .001940 .001010 .000210 .008060 \quad 0 \quad 0.049750 .000370 .054820 .024260 .00489$ 0.120490 .002150 .002080 .007770 .001530 .002610 .051390 .00070 .000080 .000070 .000240 0.000150 .00187
$\begin{array}{llllllllllll}1982 & 1 & 1 & 3 & 0 & 1 & 1 & 52 & 135 & 0 & 0.00006 & 0.00795 \\ 0.02247\end{array}$ $\begin{array}{lllllllllllllllll}0.05293 & 0.03563 & 0.21462 & 0.053 & 0.17273 & 0.01588 & 0.04724 & 0.04183 & 0.0206 & 0.01731 & 0.01459 & 0.00567\end{array}$ $0.007050 .002 \quad 0.011870 .000690 .012520 \quad 0 \quad 0.006460 .004620 .017030 .017670 .07607$ 0.019490 .047610 .008850 .012920 .014380 .002820 .007290 .004790 .000010 .000120000 0.000260 .00296
$\begin{array}{lllllllllllll}1983 & 1 & 1 & 3 & 0 & 1 & 1 & 52 & 254 & 0 & 0 & 0.00712 & 0.04191\end{array}$ 0.020140 .038820 .077280 .227970 .095970 .087510 .041050 .056160 .0338 0.02631 0.009680 .01863 $0.001110 .007510 .008260 .015260 .025350 \quad 0.000060 .005280 .028220 .010550 .007920 .02584$ 0.034550 .007010 .015610 .003060 .005640 .002990 .004950 .001470 .002180 .000570 .002770 0.000710 .00073
$\begin{array}{llllllllllllll}1984 & 1 & 1 & 3 & 0 & 1 & 1 & 52 & 202 & 0 & 0.00002 & 0.03783 & 0.10336\end{array}$ $0.173690 .086 \quad 0.050890 .043490 .091490 .026640 .027020 .013160 .022710 .013730 .024250 .00804$ 0.009120 .001850 .000510 .001060 .00579000003350 .010330 .046410 .030680 .017070 .013 0.015510 .033360 .027770 .013190 .019030 .005780 .004120 .002820 .010280 .002590 .000770 .00085 0.000120 .00234
$\begin{array}{lllllllllllll}1985 & 1 & 1 & 3 & 0 & 1 & 1 & 52 & 303 & 0 & 0.00002 & 0.00279 & 0.02507\end{array}$ $0.064760 .162040 .081040 .0408 \quad 0.035270 .036300 .042870 .027390 .028720 .0188 \quad 0.018710 .00889$ 0.004520 .005420 .004930 .002360 .00932000 .000060 .000110 .015360 .015440 .049360 .04948 0.032180 .029240 .047190 .036040 .02160 .019020 .026130 .006760 .006220 .005320 .003450 .00422 0.001340 .01145
$\begin{array}{llllllllllll}1986 & 1 & 1 & 3 & 0 & 1 & 1 & 52 & 111 & 0 & 0.00466 & 0.0088 \\ 0.02095\end{array}$ $0.077260 .1109 \quad 0.089030 .041270 .037360 .038830 .067670 .024470 .033810 .016990 .021670 .009$
$\begin{array}{llllllllllll}0.00728 & 0.00213 & 0.0115 & 0.00149 & 0.00566 & 0 & 0.00432 & 0.00224 & 0.00663 & 0.02418 & 0.05423 & 0.05353\end{array}$ 0.030770 .047010 .025410 .046620 .014930 .028990 .004220 .011790 .002630 .002120 .001450 .00082 0.000620 .00677
$\begin{array}{llllllllllllllll}1987 & 1 & 1 & 3 & 0 & 1 & 1 & 52 & 205 & 0.04462 & 0.03154 & 0.32482 & 0.01466\end{array}$ 0.010950 .031230 .041420 .065630 .016360 .002990 .004990 .015380 .003750 .006370 .00310 .0003 0.001240 .001500 .000910 .000210 .000330 .017850 .000090 .147460 .012240 .010890 .007330 .03271 $\begin{array}{lllllllllllll}0.05213 & 0.01475 & 0.01071 & 0.01644 & 0.0176 & 0.0049 & 0.01238 & 0.00473 & 0.00156 & 0.00458 & 0.00502 & 0.00004\end{array}$ 0.001110 .00318

1988 $\begin{array}{llllllllllllllllll}0.00423 & 0.00113 & 0.05054 & 0.01579 & 0.04125 & 0.00992 & 0.01415 & 0.00033 & 0.01861 & 0.00391 & 0.00258 & 0.00003 & 0.006\end{array}$ $0.002090 .000020 .000260 .003740 \quad 0.000290 .001180 .253770 .003710 .003550 .00840 .01968$ 0.046510 .014320 .001670 .007780 .004720 .000510 .002180 .010480 .001270 .009030 .000180 .00018 0.00099
$1989 \begin{array}{llllllllllll}1 & 1 & 1 & 3 & 0 & 1 & 1 & 52 & 174 & 0 & 0.00011 & 0.03457 \\ 0.03029\end{array}$ 0.429880 .001650 .000670 .008550 .008950 .017590 .002490 .001410 .000680 .008030 .0001000002070 $\begin{array}{lllllllllllll}0.00005 & 0.00022 & 0.00004 & 0.00045 & 0 & 0.00009 & 0.0226 & 0.03778 & 0.26056 & 0.00339 & 0.0004 & 0.02036\end{array}$ $\begin{array}{lllllllllllll}0.01849 & 0.03719 & 0.00432 & 0.00165 & 0.00124 & 0.01195 & 0.0142 & 0.00599 & 0.00869 & 0.00042 & 0.0009 & 0.00006\end{array}$ 0.00193
$\begin{array}{llllllllllll}1990 & 1 & 1 & 3 & 0 & 1 & 1 & 52 & 133 & 0 & 0.02742 & 0.05254 \\ 0.03834\end{array}$ 0.052850 .213030 .151810 .003140 .039760 .004410 .006420 .001110 .004970 .000560 .003170 .00028 0.001230 .000310 .000910 .001190 .004110 .000030 .013880 .038160 .0536 $0.004250 .013130 .014130 .0257 \quad 0.002960 .018040 .009420 .00790 .003450 .007280 .002590 .0012$ 0.000360 .00199
$\begin{array}{llllllllllllll}1991 & 1 & 1 & 3 & 0 & 1 & 1 & 52 & 66 & 0 & 0.03237 & 0.08143 & 0.08939\end{array}$ 0.065490 .049640 .150040 .035890 .009760 .011190 .012780 .009560 .001440 .0128 0 0.0008360 $\begin{array}{lllllllllllllll}0.00124 & 0 & 0 & 0.03012 & 0 & 0.01674 & 0.10708 & 0.05087 & 0.03811 & 0.01699 & 0.07145 & 0.02294\end{array}$ $0.005550 .0088 \quad 0.010730 .013340 .002110 .009110 .000720 .008270 .00010 .001990 .000120$ 0.01349
$\begin{array}{llllllllllll}1992 & 1 & 1 & 3 & 0 & 1 & 1 & 52 & 100 & 0 & 0.00306 & 0.088 \\ 0.12952\end{array}$ $\begin{array}{lllllllllllllll}0.10098 & 0.10262 & 0.05166 & 0.09095 & 0.03579 & 0.00788 & 0.01178 & 0.00858 & 0.0194 & 0.01313 & 0.01225 & 0.00157\end{array}$ $0.003010 .001570 .006110 .001280 .005510 \quad 0.0016 \quad 0.029280 .037580 .036870 .048470 .02022$ $\begin{array}{llllllllllllllll}0.06001 & 0.02501 & 0.0074 & 0.0019 & 0.00156 & 0.01092 & 0.00271 & 0.0066 & 0.00209 & 0.00136 & 0.00054 & 0.00501\end{array}$ 0.000040 .00615
$\begin{array}{llllllllllll}1993 & 1 & 1 & 3 & 0 & 1 & 1 & 52 & 75 & 0.00025 & 0.00174 & 0.02104 \\ 0.1297 & 0.118\end{array}$ $\begin{array}{lllllllllllllll}0.09357 & 0.05244 & 0.0481 & 0.07239 & 0.01097 & 0.00529 & 0.01416 & 0.0095 & 0.01103 & 0.00428 & 0.0025 & 0.00186\end{array}$ $0.002890 .000710 .005130 .001530 \quad 0.001660 .022010 .109170 .059450 .057010 .022660 .013810 .04$ 0.014380 .007940 .006440 .005070 .003060 .005830 .010280 .000960 .003550 .000570 .001920 .00717
$\begin{array}{llllllllllll}1994 & 1 & 1 & 3 & 0 & 1 & 1 & 52 & 76 & 0 & 0.00248 & 0.07104 \\ 0.0454\end{array}$ $\begin{array}{lllllllllllllllllll}0.13842 & 0.08056 & 0.09087 & 0.04623 & 0.01417 & 0.06873 & 0.02104 & 0.00153 & 0.00473 & 0.0061 & 0.00337 & 0.00383\end{array}$ $0.001470 .000610 .005880 .000620 .000980 \quad 0.0046$ $\begin{array}{lllllllllllll}0.02342 & 0.03959 & 0.03571 & 0.01772 & 0.00435 & 0.01236 & 0.00557 & 0.0056 & 0.0057 & 0.0051 & 0.00122 & 0.00013\end{array}$ 0.001050 .00494
$1995 \begin{array}{llllllllllll}1 & 1 & 1 & 3 & 0 & 1 & 1 & 52 & 57 & 0 & 0.00404 & 0.02541 \\ 0.0728\end{array}$ 0.086730 .125570 .082140 .061320 .040670 .018590 .042250 .012230 .003780 .006870 .005150 .00146 $\begin{array}{llllllllllllllll}0.00288 & 0.00047 & 0 & 0.00172 & 0.00367 & 0 & 0.00544 & 0.01632 & 0.03919 & 0.03082 & 0.05457 & 0.03673\end{array}$ $\begin{array}{lllllllllllllll}0.03411 & 0.03743 & 0.01884 & 0.03969 & 0.02024 & 0.01218 & 0.00496 & 0.00986 & 0.01253 & 0.00477 & 0.00522 & 0.00009\end{array}$ 0.009150 .01012
$\begin{array}{lllllllllllll}1996 & 1 & 1 & 3 & 0 & 1 & 1 & 52 & 64 & 0 & 0.00763 & 0.1728 & 0.01501\end{array}$
 $\begin{array}{lllllllllllllll}0.00284 & 0.00158 & 0 & 0.00624 & 0.00107 & 0 & 0.02565 & 0.11716 & 0.03339 & 0.034 & 0.04137 & 0.05519\end{array}$ $\begin{array}{lllllllllllllllllll}0.02609 & 0.02877 & 0.01265 & 0.02855 & 0.01731 & 0.01346 & 0.00214 & 0.00171 & 0.00015 & 0.00179 & 0.00063 & 0.01215\end{array}$ 0.003590 .00716
$\begin{array}{llllllllllll}1997 & 1 & 1 & 3 & 0 & 1 & 1 & 52 & 71 & 0 & 0.00132 & 0.01069 \\ 0\end{array}$ 0.073810 .065630 .062120 .059270 .045440 .031390 .016550 .012360 .011190 .001240 .004470 .00364 $\begin{array}{lllllllllllllllll}0.00324 & 0.00406 & 0.00196 & 0 & 0.00173 & 0 & 0 & 0.0152 & 0.14505 & 0.05635 & 0.04362 & 0.03408\end{array}$
$0.027590 .015790 .011250 .011110 .01760 .009230 .002090 .001230 .000560 .0022 \quad 0.005710 .00007$ 0.000990 .00552
$\begin{array}{llllllllllll}1998 & 1 & 1 & 3 & 0 & 1 & 1 & 52 & -1 & 0 & 0.00185 & 0.01358 \\ 0\end{array}$ 0.115790 .062330 .081080 .078690 .076420 .053780 .045270 .026230 .019280 .019910 .004290 .00127 $0.001870 .0018 \quad 0.00230 .000210 .007950 .000310 .000930 .018150 .014960 .064330 .010160 .04198$ 0.043950 .035720 .035410 .014610 .013510 .030560 .009850 .013850 .002310 .002310 .003260 .00503 0.002380 .00265
$\begin{array}{llllllllllll}1999 & 1 & 1 & 3 & 0 & 1 & 1 & 52 & -1 & 0 & 0.00006 & 0.00173 \\ 0.10925\end{array}$ $0.063150 .137960 .044080 .0662 \quad 0.048370 .050630 .046670 .019420 .012120 .009030 .00890 .00263$ $0.000080 .000940 .002050 .0029 \quad 0.00533000 .003320 .000070 .053040 .033790 .102620 .02641$ 0.041170 .025790 .020870 .012690 .008790 .004820 .00690 .007280 .004960 .003730 .002870 .002270 .0001 0.00702
$\begin{array}{lllllllllllll}2000 & 1 & 1 & 3 & 0 & 1 & 1 & 52 & -1 & 0 & 0.00002 & 0.00014 & 0.01344\end{array}$ 0.061780 .068350 .117760 .060010 .072940 .039550 .071040 .050610 .043650 .025050 .02180 .01716 0.002180 .000610 .003210 .005040 .00363000 .000030 .005100006830 .045770 .028920 .05689 0.019840 .033430 .009770 .02310 .012410 .036360 .002920 .009040 .004650 .007150 .000080 .00178 0.002680 .01525
$\begin{array}{lllllllllllll}2001 & 1 & 1 & 3 & 0 & 1 & 1 & 52 & 23 & 0.0009 & 0.01761 & 0.0093 & 0.02139\end{array}$ $0.035520 .132280 .070520 .132740 .054310 .048170 .026370 .026950 .028 \quad 0.025130 .005130 .004080$ $0.004050 .001020 \quad 0.005180 .0018 \quad 0.023580 .003360 .011420 .015980 .035430 .046570 .06113$ $\begin{array}{llllllllllllllllllll}0.01708 & 0.02996 & 0.0256 & 0.01227 & 0.01829 & 0.01634 & 0.00428 & 0.00515 & 0.01275 & 0.0018 & 0 & 0.00071\end{array}$ 0.00784
$\begin{array}{llllllllllll}2002 & 1 & 1 & 3 & 0 & 1 & 1 & 52 & 31 & 0.00126 & 0.00519 & 0.14825\end{array} 0.07593$ $0.033910 .034310 .073510 .046390 .095280 .029170 .040170 .020660 .052520 .0251 \quad 0.029630 .00392$ 0.010290 .016130 .001660 .000830 .003170 .00030 .003880 .072940 .038250 .008240 .012870 .02868 0.010710 .033510 .005610 .011740 .002480 .003510 .006830 .004420 .000520 .003170 .002470 0.000060 .00257
$\begin{array}{llllllllllll}2003 & 1 & 1 & 3 & 0 & 1 & 1 & 52 & 9 & 0 & 0.00016 & 0.01887 \\ 0.61473\end{array}$ 0.014140 .006930 .004840 .009610 .004410 .00410 .005120 .002210 .002760 .002210 .001020 .00307 $\begin{array}{lllllllllllll}0.00102 & 0.00118 & 0.00102 & 0 & 0 & 0 & 0.00063 & 0.01768 & 0.23438 & 0.0206 & 0.00197 & 0.00228\end{array}$
 0
$\begin{array}{lllllllllllll}2004 & 1 & 1 & 3 & 0 & 1 & 1 & 52 & 33 & 0 & 0.00099 & 0.00483 & 0.02117\end{array}$ 0.326770 .073460 .025480 .034220 .053850 .026610 .033640 .013540 .013350 .007630 .016560 .01126 $0.007440 .006540 .0117 \quad 0.004010 .001430 \quad 0 \quad 0 \quad 0.003130 .014170 .202070 .024580 .0176$ $0.001180 .009830 .011180 .003680 .001480 .003460 \quad 0.002030 .000740 .000740 .004340 .002030$ 0.00327
$\begin{array}{llllllllllll}2005 & 1 & 1 & 3 & 0 & 1 & 1 & 52 & 15 & 0 & 0.00082 & 0\end{array}$ $\begin{array}{llllllllllllllllll}0.11353 & 0.4349 & 0.04918 & 0.01954 & 0.02939 & 0.01235 & 0.00348 & 0.00256 & 0.0001 & 0.00985 & 0.0098 & 0.00251\end{array}$ $\begin{array}{lllllllllll}0.00256 & 0.00005 & 0.00251 & 0 & 0 & 0 & 0 & 0 & 0.03266 & 0.0368 & 0.14335\end{array} 0.02588$ $\begin{array}{lllllllllll}0.00343 & 0.00251 & 0.00343 & 0 & 0 & 0 & 0.00082 & 0.00251 & 0 & 0 & 0\end{array}$ 0.00343
\#
\#
\#

Hook-line males

| \#Hook and Line |  |  |  |  |  |  |  |  | \# samples |  | 1 | 2 | 3 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 |
|  | 17 | 18 | 19 | 20 | plus | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 |
|  | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19 | 20 | plus |
| 1985 | 1 | 2 | 3 | 0 | 1 | 1 | 52 | 1 | 0 | 0 | 0 | 0 | 0 |


|  |  |  |  |  |  |  |  |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 1986 | 0 | 0 | 0 | 0.00086 | 0.00343 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 2 | 3 | 0 | 1 | 1 | 52 | 3 | 0 | 0 | 0.00204 | 0.00148 |
| 0 |  |  |  |  |  |  |  |  |  |  |  | $\begin{array}{llllllllllllll}0.03329 & 0.04987 & 0.02766 & 0.1301 & 0.09393 & 0.15182 & 0.082 & 0.19844 & 0.00591 & 0.07306 & 0.04547 & 0.0265 & 0.0038\end{array}$ $\begin{array}{lllllllllll}0.04702 & 0.00225 & 0.00148 & 0.00004 & 0 & 0 & 0 & 0 & 0 & 0.00732 & 0\end{array}$ $\begin{array}{llllllllll}0.00394 & 0.00183 & 0.00028 & 0.00232 & 0.00408 & 0.0019 & 0.00014 & 0.00204 & 0 & 0\end{array} 0000$

$\begin{array}{llllllllllll}1987 & 1 & 2 & 3 & 0 & 1 & 1 & 52 & 7 & 0 & 0.02078 & 0\end{array}$ $\begin{array}{llllllllllll}0 & 0.00618 & 0.46082 & 0.0254 & 0.0622 & 0.0127 & 0.0876 & 0.0127 & 0 & 0 & 0 & 0.0622\end{array} 0$ $\begin{array}{llllllllllll}0.0622 & 0 & 0.00618 & 0 & 0 & 0 & 0.03158 & 0 & 0 & 0 & 0 & 0\end{array} 0$

|  | 0 | 0.00618 | 0 | 0 | 0 | 0 | 0.0622 | 0 | 0.0622 | 0 | 0 |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 1990 | 1 | 2 | 3 | 0 | 1 | 1 | 52 | 11 | 0 | 0 | 0 | 0.1 | 0 |


| 0.6 | 0.3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |

$1991 \begin{array}{llllllllllll}1 & 2 & 3 & 0 & 1 & 1 & 52 & 17 & 0 & 0.00476 & 0.01476 & 0.02609\end{array}$
 $0.002730 \quad 0 \quad 0 \quad 0 \quad 0 \quad 0.000570 .013810 .022570 .047660 .026720 .061080 .0148 \quad 0$ $\begin{array}{llllllllllll}0.0044 & 0.00532 & 0.01512 & 0 & 0.00692 & 0 & 0.00791 & 0 & 0.0099 & 0 & 0.00419 & 0 \\ 1 & 2 & 3 & 0 & 1 & 1 & 52 & 38 & 0 & 0 & 0.0014 & 0.03133\end{array}$ $0.076050 .136210 .0988 \quad 0.221810 .051910 .015750 .024860 .035490 .027680 .029430 .009760 .00214$ $0.004970 .000630 .008 \quad 0.0009 \quad 0.01247000 .000990 .000550 .014980 .046060 .037560 .02124$ 0.030450 .008640 .002960 .011370 .010030 .001670 .009780 .007040 .000230 .002980 .002720 .000490 0.00066
$\begin{array}{lllllllllllll}1993 & 1 & 2 & 3 & 0 & 1 & 1 & 52 & 20 & 0 & 0 & 0.06322 & 0.28475\end{array}$ 0.186810 .183070 .083290 .030990 .043440 .000950 .000310 .000330 .009860 .000560 .000090 .00034

 0
$\begin{array}{llllllllllll}1994 & 1 & 2 & 3 & 0 & 1 & 1 & 52 & 11 & 0 & 0 & 0.00204 \\ 0.01527\end{array}$ 0.050330 .066990 .128420 .130830 .127130 .227050 .031460 .005270 .026740 .024520 .018320 .003420 $\begin{array}{llllllllllllllll}0 & 0.00379 & 0 & 0.00629 & 0 & 0 & 0 & 0.0049 & 0.00981 & 0.00833 & 0.01471 & 0.0049\end{array}$ $\begin{array}{llllllllllll}0.01739 & 0.04386 & 0.00972 & 0 & 0.0049 & 0 & 0.0049 & 0 & 0 & 0 & 0 & 0\end{array} 0.0087$
$\begin{array}{lllllllllllll}1995 & 1 & 2 & 3 & 0 & 1 & 1 & 52 & 8 & 0 & 0 & 0.00187 & 0.01532\end{array}$ $\begin{array}{lllllllllll}0.02451 & 0.15618 & 0.20948 & 0.10585 & 0.06084 & 0.01692 & 0.0284 & 0.00986 & 0 & 0.00475 & 0\end{array}$ $0 \quad 0.000290 .000730 \quad 0 \quad 0 \quad 0 \quad 0 \quad 0 \quad 0.051060 .067840 .074690 .05575$ $0.025520 .012070 .025560 .005790 \quad 0.010210 .004020 \quad 0.004020 \quad 0.000290 .00873$ 0.01542
$1996 \begin{array}{llllllllllll}1 & 2 & 3 & 0 & 1 & 1 & 52 & 11 & 0 & 0 & 0.00672 & 0.0158\end{array}$ $\begin{array}{llllllllll}0.08338 & 0.10917 & 0.13115 & 0.12225 & 0.13751 & 0.06567 & 0.0743 & 0.0743 & 0.0139 & 0.00463\end{array} 0$ $\begin{array}{lllllllllll}0 & 0.00427 & 0.00463 & 0 & 0 & 0 & 0 & 0.00336 & 0.01008 & 0 & 0.00672\end{array} 0.01553$ $\begin{array}{lllllllllll}0.01035 & 0.08919 & 0.00854 & 0.00854 & 0 & 0 & 0 & 0 & 0 & 0 & 0 \\ 0 & 0\end{array}$ $\begin{array}{llllllllllll}1997 & 1 & 2 & 3 & 0 & 1 & 1 & 52 & 10 & 0 & 0 & 0.04794 \\ 0.20447\end{array}$ 0.085640 .132850 .152860 .082350 .088540 .039960 .02170 .026290 .010150 .002950 .007690 .001390
 $\begin{array}{llllllllllll}0.00057 & 0 & 0.00768 & 0 & 0.00768 & 0 & 0.00809 & 0 & 0 & 0 & 0 & 0 \\ 1 & 2 & 3 & 0 & 1 & 1 & 52 & -1 & 0 & 0 & 0.00213 & 0.02347\end{array}$
$\begin{array}{lllllllllllll}1998 & 1 & 2 & 3 & 0 & 1 & 1 & 52 & -1 & 0 & 0 & 0.00213 & 0.02347 \\ & 0.05733 & 0.06901 & 0.06024 & 0.08737 & 0.13578 & 0.15112 & 0.08453 & 0.04459 & 0.03388 & 0.02155 & 0.005 & 0.00189\end{array}$ $\begin{array}{llllllllll}0.00189 & 0.00402 & 0.00991 & 0 & 0.00927 & 0 & 0 & 0 & 0 & 0\end{array}$ $0.026220 .035 \quad 0.028120 .029590 .015470 .009910 .011790 .010040 .001890 .003010 .002130 .001890$ 0
$\begin{array}{lllllllllllll}1999 & 1 & 2 & 3 & 0 & 1 & 1 & 52 & -1 & 0 & 0 & 0 & 0.04742\end{array}$ $\begin{array}{lllllllllllll}0.08607 & 0.37575 & 0.09088 & 0.0561 & 0.0608 & 0.0513 & 0.07462 & 0.0102 & 0.00748 & 0.00669 & 0.00669 & 0\end{array}$ $\begin{array}{lllllllllll}0.00079 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0.00739 & 0.05183 \\ 0.00942\end{array}$ $\begin{array}{lllllllllll}0.01883 & 0.00079 & 0.00942 & 0 & 0.01338 & 0.00669 & 0.00079 & 0.00669 & 0 & 0 & 0\end{array}$ 0
2000 $\left.\begin{array}{llllllllllll}1 & 2 & 3 & 0 & 1 & 1 & 52 & -1 & 0 & 0.00132 & 0.02549 & 0.0523 \\ 0.09041 & 0.13052 & 0.10797 & 0.0791 & 0.05472 & 0.09137 & 0.01976 & 0.03555 & 0.00624 & 0.00059 & 0.00566 & 0.0152\end{array}\right)$ $0 \begin{array}{lllllllllllllllll}0 & 0.00059 & 0 & 0 & 0 & 0 & 0 & 0.01373 & 0.01241 & 0.05369 & 0.01579 & 0.01711\end{array}$ $\begin{array}{llllllllll}0.02931 & 0.03335 & 0.02255 & 0.0282 & 0.01579 & 0.01645 & 0 & 0.01241 & 0 & 0\end{array}$ 0.01241
$\begin{array}{lllllllllllll}2001 & 1 & 2 & 3 & 0 & 1 & 1 & 52 & -1 & 0 & 0 & 0 & 0.00172\end{array}$ 0.019540 .015520 .017530 .104580 .048130 .072980 .042950 .001720 .014510 .014510 .008910 .008910 $\begin{array}{llllllllllllllllll}0 & 0 & 0 & 0.00891 & 0 & 0 & 0 & 0.00891 & 0.01781 & 0.04683 & 0.09869 & 0.12771\end{array}$ $\begin{array}{llllllllll}0.03793 & 0.08648 & 0.04683 & 0.02902 & 0.05804 & 0 & 0.01451 & 0.02342 & 0 & 0.02342\end{array} 0$
2002

| 1 | 2 | 3 | 0 | 1 | 1 | 52 | -1 | 0 | 0 | 0.02632 | 0 |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 0.05263 | 0 | 0.05263 | 0.05263 | 0.02632 | 0 | 0 | 0 | 0 | 0 | 0 | 0.02632 | 0 |
| 0.02632 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.07895 | 0 | 0.10526 |  |
| 0.18421 | 0.13158 | 0.07895 | 0.10526 | 0 | 0.02632 | 0 | 0 | 0.02632 | 0 | 0 | 0 | 0 |

\#
\#
\#
Net - females
net - males

| \#Net |  |  |  |  |  |  |  | \# sampl |  | 1 | 2 | 3 | 4 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 |
|  | 18 | 19 | 20 | plus | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 |
|  | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19 | 20 | plus |  |
| 1983 | 1 | 3 | 3 | 0 | 1 | 1 | 52 | -1 | 0 | 0 | 0 | 0 |  |
|  | 0.02676 | 0.04003 | 0.09744 | 0.18161 | 0.13584 | 0.15997 | 0.09485 | 0.05798 | 0.01296 | 0.08973 | 0 | 0.0265 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.01353 | 0 | 0.03788 | 0 |
|  | 0.02491 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
| 1984 | 1 | 3 | 3 | 0 | 1 | 1 | 52 | 7 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0.10225 | 0.10225 | 0.23027 | 0.23108 | 0.14895 | 0.05153 | 0 | 0.05636 | 0.02576 | 0 | 0.01047 | 0 |
|  | 0 | 0.04106 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |
| 1985 | 1 | 3 | 3 | 0 | 1 | 1 | 52 | 36 | 0 | 0 | 0 | 0 | 0.0004 | 0.049850 .038870 .063370 .057680 .115560 .116590 .185430 .132590 .065120 .020130 .010980 .040880 .0085 $\begin{array}{llllllllllllll}0.02041 & 0.00005 & 0.00264 & 0 & 0 & 0 & 0.00033 & 0 & 0.00323 & 0.00046 & 0.00367 & 0.00463\end{array}$ $\begin{array}{lllllllll}0.00705 & 0.00807 & 0.00897 & 0.0089 & 0.00199 & 0 & 0.0041 & 0.00195 & 0\end{array} \quad 0.009650 .005230 .00269$

$\begin{array}{lllllllllllll}1986 & 1 & 3 & 3 & 0 & 1 & 1 & 52 & 41 & 0 & 0.00039 & 0.0003 & 0.00022\end{array}$ $0.000230 .018240 .101490 .0392 \quad 0.1235 \quad 0.144380 .126030 .089130 .053110 .013790 .075710 .0592$ $0.020770 .035450 .005550 .007220 .025240 \quad 0 \quad 0 \quad 0 \quad 0.000060 .000060 .005020 .00612$ 0.005730 .014980 .003550 .003170 .00150 .007350 .003510 .000490 .005550000000026 0.000570 .00026
$\begin{array}{lllllllllllll}1987 & 1 & 3 & 3 & 0 & 1 & 1 & 52 & 63 & 0 & 0 & 0.00408 & 0.0086\end{array}$ 0.025490 .024750 .061170 .201620 .067690 .031340 .106480 .176540 .040420 .09210 .009480 .01664 $0.012340 .009560 \quad 0.009450 .006410 .000190 \quad 0.002040 .004960 .002410 .000480 .00582$ $0.034640 .007740 .002590 .002450 .015520 .002740 .013930 \quad 0.000070 .00019000$ 0.000070
$\begin{array}{llllllllllll}1988 & 1 & 3 & 3 & 0 & 1 & 1 & 52 & 42 & 0 & 0 & 0.00067 \\ 0.1144\end{array}$ 0.001120 .004820 .029160 .037240 .147490 .045650 .037010 .074020 .260090 .002130 .041720 $0.025350 \quad 0.010090 \quad 0.071330 \quad 0.001010 \quad 0.047440 .001010 .001680$ $\begin{array}{lllllll}0.00168 & 0.00594 & 0.002020 & 0.00112 & 0.0323 & 0.00112 & 0.00101\end{array} 0 \quad 0 \quad 0.001350$ 0
1989 $\begin{array}{llllllllllll}1 & 3 & 3 & 0 & 1 & 1 & 52 & 68 & 0 & 0 & 0.00031 & 0.04789 \\ 0.41627 & 0 & 0.00348 & 0.00234 & 0.03069 & 0.33092 & 0.00052 & 0.03721 & 0.01504 & 0.04579 & 0.01175 & 0.01738\end{array}$ $\begin{array}{llllllllllll}0.00009 & 0 & 0.012240 & 0 & 0 & 0 & 0 & 0.00006 & 0.01467 & 0.00003 & 0\end{array}$ $0.000030 .000310 .000650 \quad 0.000030 .000430 \quad 0.011530 \quad 0.000120 .000220 \quad 0$
$\begin{array}{llllllllllll}1990 & 1 & 3 & 3 & 0 & 1 & 1 & 52 & 79 & 0 & 0.00227 & 0.00965 \\ 0.01093 & 0.0132\end{array}$ 0.275020 .048840 .001850 .005540 .123380 .093990 .046570 .019030 .0389 0.06318 0.000140 .03748 $\begin{array}{lllllllllllllll}0.00043 & 0 & 0 & 0.000140 & 0 & 0.00099 & 0.00426 & 0.00114 & 0.05594 & 0.00852 & 0.04089\end{array}$ 0.000570 .007810 .007530 .045720 .001420 .00170 .008380 .001990 .002270 .000140 .000140 .00057 0.01945
$\begin{array}{llllllllllll}1991 & 1 & 3 & 3 & 0 & 1 & 1 & 52 & 7 & 0 & 0 & 0.01502 \\ 0.01502\end{array}$ $0.088340 .113520 .405920 .082160 \quad 0.026060 .002210 .011930 \quad 0.009280 \quad 0.023850$ $\begin{array}{llllllllllllllllllllllllll}0 & 0 & 0 & 0 & 0 & 0 & 0.03004 & 0.00221 & 0.04373 & 0.01413 & 0.06537 & 0.00707 & 0\end{array}$ $\begin{array}{lllllllllllll}1992 & 1 & 3 & 3 & 0 & 1 & 1 & 5 & 0 & 0 & 0 & 0 & 0 \\ 0\end{array}$ 0.067070 .032440 .082850 .266580 .071670 .015410 .071760 .041820 .033680 .01750 .013850 .01981 $0.023530 .016240 .014720 \quad 0.002510 \quad 0 \quad 0.000480 .011620 .002950 .014330 .02943$ $0.073710 .009640 .001450 \quad 0.016 \quad 0.005310 .004910 .010540 \quad 0.006450 .000750 .005460$ 0
$\begin{array}{lllllllllllll}1993 & 1 & 3 & 3 & 0 & 1 & 1 & 52 & 12 & 0 & 0 & 0 & 0.01679\end{array}$ 0.037430 .048860 .102780 .118660 .283060 .049270 .025590 .053820 .059690 .054120 .014870 .02802 $0.003440 .013250 \quad 0 \quad 0 \quad 0 \quad 0 \quad 0.002330 .004650 .017$ 0.01254 0.00718 $\begin{array}{lllllllll}0.00799 & 0.022260 & 0 & 0.003030 & 0 & 0.001320 .002230 & 0.009810 & 0\end{array}$
$1994 \begin{array}{lllllllllllll}1 & 3 & 3 & 0 & 1 & 1 & 52 & 9 & 0 & 0 & 0 & 0\end{array}$ 0.012780 .070360 .105570 .135740 .121170 .237430 .020580 .024150 .050760 .046520 .014380 .005040 .0153 $\begin{array}{llllllllllllllllll}0.00719 & 0 & 0 & 0 & 0 & 0 & 0.00633 & 0.00922 & 0.00596 & 0.00547 & 0.01008 & 0.02065\end{array}$ $\begin{array}{llllllllll}0.009220 .03343 & 0 & 0 & 0 & 0 & 0.00811 & 0.019970 & 0 & 0 & 0\end{array}$ 0.00461
$\begin{array}{llllllllllllll}1995 & 1 & 3 & 3 & 0 & 1 & 1 & 52 & 3 & 0 & 0 & 0 & 0 & 0.0212\end{array}$ $\begin{array}{lllllllllllllllllllllllll}0.0212 & 0.0424 & 0.09385 & 0.0212 & 0.16669 & 0.30604 & 0.05738 & 0.03618 & 0.05955 & 0.04381 & 0.04787 & 0.03072\end{array}$ $\begin{array}{llllllllllll}0.030720 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0\end{array}$ $\begin{array}{llllllllllll}0.0212 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 \\ 1 & 3 & 3 & 0 & 1 & 1 & 52 & 2 & 0 & 0 & 0 & 0\end{array}$

| 1996 | 1 | 3 | 3 | 0 | 1 | 1 | 52 | 2 | 0 | 0 | 0 | 0 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | $\begin{array}{llllllll}0.033880 & 0.03388 & 0.13553 & 0.11862 & 0.08474 & 0.06776 & 0.237370 & 0.03388\end{array} 0 \quad 0.03388$ $\begin{array}{llllllllllll}0.06783 & 0.05092 & 0.05086 & 0 & 0.01697 & 0 & 0 & 0 & 0 & 0 & 0 & 0 \\ 0 & 0.03388 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 \\ 0\end{array}$

$1997 \quad 1 \quad 3 \begin{array}{llllllllllll}1 & 3 & 0 & 1 & 1 & 52 & 2 & 0 & 0 & 0 & 0\end{array}$ $0.055710 \quad 0.024550 .092540 .135980 .235130 .095370 .166190 \quad 0.036830 .033990$ $\begin{array}{llllllllllll}0.01228 & 0 & 0.01228 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 \\ 0.021720 & 0.021720 & 0.02172 & 0 & 0 & 0.01228 & 0 & 0 & 0 & 0 & 0\end{array}$ 0.02172
$\begin{array}{llllllllllllll}1998 & 1 & 3 & 3 & 0 & 1 & 1 & 52 & 3 & 0 & 0 & 0 & 0 & 0\end{array}$ $\begin{array}{llllllllllllllllll}0 & 0.0377 & 0.06604 & 0.16985 & 0.11951 & 0.19811 & 0.0786 & 0.10374 & 0.11006 & 0 & 0 & 0 & 0\end{array}$ $\begin{array}{lllllllllll}0.025130 & 0.009450 & 0 & 0 & 0 & 0 & 0 & 0.009450 & 0.022010\end{array}$
\#
$\begin{array}{llllllllllll}2004 & 1 & 6 & 3 & 0 & 1 & 1 & 52 & 87 & 0.0481 & 0.06947 & 0.0497\end{array} 0.0034$
 $\begin{array}{lllllllllllllllllllllll}0.000370 & 0.00004 & 0.0001 & 0.04323 & 0.03786 & 0.06075 & 0.01039 & 0.23843 & 0.02529 & 0.02268 & 0.00128\end{array}$ $\begin{array}{llllllllllllllll}0.00208 & 0 & 0.0006 & 0.0006 & 0 & 0 & 0.00077 & 0.00135 & 0.00081 & 0.0006 & 0 & 0.00008 & 0\end{array}$
\#
\# Mean size at age data
0 \# number of size at age observations
\# environmental data-
0 \# num env. Variables
0 \# num env. Observations
999 \# end of file

```
# ****************************************************************
# Chilipepper rockfish .ctl file
# final model from June 2007 STAR Panel
# SS2 Version 2.00c by_Richard_Methot_(NOAA);_using_Otter_Research_ADMB_7.0.1
#****************************************************************
#
#
1 # N Growth Patterns
1 #_-N_-submorphs
1 #_N_areas
1111111111#_area_assignments_for_each_fishery_and_survey
#_recruit_design_(G_Pattern_x_birthseas_x_area)_X_(0/1_flag)
1
0 #_recr_distr_interaction
0 #_Do_migration
#_movement_pattern_(for_each_season_x_source_x_destination)_input_(0/1_flag)_minage_maxage
00
2 #_Nblock_Designs
510}#\mathrm{ # blocks per design
19701979
19801988
1 9 8 9 1 9 9 1
1 9 9 2 1 9 9 8
19992006
# block design 2
1 9 7 2 1 9 7 7
19781980
19811983
1 9 8 4 1 9 8 6
1 9 8 7 1 9 8 9
19901992
19931995
19961998
19992001
20022006
0.5 #_fracfemale
1000 #_submorph_between/within
1 #vector_submorphdist_(-1_first_val_for_normal_approx)
4 #_natM_amin
5 #_natM_amax
2 #_Growth_Age-at-L1
18 #_Growth_Age-at-L2
0.1 #_SD_ad\overline{d}_to_LAA
0 #_\overline{C}V_\overline{Growth_P}\mathrm{ Pattern}
1 #_maturity_option
1 #_First_Mature_Age
3 #_parameter_offset_approach
1 #_env/block/dev_adjust_method(1/2)
-5 #_MGparm_Dev_Phase
#_growth_parms
#_LO HI INIT PRIOR PR_type SD PHASE env-var use_dev dev_minyr dev_maxyr dev_stddev Block
Block_Fxn
```

| 0.05 | 0.3 | 0.16 | 0.22 | 0 | 0.8 | -4 | 0 | 0 | 0 | 0 | 0.5 | 0 | 0 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| \#_Gpattern:_1_Gender:_1 |  |  |  |  |  |  |  |  |  |  |  |  |  |
| -3 | 3 | 0 | 0 | 0 | 0.8 | -4 | 0 | 0 | 0 | 0 | 0.5 | 0 | 0 |
| 5 | 50 | 19.659 | 19 | 0 | 20 | -2 | 0 | 0 | 0 | 0 | 0.5 | 0 | 0 |
| 25 | 70 | 47.3 | 45 | 0 | 20 | -2 | 0 | 0 | 0 | 0 | 0.5 | 0 | 0 |
| 0.05 | 0.3 | 0.1945 | 0.1772 | 0 | 0.8 | -2 | 0 | 0 | 0 | 0 | 0.5 | 1 | 0 |
| 0.02 | 0.5 | 0.06 | 0.065 | 0 | 0.8 | -2 | 0 | 0 | 0 | 0 | 0.5 | 0 | 0 |
| -3 | 3 | 0.06 | 0.065 | 0 | 0.8 | -2 | 0 | 0 | 0 | 0 | 0.5 | 0 | 0 |
| -6 | 3 | 0.232 | 0.1279 | 0 | 0.8 | -4 | 0 | 0 | 0 | 0 | 0.5 | 0 | 0 |
| \#_Gpattern:_1_Gender:_2 |  |  |  |  |  |  |  |  |  |  |  |  |  |
| -6 | 3 | 0 | 0 | 0 | 0.8 | -4 | 0 | 0 | 0 | 0 | 0.5 | 0 | 0 |
| -3 | 3 | -0.03 | -0.1 | 0 | 0.8 | -2 | 0 | 0 | 0 | 0 | 0.5 | 0 | 0 |
| -3 | 3 | -0.35-0.3 |  | 0 | 0.8 | -2 | 0 | 0 | 0 | 0 | 0.5 | 0 | 0 |
| -3 | 3 | 0.6050 |  | 0 | 0.8 | -2 | 0 | 0 | 0 | 0 | 0.5 | 0 | 0 |
| -3 | 3 | 0 | 0 | 0 | 0.8 | -3 | 0 | 0 | 0 | 0 | 0.5 | 0 | 0 |
| -3 | 3 | 0 | 0 | 0 | 0.8 | -3 | 0 | 0 | 0 | 0 | 0.5 | 0 | 0 |
| -3 | 3 | $4.05 \mathrm{e}-0$ | $064.1 \mathrm{e}-0$ |  | 0 | 0 | -3 | 0 | 0 | 0 | 0 | 0.5 | 0 |
| 0 \#_wt-len\&maturity |  |  |  |  |  |  |  |  |  |  |  |  |  |
| -3 | 10 | 3.2 | 3.25 | 0 | 0.5 | -3 | 0 | 0 | 0 | 0 | 0.5 | 0 | 0 |
| 1 | 50 | 25.713 |  | 0 | 0.8 | -3 | 0 | 0 | 0 | 0 | 0.5 | 0 | 0 |
| -3 | 3 | -0.316 | -0.3 | 0 | 0.8 | -3 | 0 | 0 | 0 | 0 | 0.5 | 0 | 0 |
| -3 | 3 | 1 | 1 | 0 | 0.8 | -3 | 0 | 0 | 0 | 0 | 0.5 | 0 | 0 |
| -3 | 3 | 0 | 0 | 0 | 0.8 | -3 | 0 | 0 | 0 | 0 | 0.5 | 0 | 0 |
| -3 | 3 | $2.24 \mathrm{e}-0$ | $062.2 \mathrm{e}-0$ |  | 0 | 0 | -3 | 0 | 0 | 0 | 0 | 0.5 | 0 |
|  | 0 |  |  |  |  |  |  |  |  |  |  |  |  |
| -3 | 10 | 3.32 | 3.32 | 0 | 0.05 | -3 | 0 | 0 | 0 | 0 | 0.5 | 0 | 0 |
| -4 | 4 | 0 | 0 | -1 | 99 | -3 | 0 | 0 | 0 | 0 | 0.5 | 0 | 0 |
| \#_recrdistribution_by_growth_pattern |  |  |  |  |  |  |  |  |  |  |  |  |  |
| -4 | 4 | 0 | 0 | -1 | 99 | -3 | 0 | 0 | 0 | 0 | 0.5 | 0 | 0 |
| \#_recrdistribution_by_area 1 |  |  |  |  |  |  |  |  |  |  |  |  |  |
| -4 | 4 | 4 | 0 | -1 | 99 | -3 | 0 | 0 | 0 | 0 | 0.5 | 0 | 0 |
| \#_recrdistribution_by_season 1 |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 1 | 1 | 1 | 1 | -1 | 99 | -3 | 0 | 0 | 0 | 0 | 0.5 | 0 | 0 |
| \#_cohort_growth_deviation |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 0 \#_custom_MG-env_setup |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 0 \#_custom_MG-block_setup |  |  |  |  |  |  |  |  |  |  |  |  |  |
| \#K block param setup (one setup for all devs) |  |  |  |  |  |  |  |  |  |  |  |  |  |
| \#_LO | HI | INIT | PRIOR |  | SD |  |  |  |  |  |  |  |  |
| -10 | 10 | 0 | 0 | 0 | . 5 | 5 |  |  |  |  |  |  |  |


| \#_Spawner-Recruitment |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 1 \#_SR_function |  |  |  |  |  |
| \#_LO | HI | INIT | PRIOR | PR_type SD | PHASE |
| 9 | 13 | 14 | 10 | 05 | 1 |
| 0.2 | 1 | 0.57 | 0.573 | $0 \quad 0.183$ | -4 |
| 0 | 2 | 1 | 1 | $0 \quad 1$ | -3 |
| -5 | 5 | 0 | 0 | $0 \quad 1$ | -3 |
| -5 | 5 | 0 | 0 | $0 \quad 1$ | -2 |
| 0.0 | 0.5 | 0.0 | 0.0 | -1. 99 | -2 \#_reserve for future autocorrelation |
| 0 \#_SR_env_link |  |  |  |  |  |
| 1 \#_SR_env_target_1 $=$ devs;_2 $=$ R0;_3 ${ }^{\text {- }}$ steepness |  |  |  |  |  |
| 1 \#do_recr_dev: $0=$ none; $1=$ devvector; $2=$ simple deviations |  |  |  |  |  |
| 1965 2006-3 3 2 \#_recr_devs |  |  |  |  |  |
| 1492 \#_first_yr_fullbias_adj_in_MPD |  |  |  |  |  |

```
\begin{tabular}{lllllll}
\multicolumn{8}{c}{ \#initial_F_parms } \\
\#_LO & HI & INIT & PRIOR & PR_type SD & PHASE \\
0 & 0.1 & 0 & 0.01 & 0 & 0.2 & -1 \\
0 & 0.1 & 0 & 0.05 & 0 & 0.2 & -1 \\
0 & 1 & 0 & 0 & 0 & 0.2 & -1 \\
0 & 1 & 0 & 0 & 0 & 0.2 & -1
\end{tabular}
#_Q_setup
# A=do power, B=env-var, C=extra SD, D=devtype(<0=mirror, 0/1=none, 2=cons, 3=rand, 4=randwalk); E=0=num/1=bio,
F=err_type
\begin{tabular}{lllllll} 
\#_A & B & C & D & E & F & \\
0 & 0 & 0 & 0 & 1 & 0 & \\
0 & 0 & 0 & 0 & 1 & 0 & \\
0 & 0 & 0 & 0 & 1 & 0 & \\
0 & 0 & 0 & 0 & 1 & 0 & \\
0 & 0 & 0 & 0 & 1 & 0 & \\
0 & 0 & 0 & 0 & 1 & 0 & \\
0 & 0 & 0 & 0 & 1 & 0 & \\
0 & 0 & 0 & 0 & 1 & 0 & \\
0 & 0 & 0 & 0 & 1 & 0 & \\
0 & 0 & 0 & 0 & 0 & 0 & \\
\# & & & & & & \\
\#_Q_parms(if_any) & & & & \\
\# LO HI INIT PRIOR & \\
\#-10 & 20 & 0 & 0 & 0 & 10 & -3 \# juv survey1 power \\
\#-10 & 20 & 0 & 0 & 0 & 10 & -3 \# juv survey2 power \\
\#-10 & 20 & 0 & 0 & 0 & 10 & 1 \# triennial q \\
\#-10 & 20 & 0 & 0 & 0 & 10 & 1 \# NWC combo q
\end{tabular}
#_size_selex_types
#_Pattern Discard Male Special
\begin{tabular}{lllll}
1 & 0 & 1 & 0 & \(\# 1\) \\
1 & 0 & 1 & 0 & \(\# 2\) \\
24 & 0 & 1 & 0 & \(\# 3\) \\
24 & 0 & 0 & 0 & \(\# 4\) \\
1 & 0 & 0 & 0 & \(\# 5\) \\
1 & 0 & 0 & 0 & \(\# 6\) \\
0 & 0 & 0 & 0 & \(\# 7\) \\
0 & 0 & 0 & 0 & \(\# 8\) \\
30 & 0 & 0 & 0 & \(\# 9\) \\
24 & 0 & 0 & 0 & \(\# 10\)
\end{tabular}
#
#_age_selex_types
#_Pattern Discard Male Special
    10000 # 1
    10000#2
    10000 # 3
    10000 # 4
    10000# 5
    10000#6
    11000#7
    11000#8
    10000#9
20010# 10
#_selex_parms
```



| -10 | 9 | 3.4 | 3.5 | 0 | 10 | 2 | 0 | 0 | 0 | 0 | 0.5 | 0 | 0 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | \# | WIDT | exp |  |  |  |  |  |  |  |  |  |  |
| -10 | 9 | 3.68 | 5 | 0 | 10 | 2 | 0 | 0 | 0 | 0 | 0.5 | 0 | 0 |
|  | \# | WIDT | exp |  |  |  |  |  |  |  |  |  |  |
| -10 | 9 | -3.37 | -4.5 | 0 | 10 | 2 | 0 | 0 | 0 | 0 | 0.5 | 0 | 0 |
|  | \# | INIT | logistic |  |  |  |  |  |  |  |  |  |  |
| -10 | 9 | 0.79 | 2.9 | 0 | 10 | 2 | 0 | 0 | 0 | 0 | 0.5 | 0 | 0 |
|  | \# | FINAL | logistic |  |  |  |  |  |  |  |  |  |  |
| \#_size_sel: 5 |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 5 | 35 | 15.7 | 25.7 | 0 | 10 | -2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 \# |
| 0.000001 |  | 35 | 0.0002 | 5 | 0 | 10 | -2 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 \# |  |  |  |  |  |  |  |  |  |  |  |  |  |
| \# size sel 6 |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 5 | 35 | 20 | 15 | 0 | 100 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 \# |
| 0.000001 |  | 35 | 14 | 5 | 0 | 10 | 2 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 \# |  |  |  |  |  |  |  |  |  |  |  |  |
| \#_size_sel: 7,8-none- pre recruit survey |  |  |  |  |  |  |  |  |  |  |  |  |  |
| \#_size_sel: 9 set to maturity- |  |  |  |  |  |  |  |  |  |  |  |  |  |
| \#_size_sel: 10 Rec CPUE |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 1 | 60 | 33.85 | 32 | 0 | 100 | 2 | 0 | 0 | 0 | 0 | 0.5 | 0 | 0 |
|  | \# | PEAK | value |  |  |  |  |  |  |  |  |  |  |
| -6 | 4 | -1.27 | -0.75 | 0 | 10 | 2 | 0 | 0 | 0 | 0 | 0.5 | 0 | 0 |
|  | \# | TOP | logistic |  |  |  |  |  |  |  |  |  |  |
| -1 | 9 | 3.4 | 3.5 | 0 | 10 | 2 | 0 | 0 | 0 | 0 | 0.5 | 0 | 0 |
|  | \# | WIDT | exp |  |  |  |  |  |  |  |  |  |  |
| -1 | 9 | 3.68 | 5 | 0 | 10 | 2 | 0 | 0 | 0 | 0 | 0.5 | 0 | 0 |
|  | \# | WIDT | exp |  |  |  |  |  |  |  |  |  |  |
| -10 | 9 | -3.37 | -4.5 | 0 | 10 | 2 | 0 | 0 | 0 | 0 | 0.5 | 0 | 0 |
|  | \# | INIT | logistic |  |  |  |  |  |  |  |  |  |  |
| -10 | 9 | 0.79 | 2.9 | 0 | 10 | 2 | 0 | 0 | 0 | 0 | 0.5 | 0 | 0 |
|  | \# | FINA | logistic |  |  |  |  |  |  |  |  |  |  |
| \# size_se1: 10-male offsets- 4 lines |  |  |  |  |  |  |  |  |  |  |  |  |  |
| \#1 | 60 | 16 | 20 | 0 | 10 | -5 | 0 | 0 | 0 | 0 | 0.5 | 0 | 0 |
|  | \# | size@ | gleg |  |  |  |  |  |  |  |  |  |  |
| \#-10 | 10 | 0 | 0 | 0 | 10 | -5 | 0 | 0 | 0 | 0 | 0.5 | 0 | 0 |
|  | \# | $\log (\mathrm{re}$ | malesel)at |  |  |  |  |  |  |  |  |  |  |
| \#-10 | 10 | 0 | 0 | 0 | 10 | -5 | 0 | 0 | 0 | 0 | 0.5 | 0 | 0 |
|  | \# | $\log (\mathrm{re}$ | alesel)at | dog |  |  |  |  |  |  |  |  |  |
| \#-10 | 10 | 0 |  | 0 | 10 | -5 | 0 | 0 | 0 | 0 | 0.5 | 0 | 0 |
|  | \# | $\log (\mathrm{re}$ | malesel) at | m |  |  |  |  |  |  |  |  |  |
| \# |  |  |  |  |  |  |  |  |  |  |  |  |  |
| \# |  |  |  |  |  |  |  |  |  |  |  |  |  |
| \#_age_sel: 1 |  |  |  |  |  |  |  |  |  |  |  |  |  |
| \#_age_sel: 2 |  |  |  |  |  |  |  |  |  |  |  |  |  |
| \#_age_sel: 3 |  |  |  |  |  |  |  |  |  |  |  |  |  |
| \#_age_sel: 5 |  |  |  |  |  |  |  |  |  |  |  |  |  |
| \#_age_sel: 6 |  |  |  |  |  |  |  |  |  |  |  |  |  |
| \#_age_sel: 7 - juv survey 1 |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 00000010-30000000\#39 |  |  |  |  |  |  |  |  |  |  |  |  |  |
| $0000010-30000000$ \# 40 |  |  |  |  |  |  |  |  |  |  |  |  |  |
| \#_age_sel: 8 - juv survey 2 |  |  |  |  |  |  |  |  |  |  |  |  |  |
| $0000010-30000000$ \# 39 |  |  |  |  |  |  |  |  |  |  |  |  |  |
| $0000010-30000000$ \# 40 |  |  |  |  |  |  |  |  |  |  |  |  |  |
| \#_age_sel: 10 |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 1 | 10 | 1 | 1 | 0 | 1 | 2 | 0 | 0 | 0 | 0 | 0.5 | 0 | 0 |
|  | \# | PEAK | value |  |  |  |  |  |  |  |  |  |  |



```
0.68
0.35
1
1
1
2 . 5
#1111111111#_mult_by_lencomp_N
1.43714
5.41864
4.24022
1
1
0.75
1
1
1
1
#1111111111#_mult_by_agecomp_N
1111111111#_mult_by_size-at-age_N
30 #_DF_for_discard_like
30 #_DF_for_meanbodywt_like
1 #_maxlambdaphase
0 #_sd_offset
#_lambdas_(columns_for_phases)
# #_CPUE/survey:_1
# #_CPUE/survey:_2
# #_CPUE/survey:_3
# #_CPUE/survey:_4
# #_CPUE/survey:_5
1 #_CPUE/survey:_6
# #_CPUE/survey:_7
# #_CPUE/survey:_8
# #_CPUE/survey:_9
#_CPUE/survey:_10
0 #_discard:1
0 #_discard:_2
0 #_discard:_3
0 #_discard:_4
0 #_discard:_5
0 #_discard:_6
0 #_discard:_7
0 #_discard:_8
0 #_discard:_9
0 #_discard:_10
0 #_meanbodyweight
0.1 #-lencomp:1
0.1 #_lencomp:_2
0.1 #_lencomp:_3
# #_lencomp:_4
1 #_lencomp:_5
0.1 #_lencomp:_6
0 #_lencomp:_7
#_lencomp:_8
#_lencomp:_9
#_lencomp:_10
```

```
# #_agecomp:_1
# #_agecomp:_2
# #_agecomp:_3
0 #_agecomp:_4
0 #_agecomp:_5
1 #_agecomp:_6
0 #_agecomp:_7
0 #_agecomp:_8
0 #_agecomp:_9
0 #_agecomp:_10
0 #_size-age:_1
# #_size-age:_2
# #_size-age:_3
# #_size-age:_4
# #_size-age:_5
# #_size-age:_6
0 #_size-age:_7
# #_size-age:_8
# #_size-age:_9
# #_size-age:_10
0 #_init_equ_catch
1 #_recruitments
0 #_parameter-priors
# #_parameter-dev-vectors
100 #_crashPenLambda
0.9 #_maximum allowed harvest rate
999
```


## Appendix C: Detailed list of STAR Panel requests and STAT responses.

## Round 1 requests

A. Compare the length-composition of the aged fish with non-aged fish for each fishery and each year.
B. Fix the code for the recreational CPUE to be number-based rather than biomass-based.
C. Reset the lambdas on LFs to 0.1 if age data exist, and to 1 if there are no associated age data for the same samples. Run with:

- No CalCOFI or core juvenile;
- No time varying K - fix at the values of all growth parameters of the earlier conditional runs;
- Trawl CPUE indices;
- Rec CPUE;
- Triennial Survey;
- Combined survey;
- Coast-wide juvenile index;
- Fix h at something reasonable;
- Fix M for females and estimate offset for males;
- Fix CV of length at age at 0.06 [based on external analysis done by the STAT];
- Profile over M including likelihood components;
- Estimate selectivity parameters;
- Estimate SSB0;
- Estimate depletion.
D. Save the results from the un-tuned model
E. Tune the trial reference model - see fit for everything. Plots and tables of diagnostics and results.
F. Profile over M for the tuned model looking at individual likelihood components - identify inconsistencies among data sources.
G. Plot or tabulate spatial distribution of samples in recreational data from observers over time.


## Round 1 responses

A. The length-compositions of the aged and non-aged chilipepper rockfish were for approximately $50 \%$ of the samples from each fishing gear. The results suggested that the size compositions of aged versus unaged fish (plotted as individuals, rather than expanded length compositions) may be biased for some years.
B. The SS2 control switch for the CPFV survey (the recreational fishery CPUE index) was corrected to indicate that the data represented numbers of fish rather than biomass.
C. The SS2 model specified for this request was set up and run with steepness fixed at 0.57 and female natural mortality fixed at 0.16 , consistent with the point estimate of steepness associated with the informative prior and the results of profiling over natural mortality. The length-
composition data were down-weighted as requested, which was recognized by both the STAR Panel and the STAT as an ad hoc correction for non-independence of the data.
D. Results of the un-tuned model were saved as requested.
E. The revised model was tuned and the results evaluated. As with the earlier model, the relative abundance indices failed to reflect the increase in biomass associated with the large 1984 cohort apparent in observed data. Similarly, the predicted values for the CPFV survey (which began in 1988) showed no decline despite a clear downward trend in the observed values for this index.
F. The profile plot over M revealed tension between the data sets, particularly between the trawl fishery (particularly the length composition data, but including the trawl CPUE time series) and the recreational CPFV survey (with the triennial survey tending to be in agreement with the recreational CPFV survey). Higher estimates of spawning stock biomass were associated with higher values of $M$.
G. Plots of the number of observed CPFV trips and the number of chilipepper rockfish caught by depth categories and year demonstrated that a relatively small number of samples from deeper depths, each of which encountered large number of fish, were recorded in the years prior to 1994. To ensure consistency in depth ranges covered by the survey through time, trips taken in depths greater than 80 fathoms were excluded from the GLM analysis. The location of the blocks that were included in the CPUE index was also displayed graphically to the STAR Panel, and although a majority of these blocks occurred in the Cordell Bank and Monterey Bay regions, the locations ranged from just south of Point Arena to the Morro Bay region. This spatial coverage was considered adequate (albeit not optimal) for reflecting relative trends throughout the core area of the stock biomass.
H. In the spirit of the discussions with the STAR Panel, the CPUE index was also reproduced using the Stephens/MacCall filter, which was very similar to that produced by the GLM using depth and block data. This indicated that the filter was working properly to identify trips likely to catch chilipepper, although both the STAT and the STAR agreed to continue with the GLM based on location and depth data. The CVs of the results from the filter were less than those from the GLM. Based on discussions with the STAR Panel regarding the triennial survey indices developed with GLMM approaches and area-swept estimates of biomass, a more detailed description of the GLMM analysis provided by T. Helser (pers. Com) was also presented to the STAR Panel. Both the STAT and the STAR agreed that the GLMM provided good predictions of the data.

## Round 2 requests

Based on the reference run that was established on Monday evening (Round 1):
H. Test for block-year interaction in GLM for recreational observer CPFV data. If a strong interaction is detected, report back to this issue and complete points I to M , but do not undertake the additional runs at points N to P .
I. Plot length-compositions of aged versus non-aged fish in remaining samples to determine those samples which are relatively unbiased. Weed out obviously biased samples from the SS2 input including those samples that had infeasible numbers of large males.
J. Investigate samples that had extraordinarily large proportions of males.
K. Link RecFIN length-compositions to the recreational fishery and CPFV observer lengthcomposition to the CPFV CPUE survey to assist in elucidating the respective selectivity curves.
L. Remove whole of deep trips $>80$.
M. Use Helser's GLMM rather than area swept index.
N. Estimate an appropriate selectivity pattern for triennial survey.
O. Systematically set lambda for recreational observer CPFV index to $1,5,10, \ldots$ till a reasonable fit to this index is attained and investigate changes in likelihood for all other components.
P. Profile over R0 as was done for M, plotting against B0.

## Round 2 responses

H. Due to the large number of interaction parameters necessary to adequately test for interactions between year and block effects, it was not possible to detect block-year interactions in a satisfactory manner, however the indication was that there were no significant interactions.
I. The length-compositions of aged and non-aged samples were plotted for samples not examined in the initial request, and several potentially problematic years of age-composition data were excluded from further analysis (see the section on commercial age and length composition data for specific years that were effectively removed from the objective function).
J. After filtering to remove outliers, the length-composition for one sample still contained a number of unfeasibly large males. This length-composition year was also "turned off" in all subsequent analyses as well as the base model.
K. In the preliminary model the CPFV index was biomass-based and was linked with the recreational fishery along with the CPFV length-composition data. In discussions with the STAR Panel it was agreed to treat the CPFV index and length compositions as a separate survey, and use RecFIN length-composition data to represent the full range of recreational fishing modes. These changes did not have a major effect on the model results.
L. Removing the data for trips $>80$ fathoms, including associated length data, had little effect on the biomass trajectory.
M. The use of the GLMM results rather than the swept area indices for the triennial and NWFSC combination survey resulted in slightly greater depletion than in the previous run. As the GLMM analysis was agreed to more appropriately account for the highly variable nature of tow-specific catch rates, this was agreed by both the STAT and the STAR Panel to be a more appropriate index for the final model and was used in all further analyses.
N. The selectivity curve for the triennial survey was essentially a horizontal line, with the result that the parameters were poorly specified and the Hessian for this run could not be inverted. To invert the Hessian required fixing the selectivity parameters at their estimated values.
O. Elevated lambdas on the CPFV index resulted in lower biomass trajectories and apparently greater depletion, with a better fit to the CPFV and triennial indices but poorer fit to the trawl CPUE index. However, even with lambda $=25$ the predicted CPFV index failed to reflect the increase in biomass that resulted from the 1984 year class, which was evident in other data
sources. A more effective approach for capturing the signal of the 1984 year class was to set the CPFV index lambda at 5, and incorporate both length and age selectivity (similar to the sablefish model), and including time-varying growth (with a 3-year blocking pattern). The resulting predicted length-compositions for the CPFV survey reflected the bimodality present in the observed length data, which was not as well reflected when using length-based selectivity alone.
P. The STAT had insufficient time to satisfy this request.

## Round 3 requests

Q. Modify the SS2 input specification to turn off the age-composition data where samples were biased (as determined from comparison of aged and non-aged LF data) and turn lengthcomposition data back on. For the sample with an infeasible number of large males, turn off both age and length-compositions.
R. Using lambda for CPFV survey data set to 1 , run SS 2 to provide a reference for subsequent runs
S. Investigate alternative parameterisation for sex-specific selection curves for the CPFV survey using either age OR length selection (but not both) and hence determine a suitable selection pattern to use. Save runs.
T. Using the final selection curve from Request S, produce a simple profile analysis based on R0 to explore the tension among different indices and data sets.

## Round 3 responses

Q. The changes were completed to remove the effect of biased sampling for age but retain the associated length data.
R. The run was completed as requested. Turning off the biased age-composition data did not have a major impact on the predictions of biomass, nor did it help the fit to the CPFV survey data.
S. The rationale for this request was to find a selection curve for the CPFV survey that would fit the CPFV index and length-composition data without the complexity of the composite age- and length-based curve that the STAT had used in response O. The STAT replaced the CPFV length-based selection curve with an age-based curve, which went asymptotic when fitted. The resulting fit appeared slightly better than that obtained with length-based selectivity. However, the request that the selectivity curve be sex-specific was not implemented. Consequently the response to request T was not informative, and that request was repeated in the next round.

## Round 4 requests

U. Complete Request S. That is, search for alternative parameterisation for sex-specific selection curves for the CPFV survey using either age OR length selection (but not both) and hence determine a suitable selection pattern to use. Save runs.
V. Using the final selection curve from Request U, produce a simple profile analysis based on $\mathrm{R}_{0}$ to explore the tension among different indices and data sets.
W. Explore alternative blocking for time-varying growth based on external environmental variables.

## Round 4 responses

U. The STAT attempted to find an alternative parameterization for sex-specific selectivity curves, but was unable to fit an age-based or length-based, sex-specific selection curve that provided as good a model fit as that obtained by the combined age- and length-based selection curve (which were not sex-specific).
V. The relative impact on the overall likelihood of the different model components at different values of $\mathrm{R}_{0}$ could not be compared easily using the profile plots because the plots did not account for the effect of lambda, which was reduced to 0.1 for some components. Using sexspecific selection for the CPFV survey did not appear to warrant further investigation.
W. An alternative block formulation was developed based on the major shifts in the sign of the Pacific Decadal Oscillation (PDO) index, which has been shown to be related to physical ocean conditions, zooplankton production, salmon smolt survival and other indices of marine productivity. The Panel agreed with the STAT that the PDO provided an adequate basis for blocking offsets for the growth parameter K into six time-blocks. The results included a large improvement in the log-likelihood, but the value of K for the final time-block was far lower than the values for previous time-blocks.

## Round 5 requests

X. Investigate feasibility of driving K with PDO (spend no more than half hour on this task).
Y. Adopt time-varying growth based on the better of using either PDO blocks (with slightlyinformative prior on K to avoid infeasible reduction in K for last period) or using environmentally-driven growth (Request X ), and using both age and size-selectivity on the CPFV CPUE recreational survey, create tuned base. Demonstrate adequate convergence of tuned run.
Z. Produce profile plots on R0 accounting for lambda.

AA. Using base run, produce standard diagnostics for STAR Panel review.

## Round 5 responses

X. The direct forcing of the growth parameter K with a three-year running mean of the PDO index showed promise, and resulted in an improved fit (approximately 25 likelihood units) relative to the time-invariant K model. However, the improvement in fit was notably less than using blocked time intervals, and consequently it was agreed that the base model should use the timeblocking approach.
Y. A value of 0.5 was used as the standard deviation for a slightly informative prior on K for the configuration with six PDO-based time-blocks for changes in K. The convergence-test runs that used "jittered" starting parameter values revealed convergence problems, suggesting that the likelihood surface is quite irregular. Requests Z and AA were not completed due to these convergence problems.

## Round 6 requests

AB. Explore convergence and results of time-varying $K$ with (a) last two blocks combined into a single large block and (b) changing the standard deviation for the prior on the deviations on K from 0.5 to 0.35 .

AC. Use 0.5 on the K-dev prior. Run with five-block rather than 6-block model. Examine results.
AD. Turn off all priors. Run with five-block rather than 6-block model. Examine results
AE. Use run from Request AD. Clean up initial values. Make qs analytical. Clean up phasing. Do jitters and alternative phasing to confirm model convergence. If not converged, report back ASAP. If converged, produce a full set of diagnostic results and profile plots on $\mathrm{R}_{0}$ accounting for lambda. If these are satisfactory, this will be the base model.

## Round 6 responses

AB . The two requested runs explored alternative methods for constraining the growth coefficient K in the final time block. The Panel was concerned that the unconstrained estimate for the final K value was extremely small and would have a strong influence on forecasts. The run with the standard deviation for the prior probability reduced to 0.35 still produced a low value for the final K . The run that merged the last two blocks in combination with a standard deviation of 0.35 for the prior probability resulted in an intermediate value of K .

AC. The Panel sought confirmation that having the longer final block in the five-block model would provide sufficient constraint for the final K value and that the prior probability on the K-offsets could be eliminated. The use of a standard deviation value of 0.5 for the prior probability on the K-offsets had little effect on the results.

AD. As several parameters had very modest likelihood values associated with weakly informative priors other than the offsets to K, all prior probabilities were removed and the lambda on priors was set to zero in order to simplify the model configuration.
AE. Convergence test runs with jittered initial parameter values indicated there still were convergence problems associated with roughness in profile plots, although the effects did not appear too severe. The panel provided guidance to jitter the final profile plots in the revised assessment to ensure convergence to the best model fit, and this was done for all sensitivity runs.

## Round 7 requests

AF. Set process error added to CPFV survey indices to 0 . Re-run. Confirm that this is appropriate to use as a base model through jitters and alternative phasing to confirm model convergence.
AG. With settings resulting from Request AF, increase emphasis to 20 on both CPFV survey indices and length frequencies to estimate age-based, sex-specific selectivity. Assess whether this gives sensible selection patterns. If so, using the resulting parameter space and selectivity pattern (possibly fixing selectivity parameters to the resulting values), de-emphasise, re-fit, and re-tune to produce plausible alternative results (removing process error if necessary after tuning). Note - no more than $\sim 45$ minutes to be spent on this task. Produce a plot of the
biomass trajectory of this compared with the result from Request AF as a sensitivity analysis. Compare the depletion estimates.
AH. With settings resulting from Request AF, explore the following dimensions of uncertainty using low and high values for (a) historical catch prior to 1978 (half and double), (b) M, and (c) h. Retain SS2 results from each run. Produce comparative plots of the biomass trajectories of these compared with the result from Request AF. Produce a table showing comparison of likelihood contributions from different components. Produce a table of comparative depletion estimates.

## Round 7 responses

AF. Removing the variance adjustment on the CPFV survey index had the desired effect of producing a better fit to the CPFV survey. After reviewing diagnostic plots the Panel recommended acceptance of this model configuration as the base model.
AG. These sensitivity runs re-explored using an alternative configuration for the CPFV survey selection curve. Previous explorations had increased the lambda on the CPFV survey index but not on the CPFV length-composition data. The new runs produced a very good fit to the CPFV index even when lambda was decreased from 20 to 10, but the CPFV selectivity curve had been configured as age- and length-based and sex-specific. Convergence tests with jittered initial parameter values still produced fits that appeared not fully converged.
During discussions the STAT indicated that the CVs for the triennial and combination surveys had been reduced externally rather than with a variance adjustment factor in the SS2 control file. Because the model provided good fits to several survey data points that had very large input CVs, the standard variance adjustment approach would have produced negative CVs for other data points with small input CVs. The Panel notes that further consideration is needed to develop an appropriate approach for handling survey variance adjustments that could potentially become negative.

AH. The runs were completed as requested. The resulting profile plots were somewhat jagged, suggesting that the model had failed to converge fully at many values of the reference variable. Following examination of the profile plots the Panel concluded that, of the variables considered, h was likely to provide the most useful axis of uncertainty. The Panel recommended assuming a normal distribution for $h$ with a mean value of 0.573 and standard deviation of 0.183 to determine the bracketing values.

## Round 8 requests

AI. Complete Request AG to estimate age-based, sex-specific selectivity. Run and produce comparison of results.

AJ. For developing a decision table, run the base model with $\mathrm{h}=0.34$ and 0.81 [mean values of the lower and upper $25 \%$ of the prior probability distribution for h ] to obtain results likely to be representative of the lower $25 \%$ and upper $25 \%$ of values, respectively. Use the alternative phasing supplied by the STAR Panel. Jitter and ensure convergence for each value of $h$.

## Round 8 responses

AI. The response to AG had used a sex-specific, age- and length-based selection curve for the CPFV survey. Results demonstrated that, although needing further refinement, an age-based, sex-specific selectivity curve could be developed to replace the age- and length-based, sexspecific selectivity curve.

AJ. While there were still convergence issues that required jittering of input parameter values for each analysis, the jittered runs for each level of steepness produced reasonably similar results. Depletion for the base case was 0.7 , while those from the lower and higher values of $h$ were 0.46 and 0.78 , respectively. The Panel accepted that use of these values of $h$ produced the required lower and upper runs to bracket uncertainty around the base-run results.


[^0]:    | Year | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 |
    | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |

    Expl. Rate $0.1190 .1130 .1330 .098 \quad 0.0710 .0370 .028 \quad 0.0140 .0010 .008 \quad 0.006 \quad 0.004$

